FIXED-BED ADSORPTION OF CARBON DIOXIDE ONTO AMMONIA-MODIFIED ACTIVATED CARBON: EXPERIMENTAL AND MODELING STUDY

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ABSTRACT

A commercial granular activated carbon (GAC) adsorbent was modified through an oxidation–amination process in an effort to increase its surface basicity and consequently enhance its CO₂ adsorption capacity. To optimize the amination conditions of activated carbon adsorbents the effects of amination temperature, amination time, and the type of starting materials (variables) on the CO₂ adsorption/desorption capacities of the adsorbents (responses) were investigated using a central composite design. The use of a pre-oxidized sorbent as a starting material and amination at 425 °C for 2.1 h were found to be the optimum conditions for obtaining an efficient carbon dioxide adsorbent. The activated carbon modified at optimum conditions (OXA-GAC) exhibited CO₂ adsorption and desorption capacity values of 26.47 mg/g and 95.4%, respectively. The promising characteristics of the OXA-GAC in terms of adsorption capacity (exhibiting an increase of 44% in capacity compared with the capacity of the GAC at 1 atm and 105 °C) and multicycle durability make it suitable for practical applications.

The equilibrium adsorption isotherms of CO₂ on the GAC and the OXA-GAC were measured using a static volumetric method. CO₂ adsorption measurements were performed at three different temperatures (303, 318, and 333 K) and pressures up to 1 atm. The obtained equilibrium data were fitted to the Freundlich, Sips, and Toth isotherms using a semi-empirical approach to differentiate the contributions of physical and chemical adsorption to the total CO₂ uptake. The Toth semi-empirical equilibrium model provided the best fit to the experimental data, over the entire analyzed ranges of temperature and pressure. The isosteric heats of CO₂ adsorption onto the GAC and OXA-GAC adsorbents were determined using the Clausius–Clapeyron equation. The initial isosteric heats of adsorption of 68 kJ mol⁻¹ and 23 kJ mol⁻¹ corresponded to the chemisorption and physisorption of CO₂ on the OXA-GAC adsorbent, respectively, and these values were in excellent agreement with the zero-coverage heats of adsorption obtained using the temperature-dependent parameters of the proposed model.

The kinetics of CO₂ adsorption on the GAC and OXA-GAC adsorbents over the temperature range of 30-60 °C were studied using the pseudo-first-order, pseudo-secondorder, and Avrami kinetic models. The best fit with the experimental kinetic data for both of the studied adsorbents was obtained by applying the Avrami kinetic model. Fixed-bed breakthrough experiments for CO₂ adsorption onto the GAC and OXA-GAC adsorbents were performed by changing the adsorption temperature over the range of 30 to 60 °C and the feed flow rate from 50 to 100 ml min⁻¹. The largest values of the CO_2 equilibrium dynamic capacity (0.67 mol kg⁻¹) and breakthrough time (10.9 min) over the range of operating conditions investigated were obtained using OXA-GAC adsorbent at 30 °C under a 50 ml min⁻¹ feed flow rate. To predict the breakthrough behavior of the fixed-bed adsorption of CO₂, a simple model was developed, including the Toth and Avrami equations to describe the equilibrium and kinetics of adsorption, respectively. The set of coupled differential equations was solved using a numerical approach based on the finite element method implemented in COMSOL Multiphysics software. The validity of the model predictions was evaluated by a comparison with the experimental data. The findings showed that the model predictions successfully fit the experimental data over the studied range of feed gas flow rates and adsorption temperatures.

ABSTRAK

A penjerap komersial karbon teraktif berbutir (GAC) telah diubahsuai melalui proses pengoksidaan-aminasi dalam usaha untuk meningkatkan permukaan kealkalian dan seterusnya meningkatkan keupayaan penjerapan CO₂. Untuk mengoptimumkan keadaan ammoksidaan pada penjerap karbon teraktif, kesan suhu aminasi, masa aminasi dan jenis permulaan bahan (pembolehubah) pada penjerapan CO₂/ kapasiti penyahjerapan penyerap yang diubahsuai (tindakbalas) telah dikaji menggunakan reka bentuk komposit pusat. Penggunaan penyerap pra-teroksida sebagai bahan permulaan dan aminasi pada 425 °C selama 2.1 jam didapati sebagai keadaan optimum untuk mendapatkan penyerap CO₂ yang berkesan. Karbon teraktif diubahsuai pada keadaan optimum telah menunjukkan penjerapan CO₂ dan nilai kapasiti penyahjerapan sebanyak 26.47 mg /g dan 95.4%. Ciriciri yang menggalakkan terhadap OXA-GAC dari segi kapasiti penjerapan (mempamerkan peningkatan sebanyak 44% dalam kapasiti berbanding dengan kapasiti GAC pada 1 atm and 105 °C) dan daya tahan pusingan perlbagai menjadikan ia sesuai untuk aplikasi praktikal.

Keseimbangan isoterma penjerapan CO₂ pada GAC dan penjerap yang optimum (OXA-GAC) telah diukur dengan menggunakan kaedah isipadu statik. Pengukuran penjerapan CO₂ telah dijalankan pada tiga suhu yang berbeza (303, 318, dan 333 K) dan pada tekanan sehingga 1 atm. Data keseimbangan diperolehi dipadankan dengan Freundlich, Sips dan Toth isoterma menggunakan pendekatan semi-empirikal untuk membezakan sumbangan penjerapan fizikal dan kimia untuk jumlah pengambilan CO₂. Keseimbangan analysis untuk julat suhu dan tekanan. Saringan isosteric penjerapan CO₂ pada penjerap ammonia-diubahsuai dan pada penjerap yang tidak dirawat telah ditentukan dengan menggunakan persamaan Clausius-Clapeyron. Saringan awal isosteric penjerapan adalah 68 kJ mol⁻¹ dan 23 kJ mol⁻¹ sepadan dengan chemisorption dan physisorption CO₂ pada penjerap yang

v

diubah suai, dan nilai-nilai ini sangat sepadan dengan penjerapan saringan sifar-liputan diperolehi menggunakan parameter bersandar suhu kepada model yang dicadangkan.

Kinetik penjerapan CO₂ pada GAC dan OXA-GAC penjerap pada julat suhu 30-60 °C telah dikaji menggunakan pseudo-tertib pertama, pseudo-tertib kedua, dan Avrami model kinetik. Yang sepadan dengan data kinetik eksperimen untuk kedua-dua penjerap dikaji telah diperolehi dengan menggunakan model kinetik Avrami. "Fixed-bed breakthrough" eksperimen bagi penjerapan CO_2 ke atas GAC dan OXA-GAC penjerap yang dilakukan dengan menukar suhu penjerapan ke lingkungan julat 30 hingga 60 °C dan kadar aliran masuk dari 50-100 ml min⁻¹. Nilai terbesar kapasiti dinamik keseimbangan CO_2 (0.67 mol kg⁻¹) dan masa "breakthrough" (10.9 min) ke atas pelbagai keadaan operasi yang disiasat telah diperolehi dengan menggunakan OXA-GAC penjerap pada 30 °C di bawah aliran 50 ml min⁻¹ kadar aliran masuk. Untuk meramalkan keadaan "breakthrough" daripada penjerapan "fixed-bed" CO₂, model yang mudah telah dibangunkan, termasuk Toth dan Avrami persamaan untuk menggambarkan keseimbangan dan kinetik penjerapan, masingmasing. Beberapa set persamaan pembezaan telah diselesaikan dengan menggunakan pendekatan yang berangka berdasarkan kaedah unsur terhingga dilaksanakan dalam perisian COMSOL Multifizik. Kesahihan ramalan model telah dinilai oleh perbandingan dengan data uji kaji. Keputusan kajian menunjukkan bahawa ramalan model berjaya sepadan dengan data uji kaji pada julat yang dikaji pada kadar aliran gas masuk dan suhu penjerapan.

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TABLE OF CONTENTS

ΓΙΤLE PAGEi
ORIGINAL LITERARY WORK DECLARATION FORMii
ABSTRACTiii
ABSTRAKv
ACKNOWLEDGEMENTSviii
ΓABLE OF CONTENTSix
LIST OF FIGURES xiii
LIST OF TABLESxv
LIST OF SYMBOLS AND ABBREVIATIONSxvii
CHAPTER 1: INTRODUCTION1
1.1 Background1
1.2 Research objectives
1.3 Thesis organization
CHAPTER 2: LITERATURE REVIEW6
2.1 Introduction
2.2 Overview of the prediction of adsorption column dynamics
2.3 Development and analysis of a mathematical model
2.3.1 Fluid phase material balance
2.3.2 Complexity of kinetic models
2.3.2.1 Local equilibrium model35
2.3.2.2 Mass transfer resistance models
2.3.3 Energy balance
2.3.3.1 Gas phase energy balance

	2.3	.3.2 Solid phase energy balance	.58
	2.3	.3.3 Wall energy balance	.59
2	.3.4	Momentum balance	.61
CHAF	PTER 3:	MATERIALS AND METHODS	.64
3.1	Introd	uction	.64
3.2	The aj	pplication of response surface methodology to optimize the amination	of
acti	vated ca	arbon for the preparation of carbon dioxide adsorbents	.64
	3.2.1	Adsorbent materials	.64
	3.2.2	Ammonia modification	.64
	3.2.3	Experimental design and method of analysis	.66
	3.2.4	CO ₂ adsorption/desorption measurements	.72
3.3	A sem	ni-empirical model to predict adsorption equilibrium of carbon dioxide	on
amı	nonia n	nodified activated carbon	.73
	3.3.1	Adsorbent materials	.73
	3.3.2	Equilibrium CO ₂ adsorption measurements	.73
	3.3.3	Adsorption isotherm equations	74
3.4	Model	ling of carbon dioxide adsorption onto ammonia-modified activated	
cart	oon: Kir	netic analysis and breakthrough behavior	76
	3.4.1	Adsorbent materials	76
	3.4.2	Kinetic adsorption measurements	76
	3.4.3	Fixed-bed adsorption experiments	77
	3.4.4	Model description and solution methodology	79
		3.4.4.1 Kinetic models	79
		3.4.4.2 Modeling dynamic column breakthrough experiments	84
		3.4.4.3 Solution methodology	87

CHAPTER 4: R	ESULTS AND DISCUSSION88
4.1 Introduct	tion
4.2 The app	lication of response surface methodology to optimize the amination of
activated carb	oon for the preparation of carbon dioxide adsorbent
4.2.1 E	Evaluation of CO ₂ adsorption and desorption capacity92
4.2.1.1	Effect of amination variables on the CO ₂ adsorption capacity96
4.2.1.2	Effect of amination variables on the CO ₂ desorption capacity99
4.2.2 In	nvestigation of the optimum amination conditions102
4.2.3 R	Reusability of the sorbent prepared under optimum conditions105
4.3 A semi-	empirical model to predict adsorption equilibrium of carbon dioxide on
ammonia moc	lified activated carbon106
4.3.1 A	Adsorption equilibrium study106
4.3.2 E	Equilibrium isotherms modeling109
4.3.3 Is	sosteric heat of adsorption117
4.4 Modelin	g of carbon dioxide adsorption onto ammonia-modified activated carbon:
Kinetic analys	sis and breakthrough behavior122
4.4.1 A	Adsorption kinetics
4.4.2 C	Column breakthrough experiments and model validation
4.4.2.1	Effect of temperature on the breakthrough profile of CO ₂
adsorptio	on132
4.4.2.2	Effect of feed flow rate on the breakthrough curve for the
adsorptio	on of CO ₂
4.4.2.3	Validation of the proposed model139
CHAPTER 5: C	ONCLUSION AND RECOMMANDATIONS141

5.1	Conclusion	141
5.2	Recommendations	143
REFE	RENCES	144
LIST (OF PUBLICATIONS	164

LIST OF FIGURES

Figure 2.1	Schematic diagram showing various resistances to the transport of adsorbate
as well as	concentration profiles through an idealized bidisperse adsorbent particle
demonstratin	g some of the possible regimes47
Figure 3.1	Schematic diagram of the experimental setup for the ammonia
modification	
Figure 3.2	Schematic of the experimental system used for the column breakthrough
measuremen	ts79
Figure 4.1	Response surface plot of CO ₂ adsorption capacity for (a) pre-heat treated,
and (b) pre-o	xidized adsorbent
Figure 4.2	Response surface plot of CO ₂ desorption capacity for (a) pre-heat treated,
and (b) pre-o	xidized adsorbent
Figure 4.3	Contour plot of (a) CO_2 adsorption capacity and (b) CO_2 desorption capacity
as a function	of aminaion temperature and time for pre-oxidized adsorbents104
Figure 4.4	Overlay plot of the optimal region for pre-oxidized adsorbents105
Figure 4.5	Cyclical adsorption and desorption of CO_2 at 105 °C by the optimal activated
carbon adsor	bent
Figure 4.6	Experimental adsorption isotherms of CO_2 on (a) modified and (b) untreated
activated car	bon measured at 30, 45 and 60 °C108
Figure 4.7	Calculated adsorption isotherms for CO ₂ chemisorption onto the modified
adsorbent at	30, 45 and 60 °C111
Figure 4.8	Graphical evaluation of the fit of the experimental equilibrium data to the
proposed mo	del for the modified adsorbent, whose parameters are presented in Tables 4.6
and 4.7. The	surface is the global isotherm model, and the black and white circles show
the experime	ntal data at 303, 318 and 333 K117

Figure 4.9 Adsorption isosteres of CO_2 for (a) modified and (b) untreated adsorbent in the temperature range from 303 to 333 K. The points were calculated by numerical interpolation, and the lines represent the linear fit. All of the isosteres are marked with the corresponding amount of CO_2 adsorbed in units of (cm³ STP/g)......119

Figure 4.11 Experimental CO₂ adsorption onto modified and untreated adsorbents at (a) $30 \,^{\circ}$ C, (b) 45 $\,^{\circ}$ C, and (c) 60 $\,^{\circ}$ C along with the corresponding fit to kinetic models.....123

Figure 4.15 Effect of feed flow rate on breakthrough curves of CO_2 adsorption onto GAC and OXA-GAC adsorbents at (a) 30 °C (b) 45 °C and (c) 60 °C.....138 Figure 4.16 Verification of the proposed model with experimental breakthrough curves at three representative operating conditions.

Discrete symbols are experimental data, and solid lines are predictions by the model....140

LIST OF TABLES

Table 2.1	Summary of the dynamics models for fixed-bed adsorption of CO_2 12
Table 3.1	Independent numerical variables and their levels (actual and coded)67
Table 3.2	Experimental design layout and experimental results of the responses70
Table 3.3	Physical properties of the adsorbent and characteristics of the adsorption bed
along with th	e operating conditions used for the fixed-bed experiments79
Table 3.4	Optimal values of the proposed Toth equilibrium isotherm parameters82
Table 3.5	Correlations used for estimation of the model parameters
Table 4.1	Analysis of variance (ANOVA) for the CO ₂ adsorption capacity93
Table 4.2	Analysis of variance (ANOVA) for the CO ₂ desorption capacity94
Table 4.3	Predicted and experimental values of the studied responses obtained at
optimum con	ditions105
Table 4.4	Freundlich isotherm parameters with R^2 and ARE for each independent
mechanism a	t temperatures of 303, 318, and 333 K114
Table 4.5	Sips isotherm parameters with R^2 and ARE for each independent mechanism
at temperatur	res of 303, 318, and 333 K
Table 4.6	Toth isotherm parameters with R^2 and ARE for each independent
mechanism a	t temperatures of 303, 318, and 333 K115
Table 4.7	Optimal values of the proposed Toth temperature-dependent
parameters	
Table 4.8	The calculated parameters of the kinetic models and associated R^2 and Δq
(%) for the C	O ₂ adsorption onto GAC at different temperatures
Table 4.9	The calculated parameters of the kinetic models and associated R^2 and Δq
(%) for the C	O ₂ adsorption onto OXA-GAC at different temperatures128

LIST OF SYMBOLS AND ABBREVIATIONS

a_a	Ratio of the external surface area to the volume of the column wall
a_s	Ratio of the particle external surface area to volume
a_w	Ratio of the internal surface area to the volume of the column wall
ANOVA	Analysis Of Variance
ARE	Average Relative Error
BET	Brunauer, Emmette and Teller
Bi	Biot number
С	Adsorbate concentration in the fluid phase
Ср	Adsorbate concentration in the macropore
С	Total concentration in the bulk phase
C_g	Gas phase heat capacity
C_s	Adsorbent heat capacity
C_w	Column wall heat capacity
CCD	Central Composite Design
d _{ext}	Column external diameter
d _{int}	Column internal diameter
d_p	Particle mean diameter
D_m	Molecular diffusivity
D_p	Macropore diffusivity
D_z	Effective axial dispersion coefficient
Dμ	Micropore diffusivity
$D_{\mu i}{}^\infty$	Micropore diffusivity at infinite dilution
DDW	Distilled Deionized Water

DF	Degree of Freedom
E_a	Activation energy
8	Gravity acceleration
GAC	Granular Activated Carbon
h _{ext}	External convective heat transfer coefficient
h_{f}	Film heat transfer coefficient
h_w	Internal convective heat transfer coefficient
$-\Delta H$	Isosteric heat of adsorption
НТА	Pre-heat-treated GAC
IUPAC	International Union of Pure and Applied Chemistry
<i>k</i> _A	Avrami kinetic constant
k_b	Barrier transport coefficient
<i>k</i> _{ext}	Column external air conductivity
k_{f}	External film mass transfer coefficient
k_F	Pseudo-first-order kinetic rate constant
k_g	Gas phase thermal conductivity
k_w	Column wall conductivity
Κ	Linear driving force rate coefficient
K_s	Pseudo-second-order kinetic rate constant
L	Bed length
LDF	Linear driving force model
L _{MTZ}	Length of the Mass Transfer Zone
LUB	Length of Unused Bed
М	Molecular weight
MFC	Mass Flow Controller

MRA	Multiple Regression Analysis
OFAT	One Factor At Time
OXA	Pre-oxidized GAC
OXA-GAC	Optimal adsorbent
р	Partial pressure
Р	Total pressure
P/P_0	Relative pressure
Pr	Prandtl number
q	Adsorbed concentration
\overline{q}	Average adsorbed phase concentration in the micropore
=	Average concentration in adsorbent particle
ч «*	Adsorbed phase concentration in equilibrium with the fluid phase
q	concentration
q_m	Maximum loading capacity
Δq	Normalized standard deviation
r	Distance along the microparticle radius
<i>r</i> _p	Mean macropore radius
R^2	Regression coefficient
R	Universal gas constant
Ra	Rayleigh number
R_c	Microparticle radius
Re	Reynolds
R_p	Macroparticle radius
RSM	Response Surface Methodology
Sc	Schmidt

SS	Sum of Squares
STP	Standard Temperature and Pressure
t	Time
TC	Thermocouple
TGA	Thermo Gravimetric Analyser
U	Fluid velocity
U	External overall heat transfer coefficient
У	Mole fraction
Z.	Distance along the bed length
3D	Three Dimensional
Еь	Bed void fraction
\mathcal{E}_p	Adsorbent porosity
$ ho_p$	Particle density
$ ho_{g}$	Gas phase density
$ ho_w$	Column wall density
μт	Micrometre
σ^2	Residual Mean Square
τ	Pore tortuosity factor
λ_L	Effective axial heat dispersion
ν	Kinematic viscosity
α	Thermal diffusivity
μ	Gas mixture viscosity
Φ	Column efficiency

CHAPTER 1: INTRODUCTION

1.1 Background

Global warming and related environmental damage associated with emissions of carbon dioxide (CO₂), the most significant greenhouse gas, have long been recognized to represent a potential serious threat to the future of the earth's environment (Bezerra, Oliveira, Vieira, Cavalcante, & Azevedo, 2011; Jadhav et al., 2007). The primary source of anthropogenic CO₂ emissions is the combustion of fossil fuels such as coal or natural gas for the production of electricity. Because fossil fuels are likely to remain the predominant energy source all over the world, CO₂ emissions must be reduced to mitigate the unfettered release of this greenhouse gas into the atmosphere (Garcés, Villarroel-Rocha, Sapag, Korili, & Gil, 2013; Siriwardane, Shen, Fisher, & Poston, 2001). As a result, various CO_2 separation techniques, such as liquid solvent absorption, membrane separation, cryogenic separation, and adsorption processes including pressure/vacuum swing adsorption (PSA/VSA) and temperature swing adsorption (TSA), are currently under investigation (Gomes & Yee, 2002; Xu, Song, Andresen, Miller, & Scaroni, 2002). Currently, absorption with amine-based absorbents is the preferred technology for the large-scale separation of CO₂ from the flue-gas streams of fossil-fuel-based power plants. However, this method suffers from several significant drawbacks that impede its implementation, including low efficiency, high energy consumption during the regeneration process, a high equipment corrosion rate, oxidative degradation of the amines, and flow problems caused by viscosity (Rinker, Ashour, & Sandall, 2000; Veawab, Tontiwachwuthikul, & Chakma, 1999). The development of alternative, lower lower-cost, energy-efficient CO₂ removal technologies is therefore important. The separation and purification of gas mixtures by adsorption is a potential option because of its ease of operation, high adsorption capacity, minimal environmental impact, low cost, and efficient recovery of the solute compared to conventional absorption with liquid solvents (Cavenati, Grande, & Rodrigues, 2004; Serna-Guerrero, Belmabkhout, & Sayari, 2010a). Particularly, pressure-swing adsorption (PSA) has a number of attractive characteristics, such as its applicability over a relatively wide range of temperature and pressure conditions, its low energy requirements, and its low capital investment costs (Delgado, Uguina, Sotelo, Ruiz, & Gomez, 2006; Mutasim & Bowen, 1991). The last three decades have seen a tremendous growth in research into and commercial applications of CO₂ removal from various flue-gas mixtures by PSA processes (Siriwardane et al., 2001; Yong, Mata, & Rodrigues, 2002).

The development of an easily regenerated and durable adsorbent with fast adsorption/desorption kinetics, a high selectivity and a high adsorption capacity will undoubtedly enhance the competitiveness of adsorptive separation for CO₂ capture in flue-gas applications (Lu et al., 2009; Su, Lu, Kuo, & Zeng, 2010). Among all adsorbents, activated carbon offers several advantages as a CO₂ adsorbent: an inherent affinity for CO₂, an easy-to-design pore structure, insensitivity to moisture, ease of regeneration, stability over a large number of cycles, and an appealing low cost (Bezerra et al., 2011; Sjostrom & Krutka, 2010). Most of the processes that produce CO₂ in product streams occur at elevated temperatures (up to 100 °C), and the stream must be cooled before separation takes place. Therefore, developing an adsorbent with a high adsorption capacity at relatively high temperatures can drastically reduce the cooling cost of separation and make CO₂ capture from power plants feasible (Maroto-Valer, Tang, & Zhang, 2005; Xu et al., 2002).

The CO₂ adsorption performance of activated carbon is well known to be strongly influenced by modification of the surface chemical properties of the activated carbon (Arenillas, Smith, Drage, & Snape, 2005; M. G. Plaza et al., 2009; Shafeeyan, Daud,

Houshmand, & Shamiri, 2010). Inspired by the current liquid-phase amine scrubbing technology, researchers have incorporated different basic nitrogen functional groups onto the carbon surface for CO₂ removal from gaseous mixtures at relatively high temperatures (Przepiórski, Skrodzewicz, & Morawski, 2004; Zhijuan Zhang, Xu, Wang, & Li, 2010). This approach is expected to exploit the strong chemical interactions between CO₂ and the attached basic nitrogen functionalities on the surface, as well as the low energy requirements, to regenerate the solid adsorbent (Drage et al., 2007; Knowles, Delaney, & Chaffee, 2006; Maroto-Valer et al., 2005; Plaza, Pevida, Arenillas, Rubiera, & Pis, 2007). Several authors have proposed modifying activated carbon with gaseous ammonia in the presence or absence of oxygen as a suitable technique to produce efficient CO₂ adsorbents that maintain high uptakes despite moderately high temperatures (Pevida, Plaza, et al., 2008; Plaza et al., 2009; Plaza, Rubiera, Pis, & Pevida, 2010).

1.2 Research objectives

The purpose of this research is to investigate the fixed-bed adsorption of CO_2 onto ammonia-modified activated carbon. More specifically, the objectives of this study are:

- Optimize the amination conditions of activated carbon adsorbents in an effort to maximize their CO₂ adsorption/desorption capacities.
- Study the adsorption equilibrium of carbon dioxide onto the ammonia modified and untreated activated carbon adsorbents and develop an appropriate semiempirical isotherm model.
- Study the kinetics of CO₂ adsorption on the ammonia modified and untreated activated carbon adsorbents at different temperatures and develop an appropriate kinetic model.

• Develop a model that consists of an appropriate equilibrium and kinetics equations to predict the breakthrough behavior of the fixed-bed adsorption of CO₂.

1.3 Thesis organization

This thesis consists of five chapters dealing with different aspects related to the topic of research.

- CHAPTER 1: This chapter briefly introduces the importance of CO₂ capture and sequestration from point source emissions as a potential way to mitigate unfettered release of greenhouse gases into the atmosphere. The objectives of the study are also presented.
- CHAPTER 2: This chapter presents a review of efforts over the last three decades toward mathematical modeling of the fixed-bed adsorption of carbon dioxide. The nature of various gas-solid equilibrium relationships as well as different descriptions of the mass transfer mechanisms within the adsorbent particle are reviewed. In addition to mass transfer, other aspects of adsorption in a fixed bed, such as heat and momentum transfer, are also studied. Both single- and multi-component CO₂ adsorption systems are discussed in the review.
- CHAPTER 3: This chapter explains all the experiments procedures for the modification and characterization of activated carbon samples. Details on the raw material, equipment, and other related procedures are also presented.
- CHAPTER 4: This chapter presents results and data obtained from laboratory experiments. In this chapter the results are presented in three parts. Part 1 investigates the application of response surface methodology in predicting and

optimizing the amination conditions of activated carbon adsorbent toward CO_2 adsorption. Part 2 investigates the adsorption equilibrium of carbon dioxide onto the ammonia modified and untreated activated carbon adsorbents and develops a semi-empirical equilibrium model able to distinguish the contributions of physical and chemical adsorption to the total CO_2 uptake. Part 3 studies the kinetics of CO_2 adsorption on the ammonia modified and untreated activated carbon adsorbents. To predict the breakthrough behavior of the fixed-bed adsorption of CO_2 , the focus of Part 3 is to develop a dynamic model that consists of an Avrami equation to describe the kinetics of adsorption and a semi-empirical Toth equation to represent the gas–solid equilibrium isotherm. In addition, a comprehensive discussions and explanations on experimental results are presented.

• CHAPTER 5: The conclusions constructed from the results and discussion chapter are explained part by part. The recommendations and suggestions for future works are also presented.

CHAPTER 2: LITERATURE REVIEW

2.1 Introduction

Concerns over the gradual increase in the atmospheric concentration of CO₂ and its impact on climate change have prompted a global research effort to capture CO₂ from point source emissions and stabilize its concentration in the atmosphere (Gomes & Yee, 2002; Grande & Rodrigues, 2008; Plaza et al., 2007). The most important sources of CO₂ emissions are power plants that generate electricity from fossil fuels (coal, oil, and natural gas) (Dantas et al., 2011; Grande, Lopes, Ribeiro, Loureiro, & Rodrigues, 2008; Grande & Rodrigues, 2008; Kikkinides, Yang, & Cho, 1993; Mulgundmath, Jones, Tezel, & Thibault, 2012; Park, Beum, Kim, & Cho, 2002). Therefore, it is critical to separate and recover carbon dioxide from the flue gases emitted by power plants to avoid excess CO₂ emissions (Chou & Chen, 2004; Ko, Siriwardane, & Biegler, 2005; Mulgundmath et al., 2012). Various separation techniques, such as liquid solvent absorption, membrane separation, cryogenic techniques, and adsorption over solid sorbents, are increasingly used to reduce CO₂ emissions (Gomes & Yee, 2002; Takamura, Narita, Aoki, Hironaka, & Uchida, 2001). At present, the most widely used technology for the removal of CO_2 from gaseous mixtures is amine absorption (Delgado et al., 2006; Leci, 1996). However, this process is energy-intensive during the regeneration of solvent and is also plagued by extensive corrosion of the process equipment (Chue, Kim, Yoo, Cho, & Yang, 1995; Gray et al., 2004; Gray et al., 2005; Ko et al., 2005). It is therefore important to explore economical and energy-efficient alternative approaches for CO₂ separation (Grande et al., 2008; Xu, Song, Miller, & Scaroni, 2005).

Recently, it was reported that the cost associated with CO₂ capture can be reduced below the cost of conventional absorption with liquid solvents by using adsorption separation technologies (Ho, Allinson, & Wiley, 2008; Radosz, Hu, Krutkramelis, & Shen, 2008). Several technological advances in the field of CO₂ capture by adsorption have been developed around the world, demonstrating the attractiveness of this technique for postcombustion treatment of flue gas (Dantas et al., 2011; Dantas et al., 2011; Grande et al., 2008). Two main adsorption technologies are viewed as feasible for CO₂ separation and purification on a large scale: pressure/vacuum swing adsorption (PSA/VSA) and temperature swing adsorption (TSA) (Chue et al., 1995; Clausse, Bonjour, & Meunier, 2004; Plaza et al., 2009; Plaza, Garcia, Rubiera, Pis, & Pevida, 2011). Recent developments have demonstrated that PSA is a promising option for separating CO₂ due to its ease of applicability over a relatively wide range of temperature and pressure conditions, its low energy requirements, and its low capital investment costs (Agarwal, Biegler, & Zitney, 2010b; Cen & Yang, 1985; Delgado et al., 2006; Gomes & Yee, 2002). Many studies concerning CO₂ removal from various flue gas mixtures by means of PSA processes have been addressed in the literature (Agarwal et al., 2010b; Chaffee et al., 2007; Chou & Chen, 2004; Chue et al., 1995; Grande et al., 2008; Ho et al., 2008; Kikkinides et al., 1993; Ko, Siriwardane, & Biegler, 2003; Mulgundmath et al., 2012; Na, Koo, Eum, Lee, & Song, 2001; Reynolds, Ebner, & Ritter, 2005; Sircar & Kratz, 1988; Xiao et al., 2008). Prior to the design of an adsorption process, selecting an appropriate adsorbent with high selectivity and working capacity, as well as a strong desorption capability, is key to separating CO₂. As a result, a wide variety of adsorbents, such as activated carbons, synthetic zeolites, carbon molecular sieves, silicas, and metal oxides, have been investigated in recent years for this purpose (Chue et al., 1995; Dantas et al., 2011; Dantas et al., 2011; Moreira, Soares, Casarin, & Rodrigues, 2006; Plaza et al., 2011; Xu et al., 2005).

The design of an appropriate adsorption process requires the development of a model that can describe the dynamics of adsorption on a fixed bed with the selected adsorbent (Dantas et al., 2011; Dantas et al., 2011; Delgado, Uguina, Sotelo, & Ruiz, 2006; Lua & Yang, 2009). The absence of an accurate and efficient adsorption cycle simulator necessitates the use of data from experimental units to develop new processes. This empirical design of an adsorption column through extensive experimentation on process development units tends to be expensive and time consuming (Siahpoosh, Fatemi, & Vatani, 2009). A predictive model using independently established equilibrium and kinetic parameters may provide, in principle, a method of estimating the column dynamic capacity without extensive experimentation. A fixed-bed column mathematical simulation that considers all relevant transport phenomena is therefore required to obtain a better understanding of the behavior of new adsorbents during the adsorption/desorption cycles and for optimization purposes. Moreover, these models are capable of estimating the breakthrough curve and temperature profile for a certain constituent in the bulk gas at all locations within the packed column. This experimentally verified model is then used to conduct an extensive study to understand the effects of various process parameters on the performance of the PSA cycle. These are the main reasons why the mathematical modeling of adsorption processes has attracted a great deal of attention among researchers.

In general, prediction of column dynamics behavior requires the simultaneous solution of a set of coupled partial differential equations (PDEs) representing material, energy, and momentum balances over a fixed-bed with the appropriate boundary conditions (Hwang, Jun, & Lee, 1995). Because the simultaneous solution of a system of PDEs is tedious and time consuming, the use of simplified models capable of satisfactorily predicting fixedbed behavior is desirable. Many attempts have been made to evaluate and develop simplifying assumptions to decrease computational time and facilitate optimization studies. A review of the literature reveals the development of simplifying assumptions mainly on the representation of mass transfer phenomena within the adsorbent particles as an alternative pathway to simplify fixed-bed adsorption calculations. Modeling and optimization of the fixed-bed adsorption of CO_2 has developed over the past three decades and is still of great interest to investigators. This review presents a fairly extensive survey of previous studies on the mathematical modeling of the CO_2 adsorption process in a packed column. Various models for gas-solid adsorption equilibria as well as different descriptions of the mass transfer mechanisms within the adsorbent particle are reviewed. In addition to concentration variation, other aspects of adsorption in a fixed bed, such as temperature and pressure variations, are also studied. The purpose of this study was to investigate the mathematical models capable of simulating the dynamic behavior of the fixed-bed adsorption of carbon dioxide.

2.2 Overview of the prediction of adsorption column dynamics

In most adsorption processes, the adsorbent is in contact with a fluid in a packed bed. An understanding of the dynamics behavior of such systems is therefore required for rational process design and optimization (Rutherford & Do, 2000a). The dynamics behavior of an adsorption column system can be classified based on the nature of the gas-solid equilibrium relationship of fluid constituents and the complexity of the mathematical model required for describing the mechanism by which the mass transfer from the fluid to the solid phase occurs (Ruthven, 1984). The gas-solid adsorption equilibrium indicates the limiting capacity for solute separation from the gas phase into the solid phase. It is the most important process that controls the dynamics behavior of a packed column so that the general nature of a mass transfer zone is determined entirely by the equilibrium isotherm. Therefore, due to variations in the composition/temperature with respect to time and location within the adsorption column and the consequent effects on the adsorption equilibrium relation, a comprehensive gas-solid equilibrium model is needed. Several authors have reported experimental evidence of these effects in a column packed with microporous adsorbents (Carta, 2003). The complexity of the mathematical model, in

turn, depends on the concentration level, the choice of rate equation, and the choice of flow model (Ruthven, 1984). In addition, temperature changes may also affect the concentration profiles, particularly for high-concentration feeds in which the heat of adsorption generates thermal waves in both axial and radial directions. Therefore, apart from the mass transfer effects on adsorption rate, the effects of heat generation and heat transfer in the adsorbent bed must also be considered (Rezaei & Grahn, 2012). Moreover, the axial pressure along the bed may not be constant. As a consequence, a momentum balance also has to be included in the model. Table 2.1 provides a comprehensive classification scheme of the summary of the fixed-bed column mathematical models for carbon dioxide adsorption developed over the last three decades. All of the models assume that the gas phase follows the ideal gas law. The flow pattern is described by the plug flow or axially dispersed plug-flow model. It is further assumed that the radial gradients of concentration and, where applicable, temperature and pressure are negligible (with the exception of models 4 and 20). The assumption that the radial gradient is negligible has been widely accepted in many other studies (Jee, Park, Haam, & Lee, 2002; Kim, Bae, Choi, & Lee, 2006; Kim, Moon, Lee, Ahn, & Cho, 2004). The majority of the models reviewed here include the effects of the finite mass transfer rate, resulting in a theoretical representation that more closely approaches a real process. Most of the aforementioned models use a linear driving force approximation to describe the gas-solid mass transfer mechanism. Some of these models consider the effects of heat generation and heat transfer in the adsorbent bed, which may affect the adsorption rates. Moreover, in modeling the non-isothermal operation of adsorption processes occurring in packed beds, it is also commonly assumed that the heat transfer resistance between the gas and the solid phases is negligible and that they reach thermal equilibrium instantaneously. With the exception of models 15-17, 21, 24, 26-27, 29-30, 31, and 33, the pressure drop across the adsorbent bed is neglected, and the column is assumed to operate at constant

pressure. Most of the adsorption equilibrium is described using non-linear isotherms such as the Langmuir isotherm or a hybrid Langmuir-Freundlich isotherm; only rarely have linear isotherms been used.

			Model assumptions						
No.	Equilibrium relationship	Flow pattern	Mass transfer rate model	Heat effects	Others	Application	Solution method	Results and comments	Ref.
1	Linear	Plug	Local	Isothermal	No radial	PSA separation of	Analytical results	The model provided a qualitative or semi	(Shenda
	equilibrium	flow	equilibrium		variation in	carbon dioxide	from a linear	quantitative process description. Due to	lman &
	isotherm		model		concentration.	from a He-CO ₂	mathematical	neglecting the effects of mass transfer	Mitchell
					Negligible	mixture using silica	model obtained by	resistance some of the detailed behavior	, 1972)
					pressure drop.	gel	the method of	differed from experimental results.	
					Trace system [*] .		characteristics.		
2	A hybrid	Plug	Local	Non-	No radial	Separation of coal	The model was	Poor comparison with experimental data for	(Cen &
2	A hybrid Langmuir-	Plug flow	Local equilibrium/	Non- isothermal	No radial variations in	Separation of coal gasification	The model was solved using an	Poor comparison with experimental data for the predictive equilibrium model. The major	(Cen & Yang,
2	A hybrid Langmuir- Freundlich	Plug flow	Local equilibrium/ Linear driving	Non- isothermal	No radial variations in concentration and	Separation of coal gasification products containing	The model was solved using an implicit finite	Poor comparison with experimental data for the predictive equilibrium model. The major discrepancy was in the CO ₂ concentration.	(Cen & Yang, 1985)
2	A hybrid Langmuir- Freundlich isotherm	Plug flow	Local equilibrium/ Linear driving force (LDF)	Non- isothermal	No radial variations in concentration and temperature.	Separation of coal gasification products containing H ₂ , CO, CH ₄ , H ₂ S,	The model was solved using an implicit finite difference method	Poor comparison with experimental data for the predictive equilibrium model. The major discrepancy was in the CO ₂ concentration. The results of the LDF model were in fair	(Cen & Yang, 1985)
2	A hybrid Langmuir- Freundlich isotherm	Plug flow	Local equilibrium/ Linear driving force (LDF) approximation	Non- isothermal	No radial variations in concentration and temperature. Thermal	Separation of coal gasification products containing H ₂ , CO, CH ₄ , H ₂ S, and CO ₂ by PSA	The model was solved using an implicit finite difference method which was stable	Poor comparison with experimental data for the predictive equilibrium model. The major discrepancy was in the CO ₂ concentration. The results of the LDF model were in fair agreement with the experimental data. Mass	(Cen & Yang, 1985)
2	A hybrid Langmuir- Freundlich isotherm	Plug flow	Local equilibrium/ Linear driving force (LDF) approximation model	Non- isothermal	No radial variations in concentration and temperature. Thermal equilibrium	Separation of coal gasification products containing H ₂ , CO, CH ₄ , H ₂ S, and CO ₂ by PSA using activated	The model was solved using an implicit finite difference method which was stable and convergent.	Poor comparison with experimental data for the predictive equilibrium model. The major discrepancy was in the CO ₂ concentration. The results of the LDF model were in fair agreement with the experimental data. Mass transfer coefficient for CO ₂ was determined	(Cen & Yang, 1985)
2	A hybrid Langmuir- Freundlich isotherm	Plug flow	Local equilibrium/ Linear driving force (LDF) approximation model	Non- isothermal	No radial variations in concentration and temperature. Thermal equilibrium between the fluid	Separation of coal gasification products containing H ₂ , CO, CH ₄ , H ₂ S, and CO ₂ by PSA using activated carbon	The model was solved using an implicit finite difference method which was stable and convergent.	Poor comparison with experimental data for the predictive equilibrium model. The major discrepancy was in the CO ₂ concentration. The results of the LDF model were in fair agreement with the experimental data. Mass transfer coefficient for CO ₂ was determined empirically.	(Cen & Yang, 1985)

Table 2.1: Summary of the dynamics models for fixed-bed adsorption of CO_2

					Tuele 2.1, continued				
			Model						
			assumptions			_			
No.	Equilibrium	Flow	Mass transfer	Heat	Others	Application	Solution method	Results and	Ref.
	relationship	pattern	rate model	effects				comments	
3	Linear	Axial	LDF	Isotherm	Negligible radial gradient of	PSA separation of	The solution to the	The theoretical	(Raghava
	equilibrium	dispersed	approximation	al	concentration.	carbon dioxide	model equations	curves based on the	n, Hassan,
	isotherm	plug flow	with non-		Negligible pressure drop.	from a He-CO ₂	was obtained by	assumption of	&
			constant		Trace system.	mixture using silica	orthogonal	inverse dependence	Ruthven,
			coefficient			gel	collection and	of the mass transfer	1985)
							using finite	coefficient with the	
							difference methods	pressure provided a	
							with consistent	good representation	
							results.	of the experimental	
								results.	
4	Linear	Axial	Pore diffusion	Non-	Negligible radial concentration	Theoretical and	Analytical solution	The central-axis-	(Kaguei,
	equilibrium	dispersed	model	isotherm	gradient.	experimental	was performed in	thermal waves	Shemilt,
	isotherm	plug flow		al	Radial temperature profile in the	studies on the CO_2	the Laplace domain	measured at various	& Wakao,
					column/uniform temperature over the	capture in a column	under the condition	axial locations in the	1989;
					column cross-section. Negligible axial	packed with	of a semi-infinite	column were in	Kaguel,
					pressure gradient. Constant temperature	activated carbon	column.	good agreement	Yu, &
					of the column wall.	particles		with those predicted.	Wakao,
								-	1985)

'Table 2.1, continued'

			Model						
			assumptions						
No.	Equilibrium	Flow	Mass transfer	Heat	Others	Application	Solution method	Results and comments	Ref.
	relationship	pattern	rate model	effects					
5	A hybrid	Plug	Local	Non-	Negligible radial	Separation of gas	The models were solved	The Knudsen plus surface	(Doong &
	Langmuir-	flow	equilibrium	isothermal	gradients in	mixtures containing	numerically by	diffusion model provided the	Yang,
	Freundlich		model.		temperature and	CO_2 , CH_4 , and H_2	employing finite	best fit when compared to the	1986)
	isotherm		Pore/surface		concentrations.	(one-third each by	difference method.	experimental data. Due to the	
			diffusion models		Thermal	volume) by PSA		assumption of infinite pore	
					equilibrium	using activated		diffusion rate, the ILE model	
					between the fluid	carbon		predicted a later breakthrough	
					and particles.			plus a lower concentration for	
					Negligible			CO ₂ .	
					pressure drop in				
					the bed.				
6	Langmuir	Plug	LDF	Isothermal	Negligible radial	PSA separation of a	The model was solved	The model predictions were	(Kapoor
	isotherm	flow	approximation		concentration	CO ₂ (50%)-CH ₄	using an implicit	reasonable and the average	& Yang,
			model		gradient.	(50%) mixture using	backward finite	difference between the model	1989)
			with a cycle time		Negligible	a carbon molecular	difference scheme,	prediction and experimental	
			dependent		pressure drop.	sieve	which was both stable	result was within 3.0%.	
			coefficient				and convergent.		

'Table 2.1, continued'

			Model						
			assumptions						
No.	Equilibrium	Flow	Mass transfer	Heat	Others	Application	Solution method	Results and comments	Ref.
	relationship	pattern	rate model	effects					
7	Langmuir	Plug	Local	Non-	No radial variations in	Separation of	A set of PDEs was reduced	The adiabatic simulation of the	(R.
	isotherm	flow	equilibrium	isothermal	concentration and	carbon dioxide	to ODEs and solved by	blowdown step showed that an	Kuma
			model	(adiabatic)	temperature.	from binary	using the numerical	isothermality assumption is	r,
					Thermal equilibrium	gas mixtures	technique of finite	inadequate for process design.	1989)
					between the fluid and	(CO ₂ /N ₂ ,	differences.	However, it could be an excellent	
					particles.	CO ₂ /CH ₄ , and		tool for predicting the column	
					Negligible pressure drop.	CO ₂ /H ₂) using		behavior and trends in a semi	
						BPL carbon		quantitative manner.	
						and 5A zeolite			
8	Langmuir	Plug	LDF	Non-	Negligible radial	CO ₂ capture	The nonlinear rate equations	A comparison of experimental	(Muta
	isotherm	flow	approximation	isothermal	temperature and	from a mixture	were solved using Runge-	breakthrough and temperature	sim &
			Model		concentration gradients/	of N ₂ (90%)-	Kutta-Merson method.	profiles with model predictions	Bowe
					Thermal equilibrium	CO ₂ (10%) by	Adsorbate concentration and	revealed that the model	n,
					between the gas and	PSA using 5A	temperature profiles were	reproduced the experimental data	1991)
					solid phases.	molecular	predicted using an implicit	satisfactorily, which indicates that	
					Negligible pressure drop	sieve	backward difference	the assumptions the model is	
					through the bed.		approximation.	based on are valid for this system.	

'Table 2.1, continued'

			Model						
			assumptions						
No.	Equilibrium	Flow	Mass transfer	Heat	Others	Application	Solution method	Results and comments	Ref.
	relationship	pattern	rate model	effects					
9	Langmuir	Axial	LDF	Isothermal	No radial	Investigation	A set of PDEs was solved by the	The experimental adsorption and	(Hwang
	isotherm	dispersed	approximation		variations in	of adsorption	method of orthogonal collection.	desorption curves were predicted	& Lee,
		plug flow	model		concentration.	and	The resulting set of ODEs was	fairly well by the LDF model and	1994)
					Negligible	desorption	solved numerically in the time	the pressure dependent mass	
					pressure gradient	breakthrough	domain by using DGEAR of the	transfer coefficients calculated	
					across the bed.	behaviors of	International Mathematical and	from a single component system	
						CO and CO ₂	Statistical Library (IMSL) which	provided a reasonably good	
						on activated	employs Gear's stiff method	representation of adsorption and	
						carbon	with variable order and step size.	desorption data for a multi	
								component system.	
10	Langmuir	Plug	LDF	Non-	No radial	Fixed-bed	A set of differential equation	A comparison between	(Kim,
	isotherm/	flow	approximation	isothermal	concentration	adsorption of	with the initial and boundary	concentration and temperature	Chue,
	Ideal		model		and temperature	a N ₂ (85%)-	conditions were solved by using	history curves with theoretical	Kim,
	adsorbed				gradients.	CO ₂ (15%)	the solver LSODA.	results revealed that the presented	Cho, &
	solution				Negligible axial	mixture using		model could predict the dynamic	Kim,
	theory				pressure	a of X-type		behavior of the adsorption bed,	1994)
	(IAST)				gradient.	zeolite		even though a slight deviation was	
								observed after the maximum point.	

'Table 2.1, continued'
			Model						
			assumptions						
No.	Equilibrium	Flow	Mass transfer	Heat	Others	Applicati	Solution method	Results and comments	Ref.
	relationship	pattern	rate model	effects		on			
11	Langmuir	Plug	LDF	Non-	Negligible radial	Fixed-bed	A set of PDEs was solved by	The model provided a good	(Hwa
	isotherm	flow	approximation	adiabat	velocity, temperature,	adsorption	the numerical method of lines.	representation of the experimental	ng et
			model	ic,	and concentration	of carbon	The resulting set of ODEs was	breakthrough and temperature curves.	al.,
			with lumped	adiabat	gradients.	dioxide	solved by using the subroutine	Since the mass transfer coefficients	1995)
			mass transfer	ic, and	Negligible pressure	(with	DIVPAG of the IMSL library,	were determined by fitting the	
			coefficient	isother	gradient across the bed.	helium as	while the non-linear algebraic	experimental data, the disadvantage of	
				mal		the carrier	equation was solved by using	this model is the determination of a	
						gas) on	the subroutine DNEQNF of the	new value for the effective mass	
						activated	same library.	transfer coefficient for each run.	
						carbon			
12	Extended	Plug	LDF	Non-	Negligible radial	Separation	A set of PDAEs representing	The predicted values matched	(Yang
	Langmuir-	flow	approximation	isother	gradients in temperature	of a binary	the packed column were solved	significantly with the experimental	, Han,
	Freundlich		model with a	mal	and concentrations.	mixture H ₂	by a flux corrected third-order	results at shorter adsorption time. The	Cho,
	isotherm		single lumped		Thermal equilibrium	(70%)-	upwind method. Numerical	errors at longer adsorption time was	Lee,
			mass transfer		between the fluid and	CO_2	oscillation, which often appears	attributed to a partial breakthrough of	&
			coefficient		particles. Negligible	(30%) by	when a convection equation is	mass transfer zone during cocurrent	Lee,
					pressure drop in the	PSA using	solved, is eliminated by the flux	depressurization and/or blowdown/	1995)
					bed.	zeolite 5A	corrected scheme.	purge steps.	

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			Model						
			assumptions						
No.	Equilibrium	Flow	Mass transfer	Heat	Others	Application	Solution method	Results and comments	Ref.
	relationship	pattern	rate model	effects					
13	Langmuir	Plug	LDF	Isother	No radial variation	Removal and	A set of equations	Good agreement between the model and	(Diagn
	isotherm	flow	approximation	mal	in concentration.	concentration of CO_2	describing the system	the experimental results was obtained	e,
			model		Negligible pressure	dilute gas from air by	was solved by Euler's	particularly for values of between the	Goto,
					drop.	PSA using three	method.	ratio of feed/enriched product flow rates	&
					Trace system.	types of commercial		= 3 and 6. Also, at the point the ratio of	Hirose,
						molecular-sieve		feed/lean product flow rates <2, where	1996)
						zeolites (13X, 5A,		the solute balance was not satisfactory,	
						and 4A)		the agreement was not quite good.	
14	Langmuir	Plug	LDF	Non-	No radial	Layered-bed PSA	To solve a set of	The agreement between the	(Park,
	isotherm	flow	approximation	isother	concentration and	separation of a	coupled PDAEs, the	experimental and theoretical prediction	Kim, &
			model	mal	temperature	typical cracked gas	spatial derivatives	was good that indicate validity of the	Cho,
					gradients. Thermal	mixture (H ₂ , CH ₄ ,	were divided using a	present model. The overall mass transfer	2000;
					equilibrium between	CO, and CO ₂) using	backward difference	coefficients were obtained by fitting the	Park,
					the gas phase and	activated carbon	scheme, and the	breakthrough data. The experimental	Kim,
					adsorbents.	followed by zeolite	resulting ordinary	effluent temperature was much lower	Cho,
					Negligible axial	5A	differential equations	than the model prediction because the	Kim, &
					pressure gradient.		were solved with the	model did not take into account the heat	Yang,
							GEAR method.	loss to the end of the column.	1998)

'Table 2.1. continued'

					1 able 2.	1, continued			
			Model assumptions						
No.	Equilibrium relationship	Flow pattern	Mass transfer rate model	Heat effects	Others	Application	Solution method	Results and comments	Ref.
15	Extended	Axial	LDF	Non-	Negligible radial	Layered bed	A set of coupled PDEs	In spite of the frozen solid-phase	(Lee,
	Langmuir-	dispersed	approximation	isotherm	gradients of	PSA separation	was solved using a finite	model, the utilized LDF model could	Yang,
	Freundlich	plug flow	model with a	al	concentration,	of a coke oven	difference method. The	predict a transient variation of the	U,
	isotherm		single lumped		pressure and	gas containing	spatial dimension was	effluent stream during pressurization	&
			mass transfer		temperature.	H ₂ , CH ₄ , CO, N ₂ ,	discretized by using a	and depressurization steps and	Ahn,
			coefficient		Thermal equilibrium	and CO ₂ using	second-order central	simulated results of the dynamic	1999)
					between the fluid and	activated carbon	difference and a second-	model agreed well with the PSA	
					particles.	and zeolite 5A	order backward	experimental results. The	
					Pressure drop along		difference for the	experimental data resulted in slightly	
					the bed was		second-order and the	higher recovery than predicted (4%	
					calculated by the		first-order space	error).	
					Ergun equation.		derivatives, respectively.		
16	Langmuir–	Axial	LDF	Non-	No radial variations	Packed bed	A set of partial	The model provided a reasonable fit to	(Moh
	Freundlich	dispersed	approximation	isotherm	in temperature,	adsorption of	differential equations	experimental adsorption data.	amadi
	isotherm	plug flow	model	al	pressure, and	carbon dioxide,	was solved using finite	However, comparing the experimental	nejad,
					concentration.	nitrogen, and	differences and	data with the model prediction	Knox,
					The Ergun equation	water on	Newmans's method.	suggested that a 2D model is required	&
					was used to estimate	molecular sieve		for accurate simulation of the average	Smith,
					the pressure drop.	5A		column breakthrough concentration.	2000)

					,			
		Model						
		assumptions			_			
Equilibrium	Flow	Mass transfer	Heat	Others	Application	Solution method	Results and comments	Ref.
relationship	pattern	rate model	effects					
Langmuir	Axial	Local	Non-	No radial	High temperature	The equations were	The LDF model was found to	(Ding &
isotherm	dispersed	equilibrium	isotherm	concentration,	carbon dioxide	solved in the gPROMS	give a good description of the	Alpay,
	plug flow	model/	al	pressure and	adsorption on	modelling environment.	adsorption and desorption data	2000)
		LDF model		temperature gradients.	hydrotalcite	The spatial discretization	especially for high feed CO ₂	
		based on pore		Pressure distribution	adsorbent	method of orthogonal	concentrations. The ILE model	
		diffusion		was described by the		collocation on finite	failed to give an adequate	
				Ergun equation.		elements was employed.	description of the desorption	
				Thermal equilibrium			kinetics.	
				between the gas and				
				particles.				
Langmuir	Plug	LDF	Isotherm	Negligible radial	CO ₂ recovery	A set of equations	Both simulation and	(Takam
isotherm	flow	approximation	al	gradient of	from boiler	describing the system was	experimental results showed the	ura et
		Model		concentration.	exhaust gas	discretized in space and	same trend of the recovery	al.,
				Negligible pressure	(containing CO ₂	the resulting set of	efficiency and the CO ₂	2001)
				drop.	and N_2) by PSA	ordinary differential	concentration of the recovery	
					using Na-X and	equations with a variable	gas with respect to the variation	
					Na-A type	time step was solved.	of the feed gas flow rate.	
					zeolites			
	Equilibrium relationship Langmuir isotherm	Equilibrium relationshipFlow patternLangmuirAxialisothermdispersed plug flowLangmuirPlug flow	KodelEquilibriumFlowMass transferrelationshippatternrate modelLangmuirAxialLocalisothermdispersedequilibriumplug flowmodel/LDF modelLDF modelbased on porediffusionItangmuirPlugLDFLangmuirFlowApproximationModelModelApproximation	Model assumptionsEquilibriumFlowMass transferHeatrelationshippatternrate modeleffectsLangmuirAxialLocalNon-isothermdispersedequilibriumisothermplug flowmodel/alLDF modelLangmuirLDF modelbased on porediffusionLangmuirPlugLDFIsothermIsothermflowapproximationalModelIsothermflowapproximation	Model assumptionsEquilibriumFlowMass transferHeatOthersrelationshippatternrate modeleffectsLangmuirAxialLocalNon-No radialisothermdispersedequilibriumisothermconcentration,plug flowmodel/alpressure andLDF modelLDF modeltemperature gradients.based on porebased on porePressure distributiondiffusionThermal equilibriumbetween the gas and particles.LangmuirPlugLDFIsothermNegligible radialisothermflowapproximationalgradient of concentration.LangmuirPlugLDFIsothermNegligible pressure drop.	Model assumptions Model assumptions Equilibrium relationship Flow pattern Mass transfer rate model Heat effects Others Application Langmuir Axial Local Non- No radial High temperature isotherm dispersed equilibrium isotherm concentration, carbon dioxide plug flow model/ al pressure and adsorption on LDF model LDF model Pressure distribution adsorbent diffusion Was described by the Ergun equation. Thermal equilibrium between the gas and particles. Def Isotherm Negligible radial CO2 recovery isotherm flow approximation al gradient of from boiler Langmuir Plug LDF Isotherm Negligible pressure (containing CO2 drop. and N2) by PSA using Na-X and Na-A type zeolites	Model assumptions Model assumptions Equilibrium relationship Flow pattern Mass transfer rate model Heat effects Others Application Solution method Langmuir Axial Local Non- No radial High temperature carbon dioxide The equations were isotherm dispersed equilibrium isotherm concentration, plug flow model/ al pressure and pressure and adsorption on adsorption on modelling environment. LDF model/ al pressure distribution was described by the Ergun equation. the spatial discretization adsorbent method of orthogonal collocation on finite Langmuir Plug LDF Isotherm Negligible radial particles. CO2 recovery A set of equations Langmuir Plug LDF Isotherm Negligible radial gradient of drop. from boiler describing the system was discretized in space and hoged is resulting set of drop. Model K K K K equations with a variable ha-A type	Model assumptions Model assumptions Equilibrium relationship pattern Flow rate model Mass transfer effects Heat effects Others Application Solution method Results and comments Langmuir Axial Local Non- No radial High temperature on concentration, The equations were solved in the gPROMS The LDF model was found to give a good description of the adsorption on plug flow model/ al pressure and temperature gradients. hydrotalcite The spatial discretization especially for high feed CO ₂ concentrations. The ILE model based on pore Pressure distribution adsorbent method of orthogonal collocation on finite failed to give an adequate effusion Figun equation. Thermal equilibrium between the gas and particles. elements was employed. description of the desorption kinetics. Langmuir Plug LDF Isotherm Negligible radial CO ₂ recovery A set of equations experimental results showed the concentration. isotherm flow approximation al gradient of from. containing CO ₂ drop. containing CO ₂ ortinary differential concentration fithe resulting se

					Table 2.1	, continued			
			Model						
No.	Equilibrium relationship	Flow pattern	Mass transfer rate model	Heat effects	Others	Application	Solution method	Results and comments	Ref.
19	Extended	Plug	LDF	Non-	Negligible gradients in	CO ₂ recovery	A MATLAB function	The analysis of breakthrough curves	(Choi et
	Langmuir	flow	approximation	isothermal	radial concentration	from a flue gas	based on sequential	showed good agreement with	al.,
	isotherm		model	(adiabatic)	and temperature.	(containing	quadratic programming	simulation data. However, analysis	2003)
					Negligible pressure	83% N ₂ , 13%	(SQP) method was used	of temperature changes in the	
					drop.	CO ₂ , and 4%	to solve the constrained	adsorption beds revealed some	
						O ₂) by PSA	nonlinear programming	discrepancy between simulations	
						using zeolite	optimization problem.	and experiments.	
						13X			
20	Ideal	Non-	LDF	Non-	Variations in	Carbon dioxide	A set of coupled PDEs	The model prediction of	(Moha
	adsorption	Darcian	approximation	isothermal	temperature,	adsorption from	was discretized by first-	breakthrough curves definitely	madinej
	solution	flow	model		concentration, and	a mixture of	or second-order	matched the obtained experimental	ad,
	theory	model			velocity along the	$(CO_2, N_2, and$	differences in time and	data. The temperature profile results	Knox,
	(IAST)	(2D			radial direction of	H ₂ O) in a	spatial dimensions. The	of 2D model also estimated the	&
		flow)			column.	column packed	set of discretized finite	experimental data fairly well. The	Smith,
						with zeolite 5A	difference equations was	few degree discrepancy between the	2003)
							solved simultaneously by	model and experimental data was	
							the implicit method of	attributed to prediction of heat	
							Newman.	transfer coefficients.	

			Model						
			assumptions						
No.	Equilibrium	Flow	Mass transfer	Heat	Others	Application	Solution method	Results and comments	Ref.
	relationship	pattern	rate model	effects					
21	O'Brien-	Axial	LDF	Adiabatic,	Negligible radial and	Adsorption of	To solve a set of differential	For low mass transfer	(Clausse
	Myers	dispersed	approximation	near-	angular gradients in	a 30% CO ₂ -	equations, the numerical method	coefficients (<0.01) the	et al.,
	isotherm	plug flow	model with an	adiabatic,	concentration,	10% C ₂ H ₆	of lines was retained. For the	numerical	2004)
			adjustable	and	temperature and	mixture in	spatial discretization, a finite	breakthrough curves	
			mass transfer	isothermal	velocity.	nitrogen (inert	volumes scheme with two	were too dispersive	
			coefficient		Thermal equilibrium	carrier gas) by	staggered grids was chosen: one	when compared to the	
					between the gas and	TSA using	for the velocity and one for the	experimental ones. For	
					the adsorbent.	Ambersorb	temperature, pressure and	higher values, the	
					The momentum	600 adsorbent	concentrations (gas and adsorbed	numerical curves were	
					equation represented		phases). The set of PDAEs	identical the ones to	
					by Ergun's equation.		obtained was integrated by	the others and agreed	
							employing an integrator	well with the	
							(DASPK 2.0).	experimental	
								measurements.	

					ʻTab	ole 2.1, continue	d'		
			Model assumptions						
No.	Equilibrium relationship	Flow pattern	Mass transfer rate model	Heat effects	Others	Application	Solution method	Results and comments	Ref.
22	Extended	Axial	Local	Non-	The	CO ₂	The PDEs were converted to ODEs by	Comparison of the simulation	(Chou
	Langmuir	dispersed	equilibrium0	isotherm	concentration	removal	the method of lines with adaptive grid	results of the CO ₂ concentration	& Chen,
	isotherm	plug flow	model	al	and temperature	from a flue	points. The estimation of the spatial	and the CO ₂ recovery with	2004)
					gradients in the	gas	derivatives was made from the upwind	experimental results indicated	
					radial direction	(containing	difference, and the cubic spline	that the data trends were similar	
					and the pressure	$80\% N_2$ and	approximation was used to estimate	although the simulation values	
					drop in the bed	20% CO ₂)	the flow rates in the adsorptive bed.	were slightly lower than the	
					were neglected.	by VSA	The concentration, temperature, and	experimental data. The	
					Instantaneous	using zeolite	adsorption quantity were integrated	difference was attributed to the	
					thermal	13X	with respect to the time in the bed by	application of an imprecise	
					equilibrium		LSODE of ODEPACK software.	adsorptive isotherm.	
					between the solid				
					and gas phases.				

			Model assumptions						
No.	Equilibrium	Flow	Mass transfer	Heat	Others	Application	Solution method	Results and comments	Ref.
	relationship	pattern	rate model	effects					
23	Langmuir	Axial	LDF	Isothermal	No radial variations	Prediction of the	The system of partial	The model that included the detailed	(Ahn &
	isotherm	disperse	approximation		in concentration.	dynamics of CO ₂	differential algebraic	structure of the monolith provided	Brandan
		d plug	model			breakthrough in	equations (PDAEs) was	an excellent match to experimental	
		flow				a carbon	coded in gPROMS	results whereas the model based on	i, 2005)
						monolith column	software to obtain a	the equivalent single channel	
							numerical model.	approach incorrectly predicted	
								higher separation efficiencies at	
								different concentrations.	
24	Multisite	Axial	A double LDF	Non-	Negligible heat,	Separation of a	The fixed-bed model was	The proposed model was able to	(Cavena
	Langmuir	disperse	approximation	isothermal	mass, and	mixture of CH ₄	solved in gPROMS (PSE	predict well the behavior of the	ti,
	isotherm	d plug	model to	(adiabatic)	momentum transport	(55%)-CO ₂	Enterprise, London, U.K.)	binary mixture in a fixed bed.	Grande,
		flow	express		in the radial	(45%) by VSA-	using orthogonal	Darken's law provided a	&
			macropore and		direction of the	PSA technology	collocation method on	successful correction of the	Rodrigu
			micropore		column.	using a Takeda	finite elements with 25	micropore diffusion coefficients	es,
			diffusion		Pressure drop was	carbon molecular	finite elements and 2	in the nonlinear regions of the	2005)
			equations		described using	sieve 3K	interior collocation points	isotherms.	
					Ergun equation.		per element.		

					Table 2.1, contin	uea			
			Model assumptions						
No.	Equilibrium	Flow	Mass transfer rate	Heat	Others	Application	Solution method	Results and comments	Ref.
	relationship	pattern	model	effects					
25	Langmuir	Axial	LDF approximation	Isother	Negligible radial	Adsorption of	A set of PDAEs was	The dispersion and mass transfer	(Mor
	isotherm	disperse	model	mal	concentration gradient.	carbon dioxide	solved using the	coefficients were calculated by	eira
		d plug			Negligible pressure drop.	from mixtures	PDECOL package in the	theoretical correlations and the	et
		flow				of CO ₂ diluted	FORTRAN language,	model described quite very well	al.,
						in helium onto	which is based on the	the dynamics of CO2 adsorption	2006
						a hydrotalcite-	method of orthogonal	in a fixed bed.)
						like Al-Mg	collocation of finite		
						compounds in	elements for partial		
						a fixed bed	differential equations in		
							double precision.		
26	Extended	Axial	General LDF model	Non-	Negligible radial	PSA	The gPROMS modeling	Compared with the LDF model	(Ki
	Langmuir-	disperse	with constant	isother	concentration and	separation of a	tool developed by	with constant diffusivity the	m et
	Freundlich	d plug	diffusivity/	mal	temperature gradients.	mixture of	Process Systems	experimental breakthrough	al.,
	isotherm	flow	Modified LDF model		Thermal equilibrium	CH_4/CO_2	Enterprise Ltd. was used	curves and adsorption dynamics,	2006
			with concentration-		between fluid and particles.	(50/50 vol %)	to obtain the solution of	was well-predicted by using the)
			dependent diffusivity		The pressure drop along	using Takeda	the dynamic simulation	proposed non-isothermal and	
					the bed was calculated by	3A carbon	of the model.	non-adiabatic modified LDF	
					the Ergun equation.	molecular		model with concentration-	
						sieve		dependent diffusivity.	

					1 4010 2:1	, commaca			
			Model						
			assumptions						
No.	Equilibrium	Flow	Mass transfer	Heat	Others	Application	Solution method	Results and comments	Ref.
	relationship	pattern	rate model	effects					
27	Extended	Axial	LDF	Non-	Negligible gradients	Fixed-bed	The complete	The model described	(Delgado et al., 2006;
	Langmuir	dispersed	approximation	isothermal	in radial	adsorption of	model was solved	adequately the breakthrough	Delgado, Maria A.
	isotherm	plug flow	model with a		concentration and	binary gas	numerically using	curves for the experiments	Uguina, Sotelo,
			single lumped		temperature.	mixtures	the PDECOL	with low CO ₂ concentration,	Beatriz Ruiz, &
			mass transfer		Thermal equilibrium	(CO ₂ /He,	program that uses	whereas the error was	Marcio Rosario,
			coefficient		between the gas and	$CO_2/N_{2,}$ and	orthogonal	higher for the runs with	2007)
					particles.	CO ₂ /CH ₄) onto	collocation on	higher CO ₂ concentration.	
					The momentum	silicalite pellets,	finite elements	The quality of the prediction	
					balance represented	sepiolite, and a	technique.	was improved introducing	
					by Ergun's equation.	basic resin		interaction factors in this	
								model, because of the strong	
								interactions between the	
								adsorbed CO ₂ molecules.	

			AC 1.1		1 0010 2.	ii, continued			
			Model assumptions			-			
No.	Equilibrium	Flow	Mass transfer rate	Heat	Others	Application	Solution method	Results and comments	Ref.
	relationship	pattern	model	effects					
28	Langmuir	Axial	A double LDF	Non-	Negligible radial	Adsorption of	Orthogonal	The presented model fitted well with	(Lein
	Isotherm	disperse	approximation	isother	variations in	a mixture of CH ₄	collocations were	experimental data, for both outlet	ekuge
		d plug	model	mal	concentration and	(70%)-CO ₂ $(30%)$	used as a spatial	composition and bed temperature. This	l-le-
		flow			temperature.	in a column	discretization	indicated that the approximations	Cocq,
					Thermal equilibrium	packed with	method. The resulting	proposed in this study give a good	Tayak
					between the gas and	bidisperse	ordinary differential-	representation of the intra-particle mass	out-
					solid phases.	adsorbent (5A	algebraic system of	transfer.	Fayol
					Negligible pressure	zeolite)	equations was solved		le, Le
					drop.		by the DDASPG		Gorre
							integration subroutine		c, &
							(IMSL library), based		Jallut,
							on the Petzold–Gear		2007)
							BDF method.		
29	Multi-site	Axial	A rigorous	Non-	No radial variations	Low-	The mathematical	The results showed that both adsorption	(Gran
	Langmuir	disperse	description for	isother	in temperature,	concentration CO ₂	model was solved	and desorption curves were not	de &
	isotherm	d plug	macropore	mal	pressure, and	removal from flue	using gPROMS 3.01	symmetrical, which cannot be accurately	Rodri
		flow	diffusion model as		concentration.	gas streams by	(PSE Enterprise,	described by the employed mathematical	gues,
			well as an LDF		The pressure drop	electric swing	United Kingdom).	model. This was attributed to the	2008)
			approximation for		was described by	adsorption using		channels with different sizes in the	
			micropre diffusions		Darcy's law.	monolith		boundaries of the honeycomb.	

'Table 2.1, continued'

			Model						
			assumptions			_			
No.	Equilibrium	Flow	Mass transfer	Heat	Others	Application	Solution method	Results and comments	Ref.
	relationship	pattern	rate model	effects					
30	Virial	Axial	A double LDF	Non-	No mass, heat or	Separation of	The numerical solutions were	A good agreement was observed	(Grande
	isotherm	disperse	approximation	isothermal	velocity gradients in	gas mixtures	performed with gPROMS (PSE	between the experimental and the	et al.,
	model	d plug	model to		the radial direction.	containing	Enterprise, UK) using the	predicted concentration history at	2008)
		flow	express		Thermal equilibrium	CO ₂ ,CH ₄ ,	orthogonal collocation on finite	the end of the column and also the	
			macropore and		between the gas and	CO, N_2 and	elements as the numerical	temperature evolution within the	
			micropore		solid phases.	H ₂ by PSA	method.	column. A deviation between the	
			diffusion		The momentum	using		adsorbed amount obtained	
			equations		balance represented	activated		experimentally and the predicted	
					by Ergun's equation.	carbon		by the model was lower than 10%.	
31	Toth isotherm	Axial	LDF	Non-	Negligible radial	Fixed-bed	The mathematical model was	The model acceptably reproduced	(Dantas et
		disperse	approximation	isothermal	concentration and	adsorption of	solved using the commercial	the experimental data for the	al., 2009;
		d plug	model with a	(adiabatic/	temperature	binary gas	software gPROMS (Process	different feed concentrations and	Dantas et
		flow	single lumped	non-	gradients.	mixtures	System Enterprise Limited,	temperatures. By using the Toth	al., 2011;
			mass transfer	adiabatic)	The momentum	(CO ₂ /He and	UK). The orthogonal	equation for pure components, the	Dantas et
			coefficient		balance represented	CO ₂ /N ₂)	collocation method on finite	simulated curve fitted well the	al., 2011)
					by Ergun's equation.	using	elements was used with six	experimental data, whereas a	
						activated	finite elements and three	deviation was observed for	
						carbon and	collocation points in each	multicomponent.	
						zeolite 13X	element of the adsorption bed.		

Table 2.1. continued	'Table	2.1.	continued'
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					Table 2.1, continued	d		
_		Model assumptions						
Equilibriu	Flow	Mass	Heat	Others	Application	Solution method	Results and comments	Ref.
m	patter	transfer rate	effects					
relationship	n	model						
Virial	Axial	A rigorous	Isother	Negligible radial	Adsorption of	Simulations of the	An exponential dependence of the	(Shen, Grande,
isotherm	disper	description of	mal	concentration	CO ₂ on pitch-	presented mathematical	micropore diffusivity with	Li, Yu, &
model /	sed	macropore		gradient.	based activated	model were performed in	temperature was shown to	Rodrigues,
Multisite	plug	and		Constant velocity	carbon	gPROMS (PSE	correctly describe the experimental	2010)
Langmuir	flow	micropore		within the column.		Enterprise, UK) using the	data within the temperature range	
isotherm		diffusion				orthogonal collocation on	studied.	
		models				finite elements.		
	Equilibriu m relationship Virial isotherm model / Multisite Langmuir isotherm	Equilibriu mFlow patterrelationshipnVirialAxialisothermdispermodel /sedMultisiteplugLangmuirflowisotherm	Equilibriu relationshipFlow patterModel assumptionsEquilibriu m relationshipFlow patterMass transfer rate modelVirialAxialA rigorousisothermdisper seddescription of macroporeMultisiteplug flowandLangmuirflowmicropore diffusion models	Model assumptionsEquilibriu mFlow patterMass transfer rate modelHeat effectsrelationship nnmodelHeatVirialAxialA rigorousIsotherisothermdisper seddescription of macroporemalMultisiteplug flowandisothermflowmicropore	Model assumptionsEquilibriu m meterFlow patterMass transfer rate effectsHeat effectsOthersrelationship relationshipnmodelVirialAxial disperA rigorousIsotherNegligible radialisothermdisper description of madelmalconcentrationmodel /sedmacroporegradient.Multisiteplug andandConstant velocityLangmuirflowmicroporewithin the column.isothermdiffusion modelsmodels-	Table 2.1, continuedModel assumptionsApplicationEquilibriu m patterFlow transfer rate modelHeat effectsOthers effectsApplicationVirial isothermAxial disperA rigorous description of malIsotherNegligible radial gradient.Adsorption of CO2 on pitch-Multisite isothermplug andandConstant velocity within the column.carbonIsothermdiffusion modelsmicroporewithin the column.	Table 2.1, continuedModel assumptionsEquilibriu m m patterFlow transfer rate modelHeat effectsOthers effectsApplicationSolution methodVirial isothermAxial disperA rigorousIsotherNegligible radial ocncentrationAdsorption of CO2 on pitch-Simulations of the presented mathematical model were performed in gradient.Multisite isothermplug andandConstant velocity within the column.carbongPROMS (PSE Enterprise, UK) using the orthogonal collocation on finite elements.	Table 2.1, continued Model assumptions Equilibriu Flow Mass Heat Others Application Solution method Results and comments relationship n model Model Adsorption of Simulations of the An exponential dependence of the Virial Axial A rigorous Isother Negligible radial Adsorption of Simulations of the An exponential dependence of the isotherm disper description of mal concentration CO2 on pitch- presented mathematical micropore diffusivity with model / sed macropore gradient. based activated model were performed in temperature was shown to Multisite plug and Constant velocity carbon gPROMS (PSE correctly describe the experimental Langmuir flow micropore within the column. enterprise, UK) using the data within the temperature range isotherm diffusion models finite elements. finite elements.

(Table 0.1

			Model			,			
			assumptions						
No.	Equilibrium	Flow	Mass transfer	Heat	Others	Application	Solution method	Results and comments	Ref.
	relationship	pattern	rate model	effects					
33	Dual-site	Plug	LDF	Non-	No radial variations	CO ₂ capture	A complete	The results indicated the	(Agarwal,
	Langmuir	flow	approximation	isothermal	in temperature,	from an 85%-	discretization	potential of the superstructure	Biegler, &
	isotherm		model		pressure, and	15% N ₂ -CO ₂	approach that uses	approach to predict PSA cycles	Zitney,
			with lumped		concentration.	feed mixture	the finite volume	with up to 98% purity and	2010a;
			mass transfer		The gas and the	using PSA	method was applied	recovery of CO ₂ . Verifications	Agarwal et
			coefficient		solid phases are in	cycles/	in both spatial and	of the accuracy of the	al., 2010b)
					thermal equilibrium.	CO ₂ capture	time domains, and	discretization scheme showed	
					The pressure drop	from a synthesis	the resulting large-	this approach is reasonably	
					along the bed was	gas feed mixture	scale nonlinear	accurate in capturing the	
					calculated by the	(55% H_2 and	programming	dynamics of PSA systems	
					Ergun equation.	45% CO ₂) using	problem (NLP) was	governed by hyperbolic PDAEs	
						PSA cycles	solved using an	and steep adsorption fronts, and	
							interior point NLP	can be used for PSA systems	
							solver.	with efficient NLP solvers.	

					Table 2.1	l, continued'			
			Model						
			assumptions						
No.	Equilibrium	Flow	Mass transfer	Heat	Others	Application	Solution method	Results and comments	Ref.
	relationship	pattern	rate model	effects					
34	Langmuir	Axial	LDF	Non-	Negligible radial	Fixed bed	-	The curvature of the concentration	(Mulgund
	isotherm	dispersed	approximation	isothermal	temperature and	adsorption of		breakthrough curve including the	math et al.,
		plug flow	model for		concentration	carbon dioxide		noted tailing was predicted with good	2012)
			external fluid		gradients.	from a CO ₂ -N ₂		accuracy. Energy profile was	
			film mass		Negligible pressure	gas mixture		predicted with lower accuracy.	
			transfer /		drop.	(10% CO ₂ in		However, the point at which the	
			A rigorous			90% N ₂) using		temperature breakthrough occurs was	
			description of			zeolite 13X		estimated with good accuracy which is	
			pore diffusion					the most important factor for	
			model					industrial applications.	
			for intra-						
			particle						
			mass transfer						

* The adsorbable component is present at only low concentration in an inert carrier.

2.3 Development and analysis of a mathematical model

The fixed-bed column mathematical models are used to predict the transient behavior of the concentration and temperature profiles for any defined changes in the initial parameters such as feed concentration, temperature, and flow rate. A complete mathematical model capable of describing the dynamics behavior of a fixed-bed adsorption system is established based on a set of fairly complex partial differential and algebraic equations (PDAEs) constructed from conservation of mass, energy, and momentum and augmented by appropriate transport rate equations and equilibrium isotherms (Hwang et al., 1995). The models used to represent a PSA process differ mainly in the form of the mass transfer rate, the form of the equilibrium isotherm, thermal effects, and the pressure drop along the bed. General descriptions of the above-mentioned items are presented in the following subsections. Many mathematical models for gas-solid adsorption in an adsorption column have been published over the past few decades, and there is still interest in developing a description of the dynamic evolution of such systems (Afzal, Rahimi, Ehsani, & Tavakoli, 2010; Leinekugel-le-Cocq et al., 2007).

2.3.1 Fluid phase material balance

The transient gas-phase component mass balance, which includes the axial dispersion term, convection flow term, accumulation in the fluid phase, and source term caused by the adsorption process on the adsorbent particles, can be represented by the following equation for a differential control volume of the adsorption column (Ruthven, 1984; Yang, 1987):

$$-D_{zi}\frac{\partial^2 c_i}{\partial z^2} + \frac{\partial}{\partial z}(uc_i) + \frac{\partial c_i}{\partial t} + (\frac{1-\varepsilon_b}{\varepsilon_b})\rho_p \frac{\partial \overline{\overline{q_i}}}{\partial t} = 0$$
(2-1)

where c_i represents the adsorbate concentration in the fluid phase; z is the distance along the bed length; u is the fluid velocity; t denotes time; ε_b is the bed void fraction; ρ_p is the particle density; $\overline{q_i}$ denotes average concentration of component *i* in adsorbent particle, which forms a link between the fluid and solid phase mass balance equations; and the effects of all mechanisms that contribute to axial mixing are lumped into a single effective axial dispersion coefficient, D_{zi} , which can be estimated using the following correlation (Da Silva, Silva, & Rodrigues, 1999; Ruthven, 1984; Wakao & Funazkri, 1978; Welty, Wicks, Wilson, & Rorrer, 2000; Yang, 1987):

$$\frac{\varepsilon_b D_{zi}}{D_{mi}} = 20 + 0.5ScRe \tag{2-2}$$

where D_{mi} is the molecular diffusivity of component *i* and *Sc* and *Re* are the Schmidt and Reynolds numbers, respectively.

The above equation, Eq. (2-1), is used to find the distribution of gas composition along the bed. Assuming no radial dependence of concentration and solid loading, c_i and $\overline{q_i}$, represent cross-sectional average values (these variables are functions of *t* and *z*). The well-known Danckwert's boundary conditions for a dispersed plug flow system can

be assumed as follows (Khalighi, Farooq, & Karimi, 2012; Wehner & Wilhelm, 1956):

$$D_{zi} \frac{\partial c_i}{\partial z}\Big|_{z=0} = -u\Big|_{z=0} (c_i\Big|_{z=0^-} - c_i\Big|_{z=0})$$
(2-3)

$$\left. \frac{\partial c_i}{\partial z} \right|_{z=L} = 0 \tag{2-4}$$

where $c_i|_{z=0^-}$ represents the feed composition for component *i* and *L* is the bed length. In the above model, Eq. (2-1), if the flow pattern is represented as plug flow, axial dispersion can be neglected, and therefore the term $-D_{zi}(\frac{\partial^2 c_i}{\partial z^2})$ can be dropped, reducing Eq. (2-1) to a first-order hyperbolic equation. This is a reasonable approximation, particularly for large industrial units, for which the term representing the axial dispersion is very small compared to the convection term (Simo, Brown, & Hlavacek, 2008). As a result of pressure and temperature variations, gas density and hence gas velocity vary along the bed. The following equation expressing the overall material balance for the bulk phase in the adsorption column is used to compute the velocity distribution through the bed (Ko et al., 2005):

$$-D_{zi}\frac{\partial^2 C}{\partial z^2} + \frac{\partial (uC)}{\partial z} + \frac{\partial C}{\partial t} + (\frac{1 - \varepsilon_b}{\varepsilon_b})\rho_p \sum_{i=1}^n \frac{\partial \overline{\overline{q_i}}}{\partial t} = 0$$
(2-5)

where C is the total concentration in the bulk phase and n is the number of components.

Applying the ideal gas law $(c_i = \frac{y_i P}{RT_g})$, the overall mass balance equation can be expressed

as follows (Ahn, Yang, & Lee, 2001; Lee et al., 1999):

$$-D_{zi}\frac{\partial^{2}P}{\partial z^{2}} + \frac{\partial P}{\partial t} + P\frac{\partial u}{\partial z} + u\frac{\partial P}{\partial z} + PT_{g}\left[-D_{zi}\frac{\partial^{2}(1/T_{g})}{\partial z^{2}} + \frac{\partial(1/T_{g})}{\partial t} + u\frac{\partial(1/T_{g})}{\partial z}\right]$$
$$-2D_{zi}T_{g}\frac{\partial(1/T_{g})}{\partial z}\frac{\partial P}{\partial z} + (\frac{1-\varepsilon_{b}}{\varepsilon_{b}})\rho_{p}RT_{g}\sum_{i=1}^{n}\frac{\partial q_{i}}{\partial t} = 0$$
(2-6)

where y_i is the mole fraction of component *i* in the gas phase, *P* is the total pressure, T_g is the gas temperature and *R* is the universal gas constant.

2.3.2 Complexity of kinetic models

The term $\frac{\partial \overline{q_i}}{\partial t}$ in Eq. (2-1) represents the overall rate of mass transfer for component *i* (at time *t* and distance *z*) averaged over a particle. The mass balance for an adsorbent particle yields the adsorption rate expression, which may be written as

$$\frac{\partial \overline{\overline{q_i}}}{\partial t} = f(q_i, c_i) \tag{2-7}$$

For an isothermal system, the expressions for the concentration profiles in both phases, $[c_i(z,t), \overline{q_i}(z,t)]$, is given by the simultaneous solution of Eqs. (2-1) and (2-7), subject to the initial and boundary conditions imposed on the column. For non- isothermal systems, an energy balance must also be taken into account. In this case, all equations are coupled because, in general, both the equilibrium concentration and the rate coefficients are temperature dependent.

Although the mass transfer rate expression, Eq. (2-7), was written here as a single equation, it commonly consists of a set of equations comprising one or more diffusion equations with their associated boundary conditions. It is worth noting that a kinetic model is basically a mass balance that involves different variables describing mass transfer mechanisms within the adsorbent particle (Chahbani & Tondeur, 2000). A variety of mass transfer kinetic models with different degrees of complexity can be found in the literature. Mass transfer kinetic models can be classified into two main categories based on the assumption of local equilibrium or the existence of mass transfer resistance between the adsorbent particle and the fluid phase. They are introduced in the following subsections.

2.3.2.1 Local equilibrium model

This model is expressed by the existence of an instantaneous local equilibrium (ILE) between the solid and fluid concentrations. If the mass transfer rate is relatively rapid, one may assume that local equilibrium is always maintained between the gas phase and the adsorbed phase within the particle at all points in the column. In other words, the local equilibrium assumption relates to the negligible effect of mass transfer resistance through the particles. As a result, it is assumed that, in this model, the adsorptive quantity is equal to the equilibrium adsorptive quantity:

$$\frac{\partial \overline{\overline{q_i}}}{\partial t} = \frac{\partial q_i^*}{\partial t}$$
(2-8)

In the above equation q_i^* is the adsorbed phase concentration of species *i* in equilibrium with the fluid phase concentration.

Equilibrium theory is aimed at identifying the general features of the dynamic response of the column without detailed calculations, as the overall pattern of the response is governed by the form of the equilibrium relationship rather than by kinetics. However, in practice, because axial mixing and mass transport resistances are neglected, breakthrough curves predicted by equilibrium models fail to give quantitatively satisfactory results and give only approximate representations of the behavior observed (Hwang et al., 1995). Although such systems are not common in practice, their analytical solution can provide useful information about the process dynamics and system behavior, which is quite valuable for preliminary design and analysis, leading to a greater understanding of the behavior of more complex systems.

Based on the classification presented in Table 2-1, the simplest case to consider is an isothermal system with no axial dispersion in which a trace level component is adsorbed from a non-adsorbing carrier gas with the assumption of negligible mass transfer resistance (model 1). Because the adsorbable component is present at a low concentration (trace level assumption), variation in the fluid velocity across the mass transfer zone is considered to be negligible, and the superficial velocity calculated based on the flow at the inlet can be treated as a constant. For these systems, the differential gas phase mass balance, Eq. (2-1), reduces to

$$u\frac{\partial c_i}{\partial z} + \frac{\partial c_i}{\partial t} + (\frac{1-\varepsilon_b}{\varepsilon_b})\rho_p \frac{\partial q_i^*}{\partial t} = 0$$
(2-9)

Analytical determination of the concentration front in a packed adsorption column is limited to a few simple cases. Using a linear equilibrium isotherm, it is possible to obtain an analytical solution for isothermal or adiabatic systems with non-disperse behavior. Assuming a constant pattern profile, an improved analytical solution was obtained for the system that used a nonlinear equilibrium isotherm such as the Langmuir isotherm, but there are some assumptions that restrict its application (Yang, 1987). In the case of the PSA processes, the first analytical solution of the equilibrium model was obtained by (Shendalman & Mitchell, 1972) for the separation of CO₂ from a He-CO₂ mixture using silica gel as an adsorbent (model 1). They implemented the linear equilibrium relation for isothermal adsorption of a trace-level component in a one-dimensional system with no axial dispersion. The assumption of linear equilibrium for one adsorbable component permitted them to obtain a solution to the equations by the characteristic method. However, the experimental data revealed rather large deviations from the equilibrium theory predictions, suggesting that effects of mass transfer resistance are likely important. (Chan, Hill, & Wong, 1981) extended the local equilibrium theory to the separation of two-component gaseous mixtures via the PSA process in an isothermal system in which both the carrier and the impurity are adsorbed. They analytically studied the assumption of instant equilibrium between the adsorbate and the adsorbent when the more strongly adsorbed component is present at a trace level and the linear isotherms for both components were considered. Later, Fernandez & Kenney (1983) and Knaebel & Hill (1985) solved the model for binary mixtures with linear adsorption isotherms, and Serbezov and Sotirchos (1997) extended the solution to multicomponent mixtures with linear adsorption isotherms.

2.3.2.2 Mass transfer resistance models

The modeling of transport equations in a packed column strongly depends on the mechanism by which the mass transfer from the fluid to the solid phase occurs. In fact, equilibrium theory is confined to systems in which the adsorptive selectivity depends on a difference in equilibrium and is not useful for systems in which separation is based on kinetic selectivity (Hassan, Ruthven, & Raghavan, 1986). An example of kinetic adsorptive separation is the separation of a CO₂/CH₄ mixture using a carbon molecular sieve, in which the separation is achieved by the large difference in diffusion rates between the two components (Diagne et al., 1996). Therefore, in modeling a real practical non-equilibrium packed column, the effects of mass transfer resistance between the fluid and the particle and within the particle must be considered (Hwang et al., 1995). In an attempt to construct a theoretical representation that more closely approximates a real process, researchers have developed dynamic models that consider effects due to dispersion and a finite mass transfer rate.

The mass transfer of solute from bulk gas into the solid phase is driven by equilibrium isotherms, whereas the mass balance equation inside the adsorbent particle depends on the adsorbent structure. At the microscopic level, the diffusion of the adsorbate into the adsorbent particles before adsorption onto the micropore surface (or adsorption onto the macropore surface, if no micropores exist) involves different transfer mechanisms. The adsorbate molecules initially must cross the external film surrounding each adsorbent particle and then diffuse through and along the porous structure of the adsorbent, as illustrated in Fig. 2.1. Depending on the specific system and the conditions, any one of the three different types of potential resistance to mass transfer may be dominant, and more than one resistance may be significant. These three potential resistances are the external fluid film resistance and the intraparticle diffusional resistances, the macropore diffusional resistance (LeVan, Carta, & Yon,

1999). In general, the mass transfer process through such a heterogeneous system can be expressed by detailed models identifying the film resistance around the solid particles and macropore/micropore resistances inside the particles. The most general case in adsorption process modeling is the case of macropore/micropore diffusion with external film resistance. Consequently, the discussion in the following subsections will focus on the case of the bidisperse pore-diffusion model with clearly distinct macropore/micropore diffusion.

2.3.2.2.1 External fluid film resistance

External fluid film mass transfer is defined based on the concentration difference across the boundary layer surrounding each adsorbent particle and is strongly affected by the hydrodynamic conditions outside the particles (as characterized by the system's Sherwood, Reynolds, and Schmidt numbers) (LeVan et al., 1999). Indeed, it is supposed that the mass transfer resistance between the bulk phase and the macro-porous gas phase is localized to an external film around the adsorbent particles. By assuming steady-state conditions at the fluid-solid interface, the mass transfer rate across the external film is supposed to be equal to the diffusive flux at the particle surface (Farooq, Qinglin, & Karimi, 2001). In fact, because no accumulation of adsorbates is allowed, the film transfer and macropore diffusion can be treated as sequential steps, and mass conservation assumption is applicable. It can be expressed as the following equation (Jin, Malek, & Farooq, 2006; LeVan et al., 1999):

$$\frac{\partial \overline{q_i}}{\partial t} = \frac{3k_{fi}}{R_p} (c_i - c_{pi} \Big|_{(t,R_p)}) = \frac{3}{R_p} \varepsilon_p D_{pi} \frac{\partial c_{pi}}{\partial R} \Big|_{(t,R_p)}$$
(2-10)

where k_{fi} is the external film mass transfer coefficient, R_p is the macroparticle radius, c_{pi} is the adsorbate concentration in the macropore, which is a function of radial position in

the particle, ε_p is the adsorbent porosity, D_{pi} is the macropore diffusivity, and R is the distance along the macroparticle radius.

The external film mass transfer coefficient, k_{fi} , around the particles can be estimated from the following correlation, which is applicable over a wide range of conditions (Wakao & Funazkri, 1978):

$$Sh = \frac{2k_{fi}R_p}{D_{mi}} = 2 + 1.1 Sc^{1/3} Re^{0.6}$$
(2-11)

In most gas adsorption studies, the intraparticle diffusional resistance is normally much greater than the external fluid film resistance (intraparticle transport of the adsorbate is the slower step). Therefore, it is reasonable to assume negligible gas-side resistance and simulate adsorption systems based on a diffusion model (Carta & Cincotti, 1998; S. Farooq et al., 2001; Raghavan et al., 1985). An accurate kinetic model that accounts for the intraparticle diffusional resistances can provide reliable simulations of kinetically controlled PSA processes. Indeed, neglecting intraparticle mass transfer kinetics leads to significant deviations from the exact solution (Chahbani & Tondeur, 2000).

2.3.2.2.2 *Macropore diffusional resistance*

Diffusion in sufficiently large pores (macro- and mesopores) such that the diffusing molecules escape from the force field of the adsorbent surface is often referred to as macropore diffusion (or pore diffusion). Depending on the relative magnitude of the pore diameter and the mean free path of the adsorbate molecules, transport in a macropore can occur by different mechanisms (Karger & Ruthven, 1992). For gas phase diffusion in small pores at low pressure, when the molecular mean free path is much greater than the pore diameter, Knudsen diffusion dominates the transport mechanism. In this case, the resistance to mass transfer mainly arises from collisions between the diffusing molecules and the pore wall. The Knudsen diffusivity (D_{ki}) is independent of pressure and varies

only weakly with temperature as follows (Karger & Ruthven, 1992; Ruthven, 1984; Suzuki, 1990; Yang, 1987):

$$D_{ki} = 9700 r_p \sqrt{\frac{T}{M}}$$
(2-12)

where r_p is the mean macropore radius in cm, T is the temperature, and M is the molecular weight of the adsorbate.

By contrast, when the molecular mean free path is small relative to the pore diameter, the bulk molecular diffusion will be the dominant transport mechanism and can be estimated from the Chapman-Enskog equation (Bird, Stewart, & Lightfoot, 2002; Ribeiro et al., 2008; Ruthven, 1984; Sherwood, Pigford, & Wilke, 1975) for binary systems or the Stefan–Maxwell equation for multi-component systems (Suzuki, 1990). In the case of molecular diffusion, the collisions between diffusing molecules are the main diffusional resistance. For the intermediate case, both mechanisms are of comparable significance, and thus the combined effects of the Knudsen and the molecular diffusion constitute the rate-controlling mechanism. The effective macropore diffusivity (D_p) is obtained from the Bosanquet equation (Grande et al., 2008; Yang, 1987):

$$\frac{1}{D_{pi}} = \tau (\frac{1}{D_{ki}} + \frac{1}{D_{mi}})$$
(2-13)

where τ is the pore tortuosity factor.

As discussed above, in macropore diffusion, transport occurs within the fluid-filled pores inside the particle (LeVan et al., 1999; Ruthven, 1984). In this situation, a differential mass balance equation for species *i* over a spherical adsorbent particle may be written as follows (Do, 1998b; Gholami & Talaie, 2009; Jin et al., 2006; LeVan et al., 1999; Qinglin, Farooq, & Karimi, 2003):

$$\frac{\partial c_{pi}}{\partial t} + \left(\frac{1 - \varepsilon_p}{\varepsilon_p}\right) \frac{\partial \overline{q_i}}{\partial t} = \frac{1}{R^2} \frac{\partial}{\partial R} \left(R^2 D_{pi} \frac{\partial c_{pi}}{\partial R}\right)$$
(2-14)

This equation is used to determine the composition of the gas penetrating macropore volume at each radial position. In the above equation, $\overline{q_i}$ is the average adsorbed phase concentration of component *i* in the micropore, which is related to the adsorbate flux at the micropore mouth by either equation 2-18 or 2-19, depending on the expression of the dominant transport mechanism in the micropore. The corresponding boundary conditions for macropore balance are as follows (Do, 1998b; Gholami & Talaie, 2009; Jin et al., 2006; LeVan et al., 1999; Qinglin, Farooq, et al., 2003; Qinglin, Sundaram, & Farooq, 2003):

$$\left. \frac{\partial c_{pi}}{\partial R} \right|_{(t,0)} = 0 \tag{2-15}$$

The external fluid – film resistance can be reflected in the boundary condition as follows:

$$\varepsilon_p \left. D_{pi} \left. \frac{\partial c_{pi}}{\partial R} \right|_{(t,R_p)} = k_{fi} (c_i - c_{pi} \Big|_{(t,R_p)})$$

$$(2-16)$$

or $c_{pi}(t, R_p) = c_i$ for no external film resistace (when pure adsorbate is fed to the column) (2-17)

Micropore diffusional resistance

In very small pores in which the pore diameter is not much greater than the molecular diameter, the adsorbing molecules can never escape from the force field of the pore wall, even at the center of the pore. Such a mechanism, in which transport may occur by an activated process involving jumps between adsorption sites, is often called micropore diffusion (also known as solid diffusion) (LeVan et al., 1999; Ruthven, 1984). In this situation, the intraparticle gas phase is neglected, and diffusion through it is supposed to be null (Chahbani & Tondeur, 2000). Consequently, the material balance equation in the

micropores does not contain any gas phase accumulation term. As illustrated in Fig. (2.1), transport in the micropores may occur by three different mechanisms: barrier resistance (confined at the micropore mouth), distributed micropore interior resistance, and the combined effects of both resistances (Cavenati et al., 2005; S. Farooq et al., 2001; Srinivasan, Auvil, & Schork, 1995).

The mass transfer rate across the micropore mouth can be expressed by the following equations (Buzanowski & Yang, 1989; Jin et al., 2006; LeVan et al., 1999; Qinglin, Sundaram, et al., 2003):

$$\frac{\partial q_i}{\partial t} = k_{bi}(q_i^* - \overline{q_i}) \quad \text{when the gas diffusion is controlled by the barrier}$$
resistance
$$(2-18)$$
or
$$= \frac{3}{R_c} D_{\mu i} \frac{\partial q_i}{\partial r}\Big|_{(t,R_c)} \quad \text{when the distributed micropore interior resistance is}$$

dominant

where k_{bi} is the barrier transport coefficient, R_c is the microparticle radius, Q_i is the distributed adsorbate concentration in the micropore, $D_{\mu i}$ is the micropore diffusivity of component *i*, and *r* is the distance along the microparticle radius.

The strong dependence of the micropore diffusivity on concentration can be expressed using Darken's equation (Cavenati et al., 2005; Chihara, Suzuki, & Kawazoe, 1978; Do, 1998a; Kawazoe, Suzuki, & Chihara, 1974; Khalighi et al., 2012; Ruthven, Farooq, & Knaebel, 1994):

$$D_{\mu i} = D_{\mu i}^{\infty} \left. \frac{d \ln(p_i)}{d \ln(q_i)} \right|_T$$
(2-20)

where $D_{\mu i}^{\infty}$ is the micropore diffusivity of component *i* at infinite dilution and p_i is the partial pressure of component *i*, which is in equilibrium with the adsorbed concentration in the micropore.

(2 - 19)

The temperature dependence of the corrected diffusivity and the surface barrier mass transfer coefficients follows an Arrhenius-type form, as described by the following (Cavenati et al., 2005; Gholami & Talaie, 2009; Grande & Rodrigues, 2004; Grande & Rodrigues, 2005; Khalighi et al., 2012; Qinglin, Sundaram, et al., 2003):

$$D_{\mu i}^{\ \infty} = D_{\mu i}^{\ 0} \exp(-\frac{E_{ai}}{R_g T_s}) \tag{2-21}$$

$$k_{bi} = k_{bi}^{0} \exp(-\frac{E_{bi}}{R_g T_s})$$
(2-22)

where $D_{\mu i}{}^{0}$ and $k_{bi}{}^{0}$ are the temperature-independent pre-exponential constants, R_{g} is the universal gas constant, T_{s} is the solid temperature, and E_{ai} and E_{bi} are the activation energy of micropore diffusion and the activation energy of surface barrier resistance for component *i*, respectively.

When the resistance distributed in the micropore interior dominates the transport of species *i*, the mass balance equation for micropore diffusion is the following (Jin et al., 2006; LeVan et al., 1999; Qinglin, Sundaram, et al., 2003):

$$\frac{\partial q_i}{\partial t} = \frac{1}{r^2} \frac{\partial}{\partial r} \left(r^2 D_{\mu i} \frac{\partial q_i}{\partial r} \right)$$
(2-23)

The corresponding boundary conditions for the microparticle balance are as follows (Jin et al., 2006; LeVan et al., 1999; Qinglin, Sundaram, et al., 2003):

$$\left. \frac{\partial q_i}{\partial r} \right|_{(t,0)} = 0 \tag{2-24}$$

when a combination of barrier and distributed micropore interior resistaces is dominant,

the barrier resistance can be reflected in the boundary condition as follows:

$$\frac{3}{R_c} D_{\mu i} \frac{\partial q_i}{\partial r} \bigg|_{(t,R_c)} = k_{bi} (q_i^* - q_i \big|_{(t,R_c)})$$
(2-25)

or
$$q_i(t, R_c) = q_i^*$$
 for no barrier resistace (2-26)

The adsorbed amount at a certain time for component *i* based on particle volume can be calculated by volume integration of the concentration profiles in the macropores and micropores (Jin et al., 2006; Khalighi et al., 2012; Qinglin, Farooq, et al., 2003; Qinglin, Farooq, & Karimi, 2004; Qinglin, Sundaram, et al., 2003):

$$\overline{\overline{q}}_{i} = \varepsilon_{p} \frac{3}{R_{p}^{3}} \int_{0}^{R_{p}} c_{pi} R^{2} dR + (1 - \varepsilon_{p}) \frac{3}{R_{p}^{3}} \int_{0}^{R_{p}} \overline{q}_{i} R^{2} dR$$
(2-27)

where:

$$\bar{q}_{i} = \frac{3}{R_{c}^{3}} \int_{0}^{R_{c}} q_{i} r^{2} dr$$
(2-28)

In most kinetically selective processes, the controlling resistance for the uptake of sorbates is typically diffusion in the micropores (Cavenati et al., 2005; S. Farooq et al., 2001; Lamia et al., 2008). Micropore diffusion can contribute significantly to the overall intraparticle mass transport, primarily due to the higher concentration of the adsorbed phase, although the mobility of molecules in the adsorbed phase is generally much smaller than in the gas phase (Kapoor & Yang, 1990). Doong & Yang (1986) reported that micropore diffusion contributed as much as 50% to the total flux in the activated carbon pores during the PSA separation of CO₂, H₂, and CH₄ (model 5). Liu and Ruthven (1996) gravimetrically measured the diffusion of CO_2 in a carbon molecular sieve sample and concluded that the data were consistent with the barrier resistance model at lower temperatures, while the distributed micropore interior resistance model adequately fitted the data at higher temperatures. They found that the results suggested a dual resistance model with varying importance of the two components depending on pressure and temperature. In another study, Rutherford and Do (2000b) fitted the uptake of CO_2 in a sample of a carbon molecular sieve (Takeda 5A) using a model based on distributed diffusional resistance in the micropore interior. The model simulation results were in fair agreement with the experimental data. Qinglin et al. (2003a) and Qinglin et al. (2003b)

investigated the diffusion of carbon dioxide in three samples of carbon molecular sieve adsorbent. They indicated that transport of gases in the micropores of these samples is controlled by a combination of barrier resistance at the micropore mouth followed by a distributed pore interior resistance acting in series. The proposed dual resistance model was shown to be able to fit the experimental results over the entire range covered in that study. Cavenati et al. (2005) studied diffusion of CO₂ on the carbon molecular sieve 3K and reported that the initial difficulty associated with diffusion due to the surface barrier resistance was not observed in the uptake of CO₂. A successful description of diffusion in micropores was achieved using the distributed micropore interior resistance model without the need for the surface barrier resistance model at the mouth of the micropore (model 24). They attributed the absence of surface barrier resistance to performing the activation protocol at a higher temperature. Shen et al. (2010) studied diffusion of CO₂ on pitch-based activated carbon beads using diluted breakthrough experiments performed at different temperatures. To simulate the breakthrough curves, they developed a mathematical model based on a rigorous description of macropore and micropore diffusion with a nonlinear adsorption isotherm and assumed that the process was isothermal (model 32). The experimental results demonstrated that micropore resistances control the diffusion mechanism within the adsorbent. More recently, Mulgundmath et al. (2012) investigated concentration and temperature profiles of CO₂ adsorption from a CO_2 -N₂ gas mixture in a dynamic adsorption pilot plant unit to better understand the adsorbent behavior. A dynamic model based on an exact description of pore diffusion was developed for the simulation of non-isothermal adsorption in a fixed-bed (model 34). The proposed model was able to adequately predict the experimental data at all three ports for the duration of the experiment.



Figure 2.1: Schematic diagram showing various resistances to the transport of adsorbate as well as concentration profiles through an idealized bidisperse adsorbent particle demonstrating some of the possible regimes: (1)+(a) rapid mass transfer, equilibrium through particle; (1)+(b) micropore diffusion control with no significant macropore or external resistance; (1)+(c) transport controlled by the resistance at the micropore

interior; (1)+(d) controlling resistance at the surface of the microparticles; (2)+(a) macropore diffusion control with some external resistance and no resistance within the

microparticle; (2)+(b) all three resistances (micropore, macropore, and film) are significant; (2)+(c) diffusional resistance within the macroparticle with some external film resistance together with a restriction at the micropore interior (2)+(d) diffusional resistance within the macroparticle in addition to a restriction at the micropore mouth with some external film resistance.

2.3.2.2.3 Linear driving force model

Although the diffusional models are closer to reality, due to the mathematical complexities associated with such equations for the exact description of intraparticle diffusion in adsorbent particles, simpler rate expressions are often desirable (Carta & Cincotti, 1998; Zhang & Ritter, 1997). Simplified models are generally adopted by using an expression of the particle uptake rate, which does not involve the spatial coordinates. The approximations express the mass exchange rate between the adsorbent and its surroundings in terms of the mean concentration in the particle, regardless of the actual nature of the resistance to mass transfer (Lee & Kim, 1998). Simplifying assumptions should increase the practical applicability of the model without reduction of accuracy. The most frequently applied approximate rate law is the so-called linear driving force (LDF) approximation, which was first proposed by Glueckauf and Coates (1947). They originally suggested that the uptake rate of a species into adsorbent particles is proportional to the linear difference between the concentration of that species at the outer surface of the particle (equilibrium adsorption amount) and its average concentration within the particle (volume-averaged adsorption amount):

$$\frac{\partial \overline{q_i}}{\partial t} = k_i (q_i^* - \overline{q_i})$$
(2-29)

As can be seen, the overall resistance to mass transfer is lumped into a single effective linear driving force rate coefficient, k_i . Glueckauf demonstrated that the LDF overall mass transfer coefficient for spherical particles was equal to $15D_e/R_p^2$ (Glueckauf, 1955). The above equation has been shown to be valid for dimensionless times (D_et/R_p^2)>0.1, where D_e is the effective diffusivity (accounts for all mass transfer resistances) and *t* is the time of adsorption or desorption (Yang, 1987). Although the LDF model deals with the average concentrations of the adsorbate within the adsorbent particle, (Liaw, Wang, Greenkorn, & Chao, 1979) demonstrated that the same value for k_i could be simply obtained by assuming a parabolic concentration profile within the particle. This assumption was later shown to be acceptable, as the exact solution to the concentration profile has almost always been found to be a parabolic function (Do & Rice, 1986; Patton, Crittenden, & Perera, 2004; Tsai, Wang, & Yang, 1983; Tsai, Wang, Yang, & Desai, 1985; Yang & Doong, 1985). Sircar and Hufton (2000a) demonstrated that the LDF model approximation is in accordance with any continuous intraparticle concentration profile within a spherical particle when a numerical constant other than 15 is used in the expression of the LDF rate coefficient. The literature includes many attempts to develop new correlations for the accurate prediction of the overall LDF rate constant (Gholami & Talaie, 2009). When both the macropore and the micropore diffusions are dominant, the overall LDF mass transfer coefficient can be expressed by defining a single effective diffusivity related to both macropore and micropore diffusivities. The following correlation was proposed by Farooq and Ruthven (1990), in which more than one mass transfer resistance (i.e., film, macropore, and micropore resistances) is considered significant:

$$\frac{1}{k_i} = \frac{R_p q_0}{3k_{fi} c_0} + \frac{R_p^2 q_0}{15\varepsilon_p D_{pi} c_0} + \frac{R_c^2}{15D_{\mu i}}$$
(2-30)

where q_0 is the value of q at equilibrium with c_0 at feed temperature.

The above equation is actually an extension of the Glueckauf approximation, which, apart from validity for a linear isothermal system, is also known to work reasonably well for nonlinear systems.

Recently, the Stefan-Maxwell approach (Do & Do, 1998; Liow & Kenney, 1990) or the dusty gas model (Mendes, Costa, & Rodrigues, 1995; Atanas Serbezov & Sotirchos, 1998) has been proposed to describe adsorption kinetics. However, Sircar & Hufton (2000b) indicated that the LDF model is adequate to capture gas adsorption kinetics because in the estimation of the final process performance, the detailed characteristics of

a local adsorption kinetic model are lumped during repeated integrations (Agarwal et al., 2010a, 2010b). Indeed, although this adsorption rate model is rather simple, it can predict the experimental data with satisfactory accuracy (Yang & Lee, 1998). Consequently, this approximation has found widespread application in modeling fixed-bed and cyclic CO₂ adsorption processes (Hwang & Lee, 1994; Raghavan et al., 1985).

A dynamic model that included finite mass transfer resistance based on a linear driving force assumption was first developed by (Mitchell & Shendalman, 1973) for the isothermal removal of CO₂ (a strongly adsorbed component in a trace amount) from He (an inert product) using silica gel. However, the model was found to provide a poor representation of the experimental data. Cen and Yang (1985) performed separation of a five-component gas mixture containing H₂, CO, CH₄, H₂S, and CO₂ by PSA. Both equilibrium and LDF models were employed to develop a mathematical model for simulating the PSA process (model 2). The results predicted by the equilibrium model, particularly for CO₂ concentration, were in poor agreement with the experimental data, indicating the significant role of mass transfer resistance in CO₂ adsorption/desorption. The simulation results of the LDF model were in generally good agreement with the experimental data. (Raghavan et al., 1985) simulated an isothermal PSA separation of a trace amount of an adsorbable species from an inert carrier using a linear equilibrium isotherm and with the assumption of a linear driving force for mass transfer resistance (model 3). The theoretically predicted behavior of the system was shown to provide a good fit with the experimental data of (Mitchell & Shendalman, 1973) for the CO₂-Hesilica gel system. The major difference between this model and the model of (Mitchell & Shendalman, 1973) is the assumption of an inverse dependence of the effective mass transfer coefficient on the total pressure. Such behavior is to be expected for a system in which the uptake is controlled by external film or pore diffusional resistance (Raghavan et al., 1985).

Kapoor and Yang (1989) also studied the kinetic separation of a CO₂/CH₄ mixture on a carbon molecular sieve. The experimental results were simulated using a linear driving force model approach with a cycle time-dependent LDF rate coefficient (model 6). The cycle time-dependent LDF coefficient included all mass transfer resistances such as film and intraparticle diffusion and was determined by matching the model simulation results with the experimental results. However, the experimental estimates of this parameter differed considerably from the predictions of a priori correlations developed by (Nakao & Suzuki, 1983) and (Raghavan, Hassan, & Ruthven, 1986). (Diagne et al., 1996) developed a new PSA process with the intermediate feed inlet position operated with dual refluxes for separation of CO₂ dilute gas from air. They studied the influence of different CO₂ feed concentrations and feed inlet positions on CO₂ product concentration. An isothermal model based on LDF approximation was developed (model 13) to explore the effects of various combinations of the operating variables and to analyze semiquantitatively the effects of the main characteristic parameters such as the dimensionless feed inlet position and the stripping-reflux ratio. Good agreement between the model prediction and the experimental results was obtained.

In another study, low-concentration CO_2 separation from flue gas was performed by PSA using zeolite 13X as the adsorbent (Choi et al., 2003). To further assess the effects of adsorption time and reflux ratio on product purity and the recovery, dynamic modeling of the PSA process based on an LDF approximation was developed (model 19). The comparison of the numerical simulation-based and experimental results demonstrated that the model adequately describes the experimental breakthrough curves and temperature changes in the bed. (Delgado et al., 2006; Delgado et al., 2007) investigated the fixed-bed adsorption of binary gas mixtures (CO_2/He , CO_2/N_2 , and CO_2/CH_4) onto silicalite pellets, sepiolite, and a basic resin. The experimental breakthrough curves were simulated by a model based on the LDF approximation for the mass transfer that considered the energy

and momentum balances and used the extended Langmuir equation to describe the adsorption equilibrium isotherm (model 27). They proposed a lumped mass transfer coefficient instead of considering two mass transfer resistances in a bidisperse adsorbent. A comparison between the experimental and theoretical curves demonstrated that the model reproduces the experimental data satisfactorily for the different feed concentrations, flow rates, and temperatures used. More recently, (Dantas et al., 2011) studied the fixed-bed adsorption of carbon dioxide from CO₂/N₂ mixtures on a commercial activated carbon. A model based on the LDF approximation for the mass transfer that considered the energy and momentum balances was used to simulate the adsorption kinetics of carbon dioxide (model 31). They considered an overall LDF mass transfer coefficient in which the effects of film, macropore, and micropore resistances were assumed to be significant. The proposed LDF model acceptably reproduced the experimental data for the different feed concentrations/temperatures and was suitable for describing the dynamics of CO₂ adsorption from the mixtures. The importance of the external and internal mass transfer resistances was determined by performing a sensitivity analysis, which concluded that micropore resistances are not very important in the studied system. Moreover, it was deduced that, in the case of macropore resistances only, the molecular diffusivity is predominant.

If one neglects diffusion through macropores, the mass transfer rate through micropore volumes can be simplified by applying the LDF model approximation, which is mathematically equivalent to the modeling of transport through a barrier resistance confined at the micropore mouth (Cavenati et al., 2005; Grande & Rodrigues, 2007; Srinivasan et al., 1995):
$$\frac{\partial \overline{q_i}}{\partial t} = K_{\mu i} (q_i^* - \overline{q_i})$$
(2-31)

$$K_{\mu i} = \frac{1}{\frac{1}{k_{bi}} + \frac{R_c^2}{15D_{\mu i}}}$$
(2-32)

where $K_{\mu i}$ is the LDF constant for mass transfer in the micropores for component *i* (Grande & Rodrigues, 2007).

When there is no surface barrier resistance in the mouth of the micropores, the first term in the denominator of Eq. (2-32) vanishes (Cavenati et al., 2005). This model, which has been referred to as the LDFS model, is simply obtained from Eq. (2-23) if the intraparticle concentration profile of the adsorbate is assumed to be parabolic (Carta & Cincotti, 1998; Chahbani & Tondeur, 2000; Do & Rice, 1986; Liaw et al., 1979; Siahpoosh et al., 2009). The mathematically simple LDF approximation permits the direct use of the averaged adsorbed concentration in the interior of the adsorbent particle and thus eliminates the need for the integration step at the particle level, in contrast to the solid diffusion model (Chahbani & Tondeur, 2000; S. Sircar & Hufton, 2000b).

If the adsorbed-phase diffusion is neglected, a similar linear driving force model based on the gaseous phase can be used to approximate the diffusive process in macropore resistance as follows (Khalighi et al., 2012):

$$\varepsilon_{p} \frac{\partial \overline{c_{pi}}}{\partial t} + \rho_{p} \frac{\partial \overline{\overline{q_{i}}}}{\partial t} = K_{pi}(c_{i} - \overline{c_{pi}})$$
(2-33)

$$K_{pi} = \varepsilon_p \frac{15D_{pi}}{R_p^2} \frac{Bi_i}{Bi_i + 1}$$
(2-34)

where K_{pi} is the LDF constant for mass transfer in the macropores for component *i*, $\overline{c_{pi}}$ is the mean intraparticle gas phase concentration of species *i*, and $Bi_i = R_p k_{fi}/(5\varepsilon_p D_{pi})$ is the mass Biot number, which represents the ratio of internal macropore to external film resistances. As can be seen, the proposed effective LDF rate coefficient, K_{pi} , is a combination of external fluid film transport, molecular, and Knudsen diffusions in the macropores. This model, which has been referred to as the LDFG model, can be derived from the pore diffusion model, Eq. (2-14), based on the assumption of a parabolic gas-phase concentration profile in the particle (Chahbani & Tondeur, 2000; Leinekugel-le-Cocq et al., 2007; Atanas Serbezov & Sotirchos, 2001; Yang & Doong, 1985). Such a space-independent expression for the adsorption rate can transform the PDE expressing mass conservation for gas penetrating pores into an ODE, and therefore the solutions are mathematically simpler and faster than the solution of the diffusion models.

(Lai & Tan, 1991) developed approximate models for pore diffusion inside the particle with a nonlinear adsorption isotherm based on a parabolic concentration profile assumption for the summation of the gas and adsorbed phases. They developed a rate expression model that depends on the slope of the adsorption isotherm at the external surface of the sorbent. (Ding & Alpay, 2000) studied high- temperature CO₂ adsorption and desorption on hydrotalcite adsorbent at a semi-technical scale of operation. They presented a dynamic model based on a linear driving force approximation to describe intraparticle mass transfer processes (model 17). To address the importance of intraparticle mass transfer resistances during different steps of operation, they also developed an adsorption model based on ILE assumption between the gas and adsorbed phases. Overall, they concluded that although the ILE model failed to give an adequate description of the desorption kinetics, the LDF model based on pore diffusion and accounting for the non-linearity of the isotherm provides an adequate approximation of the adsorption and desorption processes. (Grande & Rodrigues, 2008) studied the operation of an electric swing adsorption process for low-concentration CO₂ removal from flue gas streams using an activated carbon honeycomb monolith as an adsorbent. To explore the dynamics behavior of the system, the authors developed a mathematical

model that included bidisperse resistances within the porous structure of the monolith (model 29). A rigorous description and a linear driving force approximation were employed for macropore and micropore diffusion, respectively. Adsorption/desorption breakthrough experiments were performed to determine the validity of the proposed mathematical model. A comparison of simulated breakthroughs and experimental data showed that the dynamic model incorporating mass, energy, and momentum balances agreed well with the experimental results.

If macropore or the adsorbed-phase diffusion cannot be ignored, the mass transfer rate expression can be expressed using a double LDF model, through which the macropore and the micropore diffusion are both represented by LDF approximations taken in series (Da Silva et al., 1999; Doong & Yang, 1987; Kim, 1990; Leinekugel-le-Cocq et al., 2007; Mendes, Costa, & Rodrigues, 1996). Cavenati et al. (2005) studied the separation of a methane-carbon dioxide mixture in a column packed with bidisperse adsorbent (carbon molecular sieve 3K). To reduce the computational time required for the simulations, macropore and micropore diffusion equations were described using a bi-LDF simplification instead of the mass balances in macropores and in micropores (model 24). They assumed that the macropore diffusivity and surface barrier resistance at the mouth of the micropore are not a function of the adsorbed phase concentration, whereas the Darken law describes micropore diffusivity dependence with concentration. To confirm the validity of the mathematical model and the proposed bi-LDF approximation for the prediction of experimental data, a fixed-bed experiment of the binary mixture was performed. The results indicated that the proposed mathematical model was able to adequately predict the behavior of the binary mixture in a fixed bed. Leinekugel-le-Cocq et al. (2007) presented a simplified intraparticle model based on a non-isothermal double LDF approximation to simulate breakthroughs of a CH₄/CO₂ mixture in a fixed bed of bidisperse adsorbent (5A zeolite). A bidisperse double LDF model was proposed that

considered both macropore and micropore diffusion by LDF approximations taken in series (model 28). A comparison of the model predictions with the experimental data revealed that the approximated model proposed in this study yields a good representation of the intraparticle mass transfer.

2.3.3 Energy balance

As mentioned earlier, adsorption is an exothermic process, and temperature changes may affect the adsorption equilibrium relation and, in some cases, adsorption rates. Therefore, the effects of heat generation and heat transfer in the adsorbent bed must be considered to accurately predict packed column dynamics. Some publications have addressed nonisothermal adsorption of CO_2 in a fixed bed either as a single adsorbate or multicomponent mixture, which are summarized in Table 2.1. The amplitude of the temperature variation depends primarily on the heat of adsorption, the throughput, and the heat transfer characteristics of the packed adsorbed column (Farooq, Hassan, & Ruthven, 1988; Ruthven, 1984). The generated heat is conducted to the surface of the particles and then is transferred to the gas phase by a convection mechanism. To take into account the energy transfer, three different control volumes may be considered: gas, solid, and column wall.

2.3.3.1 Gas phase energy balance

The variation of gas temperature with respect to time arises from the transfer of energy due the axial conductive solid-phase heat flux to the gas phase along the z direction as well as the energy transfer by convection through the bed due to the bulk movement of gas. Therefore, the energy balance for the gas phase, which includes the heat transfer between the flowing fluid and the surface of the solid adsorbent through a laminar film

and the energy transferred to the wall of the column, may be written as the following (Rezaei & Grahn, 2012; Ruthven, 1984):

$$-\lambda_{L}\frac{\partial^{2}T_{g}}{\partial z^{2}} + \rho_{g}C_{g}\frac{\partial(uT_{g})}{\partial z} + \rho_{g}C_{g}\frac{\partial T_{g}}{\partial t} + (\frac{1-\varepsilon_{b}}{\varepsilon_{b}})h_{f}a_{s}(T_{g}-T_{s}) + \frac{4h_{w}}{\varepsilon_{b}d_{\text{int}}}(T_{g}-T_{w}) = 0 \quad (2-35)$$

The above equation is used to find the distribution of gas temperature along the bed, where T_g represents the bulk gas temperature; C_g is the heat capacity of the gas; ρ_g is the bulk density of the gas; h_f denotes the film heat transfer coefficient between the gas and the adsorbent; a_s expresses the ratio of the particle external surface area to volume; T_s denotes the solid temperature; h_w is the internal convective heat transfer coefficient between the gas and the column wall; d_{int} is the column internal diameter; T_w denotes the wall temperature; and λ_L is the effective axial heat dispersion, which can be estimated using the following correlation (Grande & Rodrigues, 2005; Wakao & Funazkri, 1978; Yang, 1987):

$$\frac{\lambda_L}{k_g} = 7 + 0.5 PrRe \tag{2-36}$$

where k_g is the thermal conductivity of the gas mixture and Pr is the Prandtl number.

The Chilton–Colburn analogy can be applied (in analogy with Eq. (2-11)) to estimate the convective film heat transfer coefficient between the gas and the adsorbent, h_f (Chilton & Colburn, 1934). It can be estimated through the following correlation, which is particularly applicable at higher Reynolds numbers (Wakao, Kaguei, & Funazkri, 1979):

$$Nu = \frac{2h_f R_p}{k_g} = 2 + 1.1 P r^{1/3} R e^{0.6}$$
(2-37)

The following correlation can be applied for the estimation of the internal convective heat transfer coefficient, h_w (Dantas et al., 2009; Dantas et al., 2011; Dantas et al., 2011):

$$Nu_{w} = \frac{h_{w}d_{\text{int}}}{k_{g}} = 12.5 + 0.048Re$$
(2-38)

The following boundary conditions are assumed:

$$\lambda_{L} \frac{\partial T_{g}}{\partial z} \Big|_{z=0} = -\rho_{g} C_{g} u \Big|_{z=0} \left(T_{g} \Big|_{z=0^{-}} - T_{g} \Big|_{z=0} \right)$$

$$\partial T \Big| \qquad (2-39)$$

$$\left. \frac{\partial I_g}{\partial z} \right|_{z=L} = 0 \tag{2-40}$$

where $T_{g}|_{z=0^{-}}$ is feed temperature.

2.3.3.2 Solid phase energy balance

For the solid phase, a separate energy balance equation can be assumed that considers the accumulation term, the film heat transfer term, and the heat generated by the adsorption of the adsorbate, as shown by the following (Do, 1998b):

$$\rho_p C_s \frac{\partial T_s}{\partial t} = h_f a_s (T_g - T_s) + \sum_{i=1}^n (-\Delta H_i) \frac{\partial \overline{q_i}}{\partial t}$$
(2-41)

where ρ_p is the particle density, C_s represents the heat capacity of the adsorbent and (- ΔH_i) is the isosteric heat of adsorption for the *i* component at zero coverage, which can be calculated by using the Clausius-Clapeyron equation (Yang, Lee, & Chang, 1997).

When the external heat transport limitations are negligible, it is a reasonable approximation to neglect the occurrence of temperature gradients in the particles and consider the gas phase and the surface of the adsorbing particles to be isothermal (Ribeiro, Grande, Lopes, Loureiro, & Rodrigues, 2008). In this case, a single temperature equation, which is obtained from the overall local balance in the bed (combining Eqs. (2-35) and (2-41) into one), is sufficient to describe the energy transport in the bed (Hu & Do, 1995; Atanas Serbezov & Sotirchos, 1998). However, some experimental studies have demonstrated the occurrence of temperature differences between the gas phase and the

surface of the adsorbent particles (Haul & Stremming, 1984; Lee & Ruthven, 1979; Ruthven, Lee, & Yucel, 1980). This situation is particularly relevant when an adsorption process occurs at a relatively high rate such that the time needed for the released heat to be transported to the bulk phase may not be sufficient and, consequently, significant temperature gradients are encountered in the interior of the particle (Atanas Serbezov & Sotirchos, 1998).

2.3.3.3 Wall energy balance

Finally, the energy balance for the column wall, which includes the wall heat-transfer to the external environment and to the gas phase inside the column, can be expressed as follows (Da Silva et al., 1999):

$$\rho_{w}C_{w}\frac{\partial T_{w}}{\partial t} = h_{w}a_{w}(T_{g} - T_{w}) + Ua_{a}(T_{\infty} - T_{w})$$

$$(2-42)$$

where C_w and ρ_w represent the heat capacity and the density of the column wall, respectively; a_w represents the ratio of the internal surface area to the volume of the column wall (Da Silva et al., 1999; Dantas et al., 2009; Huang & Fair, 1988); a_a denotes the ratio of the external surface area to the volume of the column wall; U is the external overall heat transfer coefficient from the wall to ambient air; and T_∞ is the ambient temperature.

The external overall heat transfer coefficient, U, can be estimated through following correlation (Incropera & Witt, 1996):

$$\frac{1}{U} = \frac{1}{h_w} + \frac{d_{\text{int}}}{k_w} \ln(\frac{d_{ext}}{d_{\text{int}}}) + \frac{d_{\text{int}}}{d_{ext}h_{ext}}$$
(2-43)

where k_w is the column wall conductivity; d_{ext} is the column external diameter; and h_{ext} is the external convective heat transfer coefficient that can be estimated using the following correlation (Incropera & Witt, 1996):

$$\frac{h_{ext}L}{k_{ext}} = 0.68 + \frac{0.67Ra^{1/4}}{\left[1 + (0.492/Pr)^{9/12}\right]^{4/9}}$$
(2-44)

In the above equation, k_{ext} is the column external air conductivity and $Ra = g\beta \frac{(T_w - T_w)}{v\alpha} L^3$ is the Rayleigh number, where g is the gravity acceleration; β is the thermal expansion coefficient; and v and α are the air kinematic viscosity and thermal diffusivity at the film temperature ($(T_w + T_w)/2$), respectively.

If the area of heat transfer from the fluid to the wall is an order of magnitude larger than the area in the axial direction, the contribution of the axial heat conduction along the column wall can be neglected (Ahn, Lee, Seo, Yang, & Baek, 1999; Ahn et al., 2001; Mohamadinejad et al., 2000). In the above wall energy balance equation (Eq. 2-42), the resistance of the metal wall to radial heat transfer has been considered in the overall heat transfer coefficient, U, which makes this value lower than that of the individual convection heat transfer coefficient between the wall and the surroundings (Grande & Rodrigues, 2005).

In an industrial-scale process in which the column length-to-diameter ratio is not large, the heat loss through a wall and heat accumulation in the wall are considered negligible in comparison to the amount of heat caused by the heat of adsorption, resulting in operation close to adiabatic behavior (Lee et al., 1999). In such a situation, the overall heat transfer coefficient, U, and therefore the last term in Eq. (2-42) can be dropped (Bastos-Neto, Moeller, Staudt, Bohm, & Glaser, 2011). In the case of an isothermal system, an instantaneous thermal equilibrium is assumed to exist between the gas and solid phases ($T_g=T_s$) or between the gas, solid, and column wall ($T_g=T_s=T_w$), depending on the system's conditions. Under such assumptions, the original three energy balances of the complete model discussed above (Eqs. 2-35, 2-41, and 2-42) are reduced to two equations and one equation, respectively.

2.3.4 Momentum balance

As the bulk fluid flows through the void spaces between adsorbent particles, it experiences a pressure drop due to viscous energy losses and a drop in kinetic energy. The momentum balance considers the terms of pressure drop and velocity changes across the packed bed and relates them by the Ergun equation (Alpay, Kenney, & Scott, 1993; Z. P. Lu, Loureiro, Rodrigues, & LeVan, 1993; Yang, Park, Chang, Ko, & Lee, 1998):

$$\frac{\partial P}{\partial z} = -K_D u - K_V u^2 \tag{2-45}$$

where K_D and K_V are parameters corresponding to the viscous and kinetic pressure loss terms, respectively. For low particle Reynolds numbers (e.g., <5), the kinetic contribution to the total pressure loss is negligible, and Eq. (2-45) reduces to Darcy's law (Ding & Alpay, 2000). Semi-empirical relationships for K_D and K_V have been derived by Ergun (Macdonald, El-Sayed, Mow, & Dullien, 1979) as follows:

$$K_{D} = \frac{150\mu(1-\varepsilon_{b})^{2}}{\varepsilon_{b}^{3}d_{p}^{2}}$$
(2-46)

$$K_{V} = \frac{1.75(1 - \varepsilon_{b})\rho_{g}}{\varepsilon_{b}^{3}d_{p}}$$
(2-47)

where *P* is the total pressure, μ is the gas mixture viscosity, and d_p is the particle mean diameter.

The existence of the pressure drop in packed beds affects the system performance by reducing the working capacity. Conventional gas separation processes using packed beds of beads or granules suffer predominantly from high pressure drop and mass transfer resistance when higher throughputs are required, leading to lower productivity and recovery and higher power consumption. In order to overcome these drawbacks, novel adsorbent structures such as monoliths, laminates, and foams structures are considered as alternate candidates for the replacement of conventional adsorbent configurations as they

offer the advantage of allowing increased throughput and reduced pressure drops (Rezaei & Webley, 2010). In recent years, monolithic adsorbents have garnered considerable attention for CO₂ capture from the gas streams (Brandani, Rouse, Brandani, & Ruthven, 2004; Grande & Rodrigues, 2008; Mosca, Hedlund, Ridha, & Webley, 2008; Ribeiro et al., 2008). A detailed review of these alternate structures and their merits was recently conducted by (Rezaei & Webley, 2010).

The pressure drop of the monolithic and laminate structures can be expressed by the following equation using the Hagen–Poiseuille equation (Rezaei & Webley, 2009):

$$\frac{\Delta P}{L} = \frac{32\mu}{d^2 \varepsilon_b} u \tag{2-48}$$

where d is the channel dimension.

In the case of foam adsorbents, the pressure drop can be estimated using the equation presented by Richardson et al. (Richardson, Peng, & Remue, 2000):

$$\frac{\Delta P}{L} = \frac{\alpha s^2 (1-\varepsilon)^2}{\varepsilon_b^3} \mu u + \frac{\beta s (1-\varepsilon)}{\varepsilon_b^3} \rho_g u^2$$
(2-49)

where the first term represents flow resistance from viscous forces, the second term is the inertial or turbulent contributions, and α , β , *s* can be calculated using following expressions:

$$\alpha = 9.73 \times 10^2 d_p^{0.743} (1 - \varepsilon_b)^{-0.0982}$$
(2-50)

$$\beta = 3.68 \times 10^2 d_p^{-0.7523} (1 - \varepsilon_b)^{0.07158}$$
(2-51)

$$s = \frac{12.979 \left[1 - 0.971 (1 - \varepsilon_b)^{0.5}\right]}{d_p (1 - \varepsilon_b)^{0.5}}$$
(2-52)

Clearly, other momentum balance equations that model the pressure drop in a structured packing could also be applied.

A detailed description of the differential mass and heat balance equations for the different adsorbent structures can be found elsewhere (Grande et al., 2006; Rezaei & Grahn, 2012; Rezaei & Webley, 2009).

CHAPTER 3: MATERIALS AND METHODS

3.1 Introduction

The following sections give information about the materials and the procedures used in developing an efficient adsorbent for CO_2 separation. Details of adsorbent modification, characterization and evaluation of their adsorption performance are also described here. Commercial granular activated carbon (GAC) used in this research was supplied by BRAVO GREEN SDN BHD. All the gases were of purity greater than 99.99% and were supplied by Air Products.

3.2 The application of response surface methodology to optimize the amination of activated carbon for the preparation of carbon dioxide adsorbents

3.2.1 Adsorbent materials

The commercial palm-shell-based granular activated carbon (referred to as GAC) was used as a starting material for the preparation of CO_2 adsorbent. The precursor was ground and passed through a US mesh size 20-35 (850-500 µm) prior to all further treatments. To eliminate fines, the GAC was then thoroughly washed with distilled-deionized water (DDW), dried at 105 °C for 24 h to remove moisture, and stored in a desiccator under vacuum. The following is a brief outline of the process used to modify these samples.

3.2.2 Ammonia modification

In this work, two different types of adsorbents were used as starting materials: GACs that were either pre-heat-treated under nitrogen or pre-oxidized with air. These adsorbents are referred to as HTA and OXA, respectively. To produce carbon adsorbents with different amounts and types of N-containing functionalities, the ammonia treatment was performed for different times and at different temperatures for both adsorbents. The ammonia modification of HTA was achieved by placing approximately 1 g of activated carbon in

a tubular quartz reactor under a 75 cm³min⁻¹ flow of nitrogen. The temperature of the reactor was increased at a rate of 10 °C min⁻¹ until it reached a final modification temperature (selected temperatures from 200 to 1000 °C). The feed gas (nitrogen) was then changed to reactive gas (pure ammonia) at the same flow rate. During this treatment, the samples were maintained at the final temperature for a period of time (between 1 and 3 h) and then cooled down to 100 °C at the same cooling rate (10 °C min⁻¹). Finally, the flow was changed back to N₂ until room temperature was reached. For the modification of OXA samples, approximately 1 g of sample was placed into a quartz reactor. Air was introduced, and the reactor was heated (10 °C min⁻¹) to 400 °C. This temperature was selected for air oxidation because it was reported to be the optimum temperature at which to add surface oxygen groups (Mangun, Benak, Daley, & Economy, 1999; Polovina, Babić, Kaluderović, & Dekanski, 1997; Saha, Tai, & Streat, 2001). The sample was held at this temperature for 2.5 h and then cooled down to room temperature before the gas was switched to pure ammonia. Both the air and NH₃ flow rates were set to 75 cm³ min⁻ ¹. Next, the sample was heated in an ammonia atmosphere with a heating rate of 10 $^{\circ}$ C min⁻¹ until the desired temperature was reached. The ammonia treatment was then performed with the same reaction conditions described above. The schematic diagram of the apparatus for the ammonia modification is shown in Figure 3.1.



Figure 3.1: Schematic diagram of the experimental setup for the ammonia modification MFC, mass flow controller; TC, thermocouple

3.2.3 Experimental design and method of analysis

The conventional approach for optimizing multifactor experiments, one-factor-at-a-time method (OFAT), involves varying a single factor while all other factors are kept fixed for a particular set of experiments. This approach is time-consuming and incapable of reaching the true optimum because possible interaction effects between the variables are not taken into consideration (Frigon & Mathews, 1997; Ghafari, Aziz, Isa, & Zinatizadeh, 2009; Kusic, Jovic, Kos, Koprivanac, & Marin, 2010; Montogomery, 2005). The RSM statistical method has been proposed to optimize parameters with a minimal number of experiments and to analyze the interaction between the parameters (Azargohar & Dalai, 2005; Khuri & Cornell, 1996; Myer & Montogomery, 2002). RSM has already proven to be a reliable statistical tool in the investigation of chemical treatment processes (Can, Kaya, & Algur, 2006; Ghafari et al., 2009; Gönen & Aksu, 2008; Hameed, Tan, & Ahmad, 2008; Körbahti & Rauf, 2008a, 2008b; Kusic et al., 2010; Ölmez, 2009; Sahu, Acharya, & Meikap, 2009; Zhanmei Zhang & Zheng, 2009). Accordingly, in this study, Design-Expert 8.0.5 software (trial version) was used to generate experimental designs, statistical analyses, and regression models and to optimize the sorbent modification

conditions. Central composite design with a quadratic model and $\alpha = 0.5$ was coupled with response surface methodology to statistically design experiments and analyze data. In the present work, the effects of three independent variables, including two numerical variables (i.e., amination temperature between 200 and 1000 °C (X_1) and amination times between 1 and 3 h (X_2)), both at five levels, and one categorical variable (i.e., the use of HTA or OXA as the starting material (X_3) were investigated using central composite design. These variables, along with their respective regions of interest, were chosen based on the literature and preliminary studies (Pevida, Plaza, et al., 2008; Plaza et al., 2009; Plaza, Rubiera, et al., 2010; Przepiórski et al., 2004; Shafeeyan et al., 2010). Table 3.1 presents the range and levels of the independent numerical variables in terms of actual and coded values. The numerical variables in actual form (X_1, X_2) were converted to their coded form (x_1, x_2) using the following equation (Amini, Younesi, & Bahramifar, 2009; Banik, Santhiagu, & Upadhyay, 2007; Soloman, Ahmed Basha, Velan, Balasubramanian, & Marimuthu, 2009):

$$x_i = \frac{(X_i - X_0)}{\Delta X} \tag{3-1}$$

where x_i is the dimensionless coded value of the *i*th independent variable, and X_i , X_0 , and ΔX correspond to the actual value, actual value at the center point, and the step change of the *i*th variable, respectively.

Table 5.1: Independent numerical variables and their levels (actual and coded)					
Symbol	Coded variable levels				
	-1	- 0.5	0	+0.5	+1
X_1	200	400	600	800	1000
X_2	1	1.5	2	2.5	3
	Symbol X ₁ X ₂	SymbolCoded -1 X_1 X_2 1	SymbolCoded variable X_1 200400 X_2 11.5	SymbolCoded variable levels -1 Coded variable levels 0 X_1 200400600 X_2 11.52	Coded their levels (actual and codeSymbolCoded variable levels -1 -0.5 0 X_1 200 400 600 800 X_2 1 1.5 2 2.5

Table 2.1. Independent numerical variables and their levels (actual and acded)

For each categorical variable, 13 experiments were performed in randomized order (to minimize the effects of the uncontrolled factors), including five replications at the design center to obtain an accurate estimation of the experimental error (Table 3.2). The CO₂ sorption capacity (Y_S) and desorption capacity (Y_D) of the resulting adsorbents were taken as the two responses of the designed experiments. A quadratic polynomial equation was developed to predict the dependent variables (chosen responses) as a function of independent variables (factors) and their interactions, as given by Eq. (3-2) (Aksu & Gönen, 2006; Can et al., 2006; Yetilmezsoy, Demirel, & Vanderbei, 2009):

$$Y = \beta_0 + \sum_{i=1}^n \beta_i x_i + \sum_{i=1}^n \beta_{ii} x_i^2 + \sum_{i=1}^n \sum_{j=1}^n \beta_{ij} x_i x_j + \varepsilon$$
(3-2)

In this equation, Y is the predicted responses; β_0 is the constant; x_i , x_j are the coded value of the independent variables; β_i is the linear term coefficient; β_{ii} is quadratic term efficient; β_{ij} is the interaction term coefficient, ε is the random error, and n is the number of factors studied.

An analysis of variance (ANOVA) was applied to evaluate the fitness of the model and identify the interactions between the amination variables and the responses. The goodness of fit of the polynomial model was expressed by the coefficients of determination, R^2 and R^2_{adj} through Eqs. (3-3) and (3-4), respectively (Körbahti & Rauf, 2008b).

$$R^{2} = 1 - \frac{SS_{residual}}{SS_{model} + SS_{residual}}$$
(3-3)

$$R_{adj}^{2} = 1 - \frac{SS_{residual}/DF_{residual}}{(SS_{model} + SS_{residual})/(DF_{model} + DF_{residual})}$$
(3-4)

Here, SS is the sum of squares, DF is the degrees of freedom.

The statistical importance of the model was checked by determining the model's adequate precision ratio using Eqs. (3-5) and (3-6), and by an *F*-test in the program (Körbahti & Rauf, 2009).

Adequate Precision =
$$\frac{\max(Y) - \min(Y)}{\sqrt{\overline{V}(Y)}}$$
(3 - 5)

$$\overline{V}(Y) = \frac{1}{n} \sum_{i=1}^{n} \overline{V}(Y) = \frac{p\sigma^2}{n}$$
(3-6)

Here, *Y* is the predicted response, *p* is the number of model parameters, σ^2 is the residual mean square (presented in Tables 4-1 and 4-2 for each responses), and n is the number of experiments.

The insignificant terms were eliminated after the F-test, and the final model was obtained. Three-dimensional response surfaces were presented to visualize the individual and interactive effects of the independent variables. Finally, the independent variable optimum values for maximizing the amount of studied responses were determined using the Design-Expert program 8.0.5 trial version. Additional experiments at optimum conditions were performed to verify the prediction provided by the regression model.

Run	Sample	Coded variables		les	Actual variables			Responses	
		<i>x</i> ₁	<i>x</i> ₂	<i>X</i> 3	$X_1(^{\circ} C)$	X_2 (h) X_3 Y_S (mg/g) Y_D (v		$Y_{\rm D}({\rm wt.\%})$	
1	OXA-600-2.5	0	0.5	1	600	2.5	OXA	28.58	93.1
2	HTA-400-2	-0.5	0	-1	400	2	HTA	24.99	95.7
3	HTA-600-2	0	0	-1	600	2	HTA	26.16	94.0
4	OXA-600-2	0	0	1	600	2	OXA	28.08	93.3
5	OXA-200-3	-1	1	1	200	3	OXA	21.81	98.1
6	HTA-600-1.5	0	-0.5	-1	600	1.5	HTA	25.6	94.5
7	HTA-1000-1	1	-1	-1	1000	1	HTA	26.05	94.1
8	OXA-600-2	0	0	1	600	2	OXA	27.95	93.4
9	HTA-600-2	0	0	-1	600	2	HTA	26.08	94.0
10	OXA-200-1	-1	-1	1	200	1	OXA	20.17	99.4
11	OXA-600-2	0	0	1	600	2	OXA	27.74	93.6
12	HTA-800-2	0.5	0	-1	800	2	HTA	27.12	92.9
13	HTA-600-2.5	0	0.5	-1	600	2.5	HTA	26.43	93.7
14	OXA-600-2	0	0	1	600	2	OXA	27.88	93.5
15	HTA-1000-3	1	1	-1	1000	3	HTA	26.31	93.5

Table 3.2: Experimental design layout and experimental results of the responses

(continued on next page)

	'Table 3.2, continued'								
Run Sample		Coded variables			Actual variables			Responses	
	couc	<i>x</i> ₁	<i>x</i> ₂	<i>X3</i>	$X_1(^{\circ} C)$	X_{2} (h)	<i>X</i> ₃	$Y_{\rm S}~({\rm mg/g})$	$Y_{\rm D}({\rm wt.\%})$
16	OXA-600-2	0	0	1	600	2	OXA	28.19	93.3
17	HTA-600-2	0	0	-1	600	2	HTA	25.84	94.2
18	HTA-200-3	-1	1	-1	200	3	HTA	22.65	97.6
19	OXA-800-2	0.5	0	1	800	2	OXA	30.04	91.9
20	OXA-1000-3	1	1	1	1000	3	OXA	28.73	92.8
21	HTA-600-2	0	0	-1	600	2	HTA	25.76	94.3
22	OXA-1000-1	1	-1	1	1000	1	OXA	28.13	93.2
23	HTA-600-2	0	0	-1	600	2	HTA	25.97	94.1
24	HTA-200-1	-1	-1	-1	200	1	HTA	21.08	98.8
25	OXA-400-2	-0.5	0	1	400	2	OXA	25.97	95.3
26	OXA-600-1.5	0	-0.5	1	600	1.5	OXA	27.26	94.0

3.2.4 CO₂ adsorption/desorption measurements

The CO₂ adsorption/desorption performance of the modified adsorbents was evaluated in a TGA/SDTA851 thermogravimetric analyzer (TGA) at atmospheric pressure. An isothermal CO₂ adsorption and desorption of the adsorbents was performed. In a typical adsorption/desorption test, approximately 10 mg of the adsorbent was placed in a small pan and heated to 105 °C at a heating rate of 2 °C min⁻¹ under a flow of pure N₂ (75 cm³min⁻¹). The sample was held isothermally (30 min) until its weight became stable. The gas was then switched to pure CO₂ at the same flow rate. The sample was held isothermally at 105 °C under a flow of pure CO₂ for 90 min to achieve complete saturation. This temperature was selected because it is the expected temperature in a power plant flue gas desulfurization unit (Maroto-Valer et al., 2005; Shafeeyan, Daud, Houshmand, & Arami-Niya, 2011; Xu et al., 2002). The weight increase during this stage was taken as the CO₂ capture capacity of the sample at 105 °C. After adsorption, the gas was switched back to N₂ at the same flow rate while the temperature was maintained at 105 °C. This was done to study the desorption capacity of the adsorbents. For all of the samples, a desorption time of 30 min was selected because this is the time necessary to fully desorb CO₂ from untreated carbon. Cyclical adsorption and desorption were used to evaluate the stability of the adsorbent. The CO2 adsorption capacity (mg-CO2/gadsorbent) and desorption capacity (%) were calculated using the following equations:

$$CO_2 adsorption capacity (mg/g) = \frac{m_1 - m_0}{m_0} \times 1000$$
(3 - 7)

$$CO_2$$
 desorption capacity (%) = $\frac{m_1 - m_2}{m_1 - m_0} \times 100$ (3 - 8)

where m_0 , m_1 , and m_2 are the weight of the adsorbent (mg) after initial heat treatment in N₂, after CO₂ adsorption, and after desorption in N₂, respectively.

3.3 A semi-empirical model to predict adsorption equilibrium of carbon dioxide on ammonia modified activated carbon

3.3.1 Adsorbent materials

Earlier, our group optimized the amination conditions of activated carbon adsorbents in an effort to maximize their CO₂ adsorption/desorption capacities (Shafeeyan et al., 2012). The optimal adsorbent (a pre-oxidized sample that was aminated at 425 °C for 2.12 h) exhibited promising adsorption/desorption performance during cyclical operations, making it suitable for practical applications. Therefore, in this work, the optimal adsorbent (referred to as OXA-GAC) was used as a starting material. Further details on the adsorbent preparation and modification can be found elsewhere (Shafeeyan, Daud, Houshmand, & Arami-Niya, 2011; Shafeeyan et al., 2012).

3.3.2 Equilibrium CO₂ adsorption measurements

CO₂ adsorption isotherms of the modified and untreated activated carbon samples were measured using a Micromeritics ASAP 2020 instrument, which is a static volumetric apparatus. The equilibrium experiments were conducted at temperatures of 30, 45 and 60 °C and at pressures up to 1 atm, a typical operating range in adsorption units for CO₂ capture from power plants. The adsorption temperature was controlled by circulating water from a thermostatic bath (Jeio Tech, model: Lab Companion RW 0525G) with an uncertainty of \pm 0.1 K. Using the volumetric method with P–V–T measurements, we determined the total quantity of gas introduced into the adsorption system and the quantity that remained in the system after reaching adsorption equilibrium. Prior to the CO₂ adsorption measurements, known amounts of samples (e.g., 50-100 mg) were loaded into the sample tube and degassed by reducing the pressure to 10⁻⁵ mmHg at 473 K for 15 h to dehydrate and desorb any adsorbed gases. CO₂ was then purged into the sample cell, and the change in adsorption volume as a function of CO_2 partial pressure was recorded. The final adsorption amount at the terminal pressure and temperature was considered to be the adsorption equilibrium amount.

3.3.3 Adsorption isotherm equations

For each modified and untreated activated carbon adsorbent, three isotherms were measured at 30, 45, and 60 °C and at pressures up to 1 atm. To apply the adsorption equilibrium data to a specific gas-separation application, an accurate mathematical representation of the adsorption equilibrium is required (Esteves, Lopes, Nunes, & Mota, 2008). The equilibrium adsorption isotherm may provide useful insight into the adsorbate-adsorbent interactions and the surface properties and affinities of the adsorbent (Foo & Hameed, 2010). For this purpose, three different pure-species isotherm models the Freundlich, Sips, and Toth isotherm equations—were used to correlate experimental equilibrium results. These three isotherm models are frequently used for modeling gasseparation processes.

The Freundlich isotherm can be applied to non-ideal adsorption on heterogeneous surfaces for a multilayer adsorption with a non-uniform distribution of adsorption heat. It is represented as (LeVan et al., 1999):

$$q = K_F P^{1/m_F} \tag{3-9}$$

The Sips isotherm is a combination of the Freundlich and the Langmuir isotherm models for predicting the behavior of heterogeneous adsorption systems. At low surface coverages, it reduces to the Freundlich equation, whereas, at high adsorbate concentrations, it predicts a monolayer adsorption capacity that is typical of the Langmuir isotherm. The Sips equation is given by (LeVan et al., 1999):

$$q = \frac{q_m \left(K_s P\right)^{1/m_s}}{1 + \left(K_s P\right)^{1/m_s}}$$
(3-10)

The Toth isotherm is a Langmuir-based isotherm derived from potential theory and is commonly used to describe heterogeneous adsorption processes. It considers a quasi-Gaussian distribution of site affinities. The Toth isotherm is written as (LeVan et al., 1999):

$$q = \frac{q_m K_T P}{\left(1 + \left(K_T P\right)^{m_T}\right)^{1/m_T}}$$
(3-11)

In the above mentioned equations, q represents the adsorbed concentration, P is the equilibrium pressure, q_m is the maximum loading capacity, K_i is the equilibrium constant (K_F) or the affinity parameter $(K_S$ and $K_T)$, and m_i $(m_F, m_S, and m_T)$ is the parameter that refers to the system heterogeneity.

The next step is to fit the experimental equilibrium adsorption data to the aforementioned isotherm models and adjust each set of isotherm parameters. Because of the inherent bias associated with transforming non-linear isotherm equations into linear forms, several authors have proposed using a non-linear regression procedure (Ho, Porter, & McKay, 2002; Porter, McKay, & Choy, 1999). Accordingly, in the current study, the parameters of the isotherm equations for each temperature were obtained by non-linear regression analysis using the Marquardt-Levenberg algorithm implemented in SigmaPlot software version 12.0 (Systat Software Inc., USA), with a user-defined equation added to the Regression Wizard. To quantify and compare the goodness of fit of the above isotherm models to the experimental data and adjust each set of isotherm constants, two different error functions, the average relative error (ARE) and nonlinear regression coefficient (R^2), were evaluated. The average relative error, which measures the deviation between the

experimental equilibrium data and the fitted model values, was calculated according to the following equation (Foo & Hameed, 2010):

$$ARE(\%) = \frac{100}{n} \sum_{i=1}^{n} \left| \frac{q_{meas} - q_{cal}}{q_{meas}} \right|_{i}$$
(3-12)

where n is the number of data points at a given temperature, and subscripts "*meas*" and "*cal*" refer to the measured and calculated values of q, respectively.

3.4 Modeling of carbon dioxide adsorption onto ammonia-modified activated carbon: Kinetic analysis and breakthrough behavior

3.4.1 Adsorbent materials

Commercial granular palm shell-based activated carbon (referred to as GAC) was used as the starting material for this study. The GAC was enriched with nitrogen using oxidation followed by amination at 425 °C for 2.12 h (referred to as OXA-GAC). The selected modification condition was optimum to provide the adsorbent with promising adsorption/desorption capacity and stability during cyclical operations (Shafeeyan et al., 2012). Further details on the adsorbent preparation and modification can be found elsewhere (Shafeeyan, Daud, Houshmand, & Arami-Niya, 2011; Shafeeyan et al., 2012).

3.4.2 Kinetic adsorption measurements

The kinetics of CO₂ adsorption on the GAC and OXA-GAC adsorbents were studied using a TGA/SDTA851 thermogravimetric analyzer at atmospheric pressure. Before each kinetic measurement, the adsorbent sample (approximately 10 mg) was pretreated under a 100 ml min⁻¹ flow of pure nitrogen at 110 °C for 1 h to guarantee removal of moisture and other dissolved gases. The temperature was decreased to the desired temperature (selected temperatures ranging from 30 to 60 °C) and the nitrogen gas was then changed to pure CO₂ at the same flow rate. The sample was maintained at the final temperature under a constant flow of CO_2 until the rate of the measured mass change of the sample approached zero, implying that thermodynamic equilibrium was attained. The adsorption capacity of the adsorbent is expressed in mol CO_2/kg adsorbent and is determined from the weight change of the sample after the introduction of CO_2 .

3.4.3 Fixed-bed adsorption experiments

The experimental breakthrough apparatus shown in Fig. 3.2 consists of a stainless steel column 0.2 m in length and 0.01 m in internal diameter packed with the adsorbent sample. Two fine meshes were placed at the bottom and top of the column to ensure a uniform gas distribution and to retain the adsorbent particles in the bed. The CO₂ and N₂ flow rates were regulated using two digital mass-flow controllers (Aalborg ® GFC17) with an accuracy and repeatability of 1% and 0.1% full scale, respectively. To achieve the desired inlet composition of the feed gas (15% CO_2/N_2 , v/v), the gas flow rates of CO_2 and N_2 were adjusted by mass flow controllers before entering the column to attain a constant total flow of 50 ml min⁻¹ (flow rates: $CO_2 = 7.5$ ml min⁻¹ and $N_2 = 42.5$ ml min⁻¹) and 100 ml min⁻¹ (flow rates: $CO_2 = 15$ ml min⁻¹ and $N_2 = 85$ ml min⁻¹) of feed gas for the first and second series of experiments, respectively. Continuous monitoring of the CO2 concentration at the bed exit (after every 20 s) was performed using a CO₂ analyzer (Bacharach Inc., IEQ Chek-Indoor Air Quality Monitor) equipped with a thermal conductivity detector that was connected to a data recording system. The temperature was measured using a K-type thermocouple with an accuracy of ± 1 K located inside of the solids bed at a height of 0.05 m above the exit end of the column. The thermocouple was connected to a temperature data logger to monitor the temperature during the experiment. The column was placed inside of a tubular furnace with a programmable temperature controller.

In a typical experiment, a sample of ca. 5.5 g of adsorbent was packed into the column. At the beginning of the experiments, the adsorbent bed was pretreated by purging with a 300 ml min⁻¹ flow of N₂ at 130 °C and atmospheric pressure for at least 3 h to ensure complete desorption of volatile compounds and pre-adsorbed gases. Subsequently, the bed was cooled to the desired adsorption temperature, and the nitrogen flow rate was reduced to the predefined value according to the experimental conditions (i.e., 42.5 and 85 ml min⁻¹). The CO₂ was then introduced at flow rates of 7.5 and 15 ml min⁻¹, resulting in a feed gas mixture containing 15 vol % CO₂ in N₂. The feed gas was fed to the column through a three-way value in an upward flow pattern at the predefined flow rates. The adsorption breakthrough experiments were performed at the temperatures of 30, 45, and 60 °C while changing the feed gas flow rate from 50 to 100 ml min⁻¹ under 1 atm total pressure. The column adsorption temperatures and feed gas concentration were selected based on the fact that a typical post-combustion flue gas contains mostly N₂ and CO₂ (approximately 10–15% CO₂) at a total pressure of 1 bar and over a temperature range of 40-60 °C (Auta & Hameed, 2014; Mason, Sumida, Herm, Krishna, & Long, 2011). The experiment continued until the effluent concentration was equal to the feed concentration, i.e., $c_t/c_0=1$. Table 3.3 lists some of the physical properties of the adsorbent and characteristics of the adsorption bed along with the operating conditions used for the fixed-bed experiments. Given that the ammonia modification yielded material with very similar physical properties compared with the parent carbon (Shafeeyan, Wan Daud, Houshmand, & Arami-Niya, 2012), it is reasonable to assume that the difference in the amount of the presented characteristics of the modified and untreated adsorbents is negligible.

Parameter	Unit	Values
Particle apparent density (ρ_s)	kg m ⁻³	800
Average particle diameter (d_p)	m	0.65×10 ⁻³
Bed height (<i>L</i>)	m	0.2
Bed diameter (<i>d</i>)	m	0.01
Bed weight (<i>m</i>)	kg	5.5×10 ⁻³
Bed voidage (ε_b)	-	0.56
Feed flowrate (Q)	ml min ⁻¹	50 and 100
Feed composition (c_0)	vol %	15% CO2 in N2
Total pressure (P)	atm	1
Adsorption temperature (T)	°C	30, 45, and 60

Table 3.3: Physical properties of the adsorbent and characteristics of the adsorption bed along with the operating conditions used for the fixed-bed experiments



Figure 3.2: Schematic of the experimental system used for the column breakthrough measurements

3.4.4 Model description and solution methodology

3.4.4.1 Kinetic models

Among the key characteristics required for the design, simulation, and development of a CO_2 removal system, adsorption kinetic data are important because the residence time required for completion of an adsorption process, the adsorption bed size and, consequently, the unit capital costs are significantly influenced by the kinetic

considerations (Loganathan, Tikmani, Edubilli, Mishra, & Ghoshal, 2014; Monazam, Spenik, & Shadle, 2013). A wide variety of kinetic models with different degrees of complexity have been developed to quantitatively describe the adsorption processes and to identify the adsorption mechanism. Due to the complexities associated with the exact description of kinetic parameters, a common approach involves fitting the experimental results to a number of conventional kinetic models and selecting the model with the best fit (Loganathan et al., 2014; Serna-Guerrero & Sayari, 2010). Accordingly, three of the most common theoretical kinetic models (pseudo-first-order, pseudo-second-order, and Avrami models) that have been previously applied to describe the adsorbent–adsorbate interactions and adsorption rate behavior were employed in this study.

The pseudo-first-order kinetic model assumes reversible interactions with an equilibrium established between the gas and solid surfaces. The model states that the rate of change of a species is directly proportional to the difference between the saturation concentration of that species and its mean concentration within the particle. This model is represented by the following equation (Shokrollahi, Alizadeh, Malekhosseini, & Ranjbar, 2011):

$$\frac{dq_t}{dt} = k_F (q_e - q_t) \tag{3-13}$$

where k_F (s⁻¹) represents the pseudo-first-order kinetic rate constant, and q_e and q_t (mol kg⁻¹) denote the equilibrium uptake and the amount adsorbed at time t(s), respectively. Integration of Eq. (3-13) with the boundary conditions of $q_t = 0$ at t = 0 and $q_t = q_e$ at t = t_∞ gives the following equation:

$$q_t = q_e (1 - e^{-k_F t}) \tag{3-14}$$

The pseudo-second-order kinetic model is based on the assumption that the chemical interactions control the overall adsorption kinetics. The model assumes a linear relationship between the uptake rate and the square of the number of unoccupied

adsorption sites. This model is expressed according to the following equation (Borah, Sarma, & Mahiuddin, 2011):

$$\frac{dq_t}{dt} = k_s (q_e - q_t)^2 \tag{3-15}$$

where $k_s \pmod{\text{kg}^{-1}\text{s}^{-1}}$ is the pseudo-second-order kinetic rate constant.

Integration of Eq. (3-15) with the boundary conditions of $q_t = 0$ at t = 0 and $q_t = q_e$ at $t = t_{\infty}$ leads to the following equation:

$$q_{t} = \frac{q_{e}^{2}k_{s}t}{1 + q_{e}k_{s}t}$$
(3-16)

The Avrami model was originally developed to model phase transitions and crystal growth of materials and has recently been applied to the prediction of the adsorption kinetics of CO₂ on amine-functionalized adsorbents (Stevens et al., 2013; Wang, Stevens, Drage, & Wood, 2012). The adsorption rate of the Avrami equation is described as follows:

$$\frac{dq_t}{dt} = k_A^n t^{n-1} (q_e - q_t)$$
(3-17)

where *n* is the Avrami exponent and k_A (s⁻¹) is the Avrami kinetic constant.

The integrated form of Eq. (3-17) can be written as follows:

$$\frac{q_t}{q_e} = 1 - e^{-(k_A t)^n} \tag{3-18}$$

To establish a complete kinetic model, the equilibrium adsorption capacity, q_e , must be determined. Accordingly, in this study, the adsorption equilibrium was described using a semi-empirical Toth isotherm. This model was previously developed by our group to consider the simultaneous occurrence of two independent chemical and physical adsorption mechanisms for CO₂ adsorption. The model offers the advantage of differentiating the contributions of physical and chemical adsorption to the total CO₂ uptake, enabling its application for the description of the adsorption equilibria of CO₂ on

untreated and modified adsorbents. The proposed equilibrium model can be written as follows:

$$q_{e} = \left[\frac{q_{m}K_{T}P}{\left(1 + \left(K_{T}P\right)^{m_{T}}\right)^{1/m_{T}}}\right]_{phys} + \left[\frac{q_{m}K_{T}P}{\left(1 + \left(K_{T}P\right)^{m_{T}}\right)^{1/m_{T}}}\right]_{chem}$$
(3-19)

To express the temperature dependence of the Toth isotherm parameters, the parameters K_T , q_m , and m_T were described by the following equations (Li & Tezel, 2008):

$$K_{T} = K_{T_{0}} \exp\left[-\frac{\Delta H}{RT_{0}}\left(\frac{T_{0}-T}{T}\right)\right]$$
(3-20)

$$q_m = q_{m_0} \exp\left[\eta\left(\frac{T_0 - T}{T_0}\right)\right]$$
(3-21)

$$m_T = m_{T_0} + \alpha \left(\frac{T - T_0}{T}\right) \tag{3-22}$$

where *T* is the absolute temperature (K), T_0 is the reference temperature, K_{T0} and m_{T0} are the affinity and heterogeneity parameters at the reference temperature, respectively, α and η are constant parameters, (- ΔH) is the isosteric heat of adsorption at zero coverage (kJ mol⁻¹), and *R* is the gas-law constant (J mol⁻¹ K⁻¹).

The optimal values of the isotherm parameters are summarized in Table 3.4, using 30 °C as a reference temperature.

	$q_m^{}_{(m mol~kg^{-1})}$	K_T (atm ⁻¹)	<i>m_T</i> (dimensionless)	α (dimensionless)	η (dimensionless)	$\left(-\Delta H\right)$ (kJ mol ⁻¹)
Physical adsorption	5.01	7.1×10 ⁻¹	0.59	0.96	13.43	23.05
Chemical adsorption	0.54	1.05×10 ⁵	0.29	0. 25	1.67	68.11

Table 3.4: Optimal values of the proposed Toth equilibrium isotherm parameters

To determine each set of kinetic model parameters, the experimental data were then fitted to the previously mentioned kinetic models. The transformation of non-linear kinetic equations into linear forms is generally associated with the error distribution, depending on the method used to linearize the kinetic equations (Kumar & Sivanesan, 2006). Therefore, to rigorously estimate the kinetic parameter sets from the original form of each kinetic model, a non-linear regression analysis using the Marquardt-Levenberg algorithm implemented in SigmaPlot software version 12.0 (Systat Software Inc., USA) was employed in this study. To quantitatively compare the goodness of fit of the kinetic models with the experimental results, two different error functions, the nonlinear regression coefficient (R^2) and the normalized standard deviation (Δq), were evaluated. The regression coefficient, which determines how well the data points fit the model, was calculated as follows (Stevens et al., 2013):

$$R^{2} = 1 - \left(\frac{\sum_{i=1}^{n} (q_{t(mes)} - q_{t(mod)})^{2}}{\sum_{i=1}^{n} (q_{t(mes)} - \overline{q_{t(mes)}})^{2}}\right) \cdot (\frac{n-1}{n-p})$$
(3-23)

The normalized standard deviation, which reflects the deviation between the experimental results and the values predicted by the kinetic models, can be represented using the following equation (Vargas, Cazetta, Kunita, Silva, & Almeida, 2011):

$$\Delta q(\%) = 100 \times \sqrt{\frac{\sum \left[(q_{t(mes)} - q_{t(mod)}) / q_{t(mes)} \right]^2}{n - 1}}$$
(3-24)

where *n* is the number of experimental adsorption points of the kinetic curves, the subscripts "*mod*" and "*mes*" refer to the model predicted and measured values of the amount adsorbed, respectively, $\overline{q_{mes}}$ is the average of the experimental data, and *p* is the number of parameters of the model.

3.4.4.2 Modeling dynamic column breakthrough experiments

A basic study on the dynamics behavior of an adsorption column system is required to obtain a deeper understanding of the behavior of new adsorbents during the adsorption/desorption cycles and for process design and optimization purposes (Shafeeyan, Wan Daud, & Shamiri, 2014). To predict the breakthrough behavior of CO₂ adsorption in a fixed-bed packed with GAC and OXA-GAC adsorbents, the following model was proposed based on the mass balance concept combined with an Avrami model for the representation of the adsorption kinetics and a semi-empirical Toth model for the description of the adsorption equilibrium. To develop the fixed-bed model, we assumed the following:

(1) The gas phase behaves as an ideal gas. (2) An axially dispersed plug-flow model is adopted. (3) The radial gradient of concentration is negligible (given that the ratio of the particle to the column radius is less than 10) (Monazam et al., 2013). (4) The fixed bed is assumed to operate isothermally. (5) The feed contains a low concentration of CO₂; thus, the pressure drop throughout the column is negligible, and the linear velocity remains constant along the bed (Dantas et al., 2009). (6) The effect of nitrogen adsorption is negligible. (7) The effect of external mass transfer (i.e., macropore diffusion and film mass transfer) is considered negligible. (8) The rate of adsorption is described using an Avrami kinetic rate expression, which is explained in Section 4.4.1. (9) The adsorption equilibrium is expressed by a semi-empirical Toth model, as will be described in Section 3.1. Most of these assumptions are commonly accepted for a modeling PSA operation (Ruthven, 1984) and are clearly a compromise between accuracy and effort of parameter determination and model solution (Bonnot, Tondeur, & Luo, 2006). The correlations used for the estimation of the model parameters are summarized in Table 3.5.

Based on these assumptions, the transient gas-phase mass balance for a differential control volume of the adsorption column can be described by the following equation (Ruthven, 1984):

$$-D_{L}\frac{\partial^{2}c}{\partial z^{2}} + u\frac{\partial c}{\partial z} + \frac{\partial c}{\partial t} + (\frac{1-\varepsilon_{b}}{\varepsilon_{b}})\rho_{p}\frac{\partial q}{\partial t} = 0$$
(3-25)

where D_L is the effective axial dispersion coefficient, z represents the length in the axial direction, c denotes the CO₂ concentration in the gas phase, ε_b is the bed voidage, u is the superficial velocity, t represents time, ρ_p is the particle density, q represents the volume-averaged concentration in the adsorbed phase that may constitute a connection between the solid and gas phase mass balances, and the term $\partial q / \partial t$ expresses the adsorption rate, which is determined from an Avrami kinetic model, as described in Section 4.4.1. In Eq. (3-25), the first term is the axial dispersion term, the second term is the convective flow term, and the third and fourth terms represent the accumulation in the fluid and solid phases, respectively. The equation is applied to describe the gas composition distribution in the bed. Ignoring the radial dependence of concentration in the gas and adsorbed phases, c and q are expressed as functions of z and t, respectively. In addition, applying the ideal gas law, the concentration c can be correlated with the partial pressure as c=yP/RT, where P is the total pressure, y denotes the CO₂ mole fraction in the feed gas, R is the universal gas constant, and T expresses the feed gas temperature.

The following initial and boundary conditions are assumed (Khalighi et al., 2012):

I.C.:
$$c(z,t=0)=0; q(z,t=0)=0$$
 (3-26)

B.C.:
$$D_L \left. \frac{\partial c}{\partial z} \right|_{z=0} = -u(c_0 - c)$$
 (3-27)

$$\left. \frac{\partial c}{\partial z} \right|_{z=L} = 0 \tag{3-28}$$

Molecular	
wolecular	$D_{1.5} = \frac{T^{1.5}}{(1-1)^{1.5}} \left(\frac{1}{(1-1)^{1.5}} + \frac{1}{(1-1)^{1.5}}\right)^{0.5} \left(0.0027 - 0.0005(\frac{1}{(1-1)^{1.5}} + \frac{1}{(1-1)^{1.5}})^{0.5}\right)$
diffusivity (D_m)	$P\sigma_{i,j}^2 \Omega_D M_i M_j \left(OOOD M_i M_j \right)$
(Bird et al.,	where, 1 06036 0 19300 1 03587 1 76474
2002; Perry,	$\Omega_D = \frac{1.00050}{T^{*0.15610}} + \frac{0.17500}{\exp(0.47635T^*)} + \frac{1.05507}{\exp(1.52996T^*)} + \frac{1.70474}{\exp(3.89411T^*)};$
Green, &	$\sigma_{i,j} = (\sigma_i + \sigma_j) / 2; T^* = kT / \varepsilon_{i,j};$
,	$\varepsilon_{i,j} = (\varepsilon_i \varepsilon_j)^{0.5}; \varepsilon / k = 0.77T_c; \sigma = 2.44(T_c / P_c)^{1/3}$
Maloney, 1997)	
Axial dispersion	$\varepsilon_{L} = 20 \pm 0.5ScRe; Sc = \frac{\mu}{Re} \cdot Re = \frac{\rho_{g} u \varepsilon_{b} d_{p}}{Re}$
coefficient (D_L)	$\rho_{b} D_{m} \rho_{s} D_{m} \mu$
(Ruthven, 1984)	where D_m (m ² s ⁻¹) is the molecular diffusivity and <i>Sc</i> and <i>Re</i> are the Schmidt
	and Reynolds numbers, respectively.
Pure gas	$u = 2.6693 \times 10^{-5} \sqrt{MT}$
viscosity (11)	$\mu = 2.0055 \times 10^{-10} \sigma^2 \Omega_{\mu}$
viscosity (µ)	where μ is in g cm ⁻¹ s ⁻¹ and
(Bird et al.,	$\Omega_{\mu} = \frac{1.16145}{T^{*0.14874}} + \frac{0.52487}{cmr(0.77220T^{*})} + \frac{2.16178}{cmr(2.42787T^{*})}$
2002)	$I = \exp(0.7/320I) = \exp(2.43787I)$
Viscosity of gas	∠ >-0.5 ⊏ ⊐2
viscosity of gas	$\mu = \sum_{i=1}^{n} \frac{y_{i}\mu_{i}}{1 + (\mu_{i})^{0.5}} = \frac{1}{(1 + M_{i})^{0.5}} \left[1 + (\mu_{i})^{0.5} (M_{j})^{0.25} \right]^{-1}$
mixture (μ_{mix})	$\mu_{mix} = \sum_{i=1}^{j} \sum_{i=1}^{j} y_i \Phi_{i,j}, \Psi_{i,j} = \sqrt{8} \left(\prod_{i=1}^{j} M_j \right) \left[\prod_{i=1}^{j} \mu_j, M_i \right]$
(Bird et al.,	
2002)	
Density of gas	P T 1
	$\rho_g = \sum y_i M_i \times \frac{1}{P} \times \frac{1}{T} \times \frac{1}{V}$
mixture (pg)	$i=1$ I_0 I v_m
	where ρ_g (kg m ⁻³) is the density; T_0 and P_0 are the temperature and pressure
	at STP conditions; and V_m is 22.4 L mol ⁻¹ .

Table 3.5: Correlations used for estimation of the model parameters

 σ (A°) and ε/k (K) are the Lennard- Jones length and energy parameters; T (K) is the temperature; P (atm) is the total pressure; M_i (g mol⁻¹) is the component molecular weight; T_c (K) and P_c (atm) are the critical temperature and pressure; y_i is the component mole fraction.

3.4.4.3 Solution methodology

The simultaneous solution of a set of coupled nonlinear partial differential equations constructed from mass conservation along with ordinary differential and algebraic equations representing kinetic and equilibrium equations yields the predicted adsorption breakthrough curve. The resultant system of differential and algebraic equations (Eqs. 3-25, 3-17, and 3-19) which are coupled with one another, together with the corresponding initial and boundary conditions (Eqs. 3-26 to 3-28) detailed in section 3.5.2 are solved numerically using a chemical reaction engineering module implemented in COMSOL Multiphysics version 4.4 (Burlington, MA, USA), which uses the finite element method for the numerical solution of differential equations.

CHAPTER 4: RESULTS AND DISCUSSION

4.1 Introduction

Growing environmental concern regarding the adverse effects of anthropogenic carbon dioxide (CO_2) emissions on global warming and climate change have motivated research into the capture of CO₂ from large emission point sources and the mitigation of the unfettered release of this major greenhouse gas into the atmosphere (Delgado et al., 2006; Samanta, Zhao, Shimizu, Sarkar, & Gupta, 2011). Fossil fuel-fired power plants account for approximately one-third of global CO₂ emissions due to the combustion of fossil fuels such as petroleum, coal, and natural gas (Dantas et al., 2011; Monazam et al., 2013). Therefore, remarkable growth in research activity related to the development, improvement, and optimization of efficient separation techniques for removing CO₂ from power-plant flue gas has occurred over the past three decades (Agarwal et al., 2010a). Currently, amine-based chemical absorption processes are the most mature methods for removing CO₂ from industrial gas streams (Delgado et al., 2006; Serna-Guerrero, Belmabkhout, & Sayari, 2010b). Unfortunately, the absorption process is highly energyintensive because of the high energy requirements for regeneration. A low CO₂ loading capacity, extensive corrosion of the process equipment and toxicity are additional disadvantages associated with the amine absorption process (Gray et al., 2005; Rinker et al., 2000). A promising alternative procedure for separating CO_2 from flue gases is adsorption using solid physical adsorbents, which could reduce the costs associated with the capture step (Ho et al., 2008; Radosz et al., 2008). Other advantages of these solid physical adsorption processes over traditional aqueous amine processes include greater capacity, higher selectivity, smaller regeneration energy requirements, and easier handling (Garcés et al., 2013; Samanta et al., 2011). Many literature reports have investigated CO₂ capture from flue gas streams using pressure swing adsorption (PSA)
(Ho et al., 2008; Ko et al., 2003; Reynolds et al., 2005; Xiao et al., 2008). However, the success of a PSA process depends on the selection of an adsorbent with sufficient adsorption capacity, selectivity, and durability (Su et al., 2010). Accordingly, in recent years, various adsorbents, including zeolites, carbon molecular sieves, metal organic frameworks, silicas, and activated carbon, have been extensively investigated for their application in removing CO_2 from flue-gas streams (Loganathan et al., 2014; Sevilla & Fuertes, 2011).

Activated carbon has the advantages of low energy requirements for regeneration, an inherent affinity for CO₂, stability at higher temperatures, low cost, a large surface area and high micropore volume, hydrophobicity, and the possibility of tailoring its surface chemistry and structural properties (Bezerra et al., 2011; Garcés et al., 2013; Sjostrom & Krutka, 2010). Notably, however, most of the conventional adsorbents based on physisorption suffer from the disadvantage of poor adsorption capacities at relatively low CO₂ concentrations and high adsorption temperatures (i.e., the operating conditions characteristic of the separation of CO₂ from a typical flue gas mixture), which may represent a serious drawback for flue gas CO₂ removal using gas adsorption (Bezerra et al., 2011; Samanta et al., 2011). To improve CO₂ adsorption capacity, modifications of adsorbent surface chemical properties by the incorporation of nitrogen surface groups have gained much attention in the past few years (Drage et al., 2007; Knowles et al., 2006; Maroto-Valer et al., 2005; Plaza et al., 2007). CO₂ is a weak Lewis acid that can interact with electron donors; therefore, the presence of basic nitrogen-containing surface groups is thought to provide adsorbents capable of interacting strongly with CO₂ (Plaza et al., 2011; Zhijuan Zhang, Xu, et al., 2010). Przepiórski et al. demonstrated that the CO₂ adsorption capacity on an ammonia-treated CWZ-35 activated carbon was 76 mg g⁻¹ at 309.15 K and 102.5 kPa (Przepiórski et al., 2004). Hao et al. observed a maximum CO₂ uptake of 3.13 mmol g⁻¹ at 25 °C and 1 atm for synthesized nitrogen-doped porous carbon

monolith samples (Hao, Li, Qian, & Lu, 2010), and Wahby et al. reported a CO₂ uptake of 4.7 mmol g⁻¹ by chemically activated petroleum pitch at 25 °C and 1 atm (Wahby et al., 2010). In this study, a commercial granular activated carbon (GAC) adsorbent was modified through an oxidation-amination process in an effort to increase its surface basicity and consequently enhance its CO₂ adsorption capacity. The promising characteristics of the optimal adsorbent (OXA-GAC) in terms of adsorption capacity (exhibiting an increase of 44% in capacity compared with the capacity of the original AC at 1 atm and 105 °C) and multicycle durability make it suitable for practical applications. For the design and simulation of separation processes, the study of the adsorption equilibria is essential in supplying the basic information required for developing and validating models that represent the nature of the adsorption processes (Cavenati et al., 2004; Grande & Rodrigues, 2004). Although reliable experimental equilibrium data for some adsorbates on carbon adsorbents are extensively available in the literature (An, Feng, & Su, 2009; Garcés et al., 2013), as far as we know, very few data have been reported for CO₂ adsorption over ammonia-modified activated carbons under the conditions used here. Furthermore, no studies have been conducted to analyze the CO₂ adsorption isotherms over such adsorbents to fit to a proper isotherm model. Therefore, in the present study, the adsorption equilibria of carbon dioxide over the parent GAC and its ammonia-modified counterpart (OXA-GAC) at various operating conditions were obtained and used to fit to an appropriate semi-empirical isotherm model.

Despite the notable increase in the practical applications of adsorption for separation and purification, due to the inherent complexity associated with such systems, the industrial design and optimization of adsorptive gas-separation processes has largely remained an experimental effort (Bastos-Neto et al., 2011). The main disadvantages of the empirical design of an adsorption column are the high costs and operational effort required (Siahpoosh et al., 2009). Therefore, to develop, evaluate, and optimize an appropriate

adsorption process, several researchers have proposed the application of a mathematical model capable of describing the combined effect of equilibrium and kinetics on a fixed bed with the selected adsorbent (Agarwal et al., 2010a; Dantas et al., 2011; Shivaji Sircar, 2006). Shafeeyan et al. provided a comprehensive review of previous studies on the mathematical modeling of packed columns to capture CO₂ from various flue-gas streams for a variety of adsorption systems (Shafeeyan et al., 2014). A predictive model using independently established equilibrium and kinetic parameters may, in principle, provide a method of estimating the column dynamic capacity without extensive experimentation. Moreover, such models are capable of estimating the concentration profile for a certain constituent in the bulk gas at all locations within the packed column (Shafeeyan et al., 2014). Accordingly, in this study, the experimental kinetics data of CO₂ adsorption at different temperatures were fitted to three different kinetic models (i.e., pseudo-first order, pseudo-second order, and Avrami models), and the corresponding kinetic rate constants were estimated from the model providing the best fit. To the best of our knowledge, no studies have been conducted to develop a kinetic model for CO₂ capture over ammonia-modified activated carbon. Furthermore, using the best fitted kinetic model and an appropriate equilibrium equation, a simple dynamic model was developed. The proposed model was constructed based on the mass balance concept and used to describe the breakthrough behavior of the CO₂ adsorption in a fixed-bed reactor packed with ammonia-modified and untreated adsorbents. Finally, the model predictions were verified by comparing the simulations with the experimental results.

In this chapter the results are presented in three parts. Part 1 investigates the application of response surface methodology in predicting and optimizing the amination conditions of activated carbon adsorbent toward CO_2 adsorption. Part 2 studies the adsorption equilibrium of carbon dioxide onto the GAC and OXA-GAC adsorbents and develops a

semi-empirical equilibrium model able to distinguish the contributions of physical and chemical adsorption to the total CO₂ uptake. Part 3 studies the kinetics of CO₂ adsorption on the GAC and OXA-GAC adsorbents. To predict the breakthrough behavior of the fixed-bed adsorption of CO₂, the focus of Part 3 is to develop a dynamic model that consists of an Avrami equation to describe the kinetics of adsorption and a semi-empirical Toth equation to represent the gas–solid equilibrium isotherm. In addition, a comprehensive discussions and explanations on experimental results are presented.

4.2 The application of response surface methodology to optimize the amination of activated carbon for the preparation of carbon dioxide adsorbents

4.2.1 Evaluation of CO₂ adsorption and desorption capacity

The experimental points (in coded and actual values) together with the values of both responses are represented in Table 3.2. As can be seen from the table, adsorbents O9 and O5 displayed the highest and lowest CO₂ adsorption capacities of 30.04 and 20.17 mg/g at 105 °C, respectively. On the contrary, sorbent O5 exhibited the highest and sorbent O9 showed the lowest CO₂ desorption capacities of 99.4 %, and 91.9 %, respectively. Inspection of the data presented in Table 3.2 reveals that the sorbent preparation variables have a significant effect on the CO₂ adsorption/desorption capacity. The relationship between the responses and independent variables was analyzed using response surface methodology (RSM). When multiple regression analysis is applied to the design matrix and the obtained response values, a fitted quadratic polynomial equation is generated. This equation can be used to optimize the amination conditions. For both responses, the quadratic model was selected as suggested by the software. The models were attained based on the highest order polynomials where the additional terms were significant and the models were not aliased. The general response models in terms of coded factors

include all numerical and categorical parameters (i.e., x_1, x_2 and x_3). The predictive models in coded terms (i.e., Y_S and Y_D) are presented in Tables 4.1 and 4.2, respectively. On the contrary, the final equations in actual form are defined for each type of categorical factor separately. The final empirical equation obtained for CO₂ sorption capacity (Y_S) of the adsorbents prepared from HTA and OXA (symbolized by Y_{S-1} and Y_{S+1} , respectively) are given in Eqs. (4-1) and (4-2) respectively. For CO₂ desorption capacity (Y_D) of the adsorbents prepared from HTA and OXA (symbolized by Y_{D-1} and Y_{D+1} , respectively), the final equations are given in Eqs. (4-3) and (4-4), respectively.

$$Y_{S-1} = +12.48480 + 0.016138X_1 + 6.89030X_2 - 7.7356 \times 10^{-6}X_1^2 - 1.48770X_2^2$$
$$-7.34375 \times 10^{-4}X_1X_2$$
(4 - 1)

 $Y_{D-1} = +105.39959 - 0.016811X_1 - 4.21352X_2 + 8.50528 \times 10^{-6}X_1^2 + 0.86084X_2^2$ +4.6875×10⁻⁴X_1X_2 (4 - 3) $Y_{D+1} = +105.87736 - 0.018477X_1 - 4.20241X_2 + 8.50528 \times 10^{-6}X_1^2 + 0.86084X_2^2$ +4.6875×10⁻⁴X_1X_2 (4 - 4)

Source Sum of Degree of F-value *p*-value (Prob>*F*) Mean squares freedom square Model 141.51 8 17.69 85.09 < 0.0001, Significant X_1 78.68 78.68 378.46 < 0.0001 1 *X*₂ 2.94 2.94 14.15 0.0016 1 *X*₃ 16.15 16.15 77.67 < 0.0001 X_{1}^{2} 0.75 0.75 3.59 0.0751 $\begin{array}{c} X_2^{\ 2} \\ X_1 X_2 \end{array}$ 1.08 1.08 5.19 0.0359 0.69 0.0861 0.69 3.32 1 X_1X_3 5.79 5.79 27.86 < 0.0001 1 X_2X_3 0.048 0.048 0.23 0.6381 1 Residual 3.53 17 0.21 Final equation in terms of coded factors after excluding the insignificant terms: $Y_{\rm S} = +27.12 + 2.96x_1 + 0.57x_2 + 0.79x_3 - 1.24x_1^2 - 1.49x_2^2 - 0.29x_1x_2 + 0.8x_1x_3 + 0.073x_2x_3$ R² = 0.9756; Adjusted-R² = 0.9642; Predicted- R² = 0.8229; Adequate precision= 33.522

Table 4.1: Analysis of variance (ANOVA) for the CO2 adsorption capacity

Checking the adequacy of the model is one of the most important parts of the data analysis procedure, as the model functions would give poor or misleading results if the fit were inadequate (Körbahti, 2007; Körbahti & Rauf, 2008a; Myer & Montogomery, 2002; Yetilmezsoy et al., 2009). An analysis of variance (ANOVA) was carried out for each response to assess the significance and fitness of the model. The results are presented in Tables 4.1 and 4.2.

	•					•
Source	Sum of	Degree of	Mean	F-value	<i>p</i> -value (Pr	ob>F)
	squares	freedom	square			
Model	90.61	8	11.33	171.31	<0.0001,	Significant
<i>X</i> ₁	60.84	1	60.84	920.26	< 0.0001	
X_2	2.1	1	2.10	31.80	< 0.0001	
<i>X</i> ₃	1.63	1	1.63	24.58	0.0001	
X_{1}^{2}	0.9	1	0.90	13.66	0.0018	
X_{2}^{2}	0.36	1	0.36	5.47	0.0318	
$\bar{X_1X_2}$	0.28	1	0.28	4.25	0.0548	
$X_1 X_3$	1.00	1	1.00	15.13	0.0012	
$X_2 X_3$	0.0002	1	0.0002	0.0042	0.9491	
Residual	1.12	17	0.066	-	-	

Table 4.2: Analysis of variance (ANOVA) for the CO₂ desorption capacity

Final equation in terms of coded factors after excluding the insignificant terms: $Y_5 = +93.7 - 2.6x_1 - 0.48x_2 - 0.25x_3 + 1.36x_1^2 + 0.86x_2^2 + 0.19x_1x_2 - 0.33x_1x_3 + 5.556 \times 10^{-3}x_2x_3$ $R^2 = 0.9877$; Adjusted- $R^2 = 0.9820$; Predicted- $R^2 = 0.9342$; Adequate precision= 45.921

ANOVA is a statistical technique that subdivides the total variation in a set of data into component parts associated with specific sources of variation; this allows one to test hypotheses based on the parameters of the model (Gönen & Aksu, 2008). Table 4.3 and 4.4 illustrate the reduced quadratic models in terms of coded factors and also show other statistical parameters for both of the studied responses. Based on the ANOVA results presented in these tables, it can be concluded that the models were significant for both responses with p-values less than 0.0001(model and term p-value <0.05 indicates the model and the term are significant for 95% confidence intervals) to predict the response values (Azargohar & Dalai, 2005). In this case, amination temperature (X_1), amination time (X_2), type of starting material (X_3), X_1X_3 , and X_2^2 were significant model terms for both the CO₂ adsorption and desorption capacities with p-values less than 0.05. In

addition, the interaction between the X_2 and X_3 variables (X_2X_3) was insignificant to the responses (term p-value >0.100 indicates the model terms are not significant) which could be manually removed from the model to improve the regression model and optimization results.

The fit of model to the empirical data was tested by calculating the regression coefficient R^2 and R^2_{adj} . The R^2_{adj} values of 0.9642 and 0.9820 were obtained for the CO₂ adsorption (*Y*_b) capacity, respectively. This indicated that 96.42 and 98.20% of the total variation in the CO₂ adsorption and desorption capacity, respectively, could be explained by the quadratic model. The high R^2 value (i.e. close to unity) and reasonable agreement between the predicted and adjusted R^2 observed for both responses implies that the developed model can satisfactorily account for the experimental data (Azargohar & Dalai, 2005; Can et al., 2006; Garg, Kaur, Garg, & Sud, 2008; Yetilmezsoy et al., 2009). Adequate precision measures the signal-to-noise ratio and compares the range of the predicted values at the design points to the average prediction error. The adequate precision ratios of 33.5 and 45.9, derived for the CO₂ adsorption and desorption capacity, respectively, are much greater than 4 and indicate adequate model discrimination (Ölmez, 2009). It is worth noting that despite the lack-of-fit was significant, due to the reasonable agreement between the predicted and adjusted R^2 , the model suggested here can be used to navigate into design space to find an optimum.

It is known that the CO_2 capture capacity of activated carbons is a combination of both physical and chemical adsorption (is governed by porous structure and is influenced by the surface chemistry) (Arenillas, Drage, Smith, & Snape, 2005; Drage et al., 2007). Our previous study showed that ammonia modification did not significantly alter the pore structure of the studied adsorbents (less than 15%) (Shafeeyan et al., 2011). Therefore, the behavior of these materials in carbon dioxide adsorption may be primarily attributed to the change in their chemical surface properties. The best way of expressing the effect of amination parameters on the CO₂ adsorption/desorption within the experimental range was to generate response surface plots of the equation. The linear, quadratic and cross terms in the second-order polynomial model were used to generate a three-dimensional (3D) response surface graph and a two-dimensional contour plot. Based on the ANOVA results presented in Tables 4.3 and 4.4, amination time (X_2), type of starting material (X_3), and X_2^2 were found to have significant effects, but amination temperature (X_1) has the largest effect on both the CO₂ adsorption and desorption capacities (due to the highest *F*values). Furthermore, among the interaction terms, X_1X_3 was found to significantly affect the studied responses whereas the effect of interaction between X_2 and X_3 (X_2X_3) was no statistically significant.

4.2.1.1 Effect of amination variables on the CO2 adsorption capacity

Three-dimensional (3D) response surface plots of the predictive quadratic model for the CO₂ adsorption capacity of pre-heat treated and pre-oxidized carbons are presented in Fig. 4.1 (a) and (b), respectively. The response surfaces were generated based on Eqs. (4-1) and (4-2). For both type of adsorbents, CO₂ capture capacity increases as the temperature and time of the ammonia treatment increases. In addition, the positive sign in front of the (X_1) and (X_2) terms in Eqs. (4-1) and (4-2) also implies a favorable or synergistic effect of these terms on capture capacity. It has been reported that the CO₂ adsorption performance of modified adsorbents is strongly influenced by the types of nitrogen functionalities introduced on the carbon surface (Drage et al., 2007; Maroto-Valer et al., 2005; Pevida, Plaza, et al., 2008; Plaza, Rubiera, et al., 2010; Shafeeyan et al., 2011). However, the nature of the incorporated nitrogen functionalities differed with the temperature of ammonia treatment (Mangun, Benak, Economy, & Foster, 2001). Therefore, the synergistic effect of X_1 and X_2 factors on the Y_8 response may be related

to the formation of more N-containing functionalities particularly in the form of pyrrole and pyridinic-like functionalities which are appeared during high temperature (>600 °C) amoniation (suitable surface chemistry) (Jansen & van Bekkum, 1994; Maroto-Valer et al., 2005; Pevida, Plaza, et al., 2008; M. G. Plaza et al., 2009; Plaza, Rubiera, et al., 2010). In addition, a partial gasification of the adsorbents with free radicals that were generated during the decomposition of ammonia would lead to pore enlargement, which may contribute to the enhancement of CO₂ uptake (favorable textural properties) (Boehm, Mair, Stoehr, De Rincón, & Tereczki, 1984; Bota & Abotsi, 1994; Pevida, Plaza, et al., 2008; Plaza, Pevida, et al., 2010). This is in accordance with our previous findings, in which the surface chemistry and porous texture of the modified sorbent became more favorable as the amination temperature increased (Shafeeyan et al., 2011).

As can be seen in Table 3.2 and Fig. 4.1, in comparison with the pre heat- treated aminated samples, the pre-oxidized carbon displayed an enhanced ability to adsorb CO₂. The greatest experimental CO₂ adsorption capacity (Table 3.2), 30.04 mg/g, was obtained for the pre-oxidized sample that had been modified with ammonia at 800 °C for 2 h (O9). The increased CO₂ capture capacity observed for pre-oxidized adsorbents indicates that the presence of oxygen functionalities on the carbon surface prior to ammonia treatment plays an important role in improving the CO₂ capture performance of the modified samples. When the oxidized carbon is treated with ammonia at elevated temperatures, free radicals such as NH₂, NH, and atomic hydrogen, which are created during ammonia decomposition, may attack the surface oxides and active sites present on the carbon surface and form N-containing functional groups (Bota & Abotsi, 1994; Jansen & van Bekkum, 1995; Stöhr, Boehm, & Schlögl, 1991; Vinke, van der Eijk, Verbree, Voskamp, & van Bekkum, 1994). As a consequence of the improved surface chemistry, stronger interactions between the acidic gaseous CO₂ and the basic nitrogen groups would take place and lead to a higher CO₂ capture.

2004) modified granular activated carbons (GACs) with dry ammonia and demonstrated that the presence of a threshold number of oxygen-containing acidic functionalities on the carbon surface is a prerequisite for effective ammonia treatment. Mangun et al. (2001) reported that the reactivity of ammonia gas with carbon surfaces and the consequent formation of nitrogen surface groups increase with the oxygen content of the carbon precursor. The authors suggested that using a post-oxidation ammonia treatment may enlarge the carbon pores while increasing the surface basicity (number of N-containing functionalities). These changes in the carbon pore structure and surface chemistry are desirable because they enhance CO_2 uptake.

In the case of the amination temperature factor (X_1), the trend of enhanced CO₂ uptake is limited to mid-range temperatures (near 800 °C). As can be observed from Fig. 4.1, when the amination temperature goes beyond a certain limit (over 800 °C), there is a decrease in the amount of CO₂ uptake for both type of adsorbent. The observed decrease in CO₂ uptake may be, in part, a consequence of the thermal decomposition of nitrogen surface groups during the high temperature treatment, which would decrease chemisorption (Drage et al., 2007). However, decreased CO₂ uptake can also be associated with the collapse of adjacent pore walls during the pre-heating stage, which leads to diminished textural properties (i.e., a decrease in physisorption contributions) (Maroto-Valer et al., 2005). Furthermore, lowered CO₂ capture capacities of activated carbons treated with ammonia at higher temperatures have been reported for commercial activated carbons. This is likely a result of micropores being closed and blocked by N-containing groups (Przepiórski et al., 2004). The clear maximum values observed in the response plots (Fig. 4.1) imply that the optimum conditions for maximum capture capacity are highly dependent on the amination time and temperature in the design space.



Figure 4.1: Response surface plot of CO₂ adsorption capacity for (a) pre-heat treated, and (b) pre-oxidized adsorbent.

4.2.1.2 Effect of amination variables on the CO₂ desorption capacity

All of the studied samples were regenerated by changing the reactive gas, CO_2 , to a nonreactive gas, N_2 , at 105 °C. Fig. 4.2 a and b represents the 3D response surfaces that were constructed to show the variables' effectiveness on the CO_2 desorption capacity of the modified adsorbents. The response surfaces were generated based on Eqs. (4-3) and (4-

99

4). A decrease in desorption capacity with an increase in amination temperature can be observed for both types of adsorbents. Furthermore, the negative sign in front of the (X_1) and (X_2) terms in Eqs. (4-3) and (4-4) also indicates their unfavorable or antagonistic effect on the desorption performance. One possible reason for this observation is that for high temperature adsorption (e.g., over 100 °C), CO₂ uptake is mainly controlled by chemical adsorption (Plaza, Rubiera, et al., 2010; Shafeeyan et al., 2011). In samples with favorable surface chemistries, the stronger interactions between acidic, gaseous CO₂ and the incorporated basic N-functionalities lead to the formation of stronger chemical bonds that hinder regeneration. Accordingly, the sample with the highest CO₂ adsorption capacity (O9), demonstrated the lowest desorption capacity (91.9%). In contrast, as can be seen in Table 3-2 and Fig. 4.2, the sample with the lowest capture capacity (O5) had the highest desorption capacity (99.4%), which may be the result of either weak chemical interactions between the CO₂ and N-functionalities or its poor textural properties compared to the other adsorbents. The aforementioned conclusion was based on the chemical and textural characterization results reported from our group (Shafeeyan et al., 2011).

It should be noted that higher temperature of amination (over 800 °C) did not result in additional decreases in the regeneration potential of either type of adsorbent. The results presented in Table 3.2 demonstrates that the samples treated with ammonia at temperatures above 800 °C showed a slight increase in the CO₂ desorption capacity. As has been previously discussed, this may be attributed to either the elimination of basic N functionalities suitable for CO₂ adsorption or poorly developed textural properties that appear during high temperature treatments. Consequently, the adsorbed gas on the surface of the activated carbon becomes unstable, which results in more desorption of adsorbed CO₂ molecules.



Figure 4.2: Response surface plot of CO₂ desorption capacity for (a) pre-heat treated, and (b) pre-oxidized adsorbent.

4.2.2 Investigation of the optimum amination conditions

In the production of efficient CO₂ adsorbents, relatively high capture capacities are necessary for production to be economically feasible. For practical applications, the modified adsorbents should not only possess high adsorption capacities but also show high desorption capacities for economical viability. However, the CO₂ adsorption (Y_S) and desorption (Y_D) capacities are compensatory (i.e., when Y_S increases, Y_D decreases and vice versa). Thus, it is difficult to find a balance between high adsorption and desorption capacities. The determination of the optimum amination conditions for maximizing the amounts of Y_S and Y_D within the studied experimental range was carried out using the point prediction function in the Design-Expert software. In addition, the optimization was also used to determine whether the use of HTA or OXA is more favorable to the production of adsorbents with higher CO₂ adsorption/desorption capacities.

Multiple response optimizations could be generated from the contour plots of the second order model for each response in the experimental ranges of each factor (Fig. 4.3a and b). Furthermore, the optimum conditions could be graphically visualized by superimposing the contours of the response surfaces in an overlay plot (Fig. 4.4). The shaded area in the overlay plot of the responses (Fig. 4.4) corresponds to the optimum variable ranges for the production of an adsorbent with a high CO₂ adsorption (\geq 26 mg/g) and desorption capacity (\geq 94%). To identify the optimum conditions, the targeted criteria were set as maximum values for the responses (CO₂ adsorption and desorption capacity), whereas the values of the numerical variables (amination temperature and amination time) were set in the ranges studied for both HTA and OXA adsorbents. The amination conditions that showed the highest desirability were experimentally verified (desirability is an objective function that ranges from zero outside of the limits to one at the goal; the program seeks to maximize this function (Amini et al., 2008). The optimal point indicated by numerical optimization corresponds to a pre-oxidized adsorbent that has been aminated at 425 °C for 2.1 h, as indicated in Fig. 4.3. Fig. 4.3a and b shows the contour plots for predicting the maximum CO_2 adsorption and desorption capacities, respectively, under the determined optimum conditions for pre-oxidized adsorbents. It was predicted that maximum CO_2 adsorption and desorption capacities of 26.1 mg/g and 94.9%, respectively, could be obtained for the activated carbon adsorbent. From the aforementioned optimization results, the application of post-oxidation ammonia treatment (preparation method) at 425 °C for 2.12 h (experimental conditions) was found to produce a high adsorption/ desorption capacity activated carbon for carbon dioxide adsorption.

To evaluate the accuracy of the second order equations, three additional CO_2 adsorption/desorption experiments were carried out using adsorbents that were prepared under the RSM-predicted optimum conditions. The average values of the repeated experiments and predicted results are presented in Table 4.3. As can be seen in this table, the deviation errors between the experimental and predicted (by regression model) values of CO_2 adsorption and desorption capacity were 1.41% and 0.61%, respectively. The close agreement between the experimental values and those predicted from the CCD (regression) model suggests that the developed model can correlate the amination variables to CO_2 adsorption and desorption capacities with a high degree of accuracy. It is expected that the optimization results presented in this paper may provide background information for detailed process improvement research.



Figure 4.3: Contour plot of (a) CO₂ adsorption capacity and (b) CO₂ desorption capacity as a function of aminaion temperature and time for pre-oxidized adsorbents.

a)



Figure 4.4: Overlay plot of the optimal region for pre-oxidized adsorbents.

optimum conditions.				
Response	Experimental results ^a	Predicted value	Error (%)	
CO ₂ adsorption capacity (mg/g)	26.47	26.10	1.41	
CO ₂ desorption capacity (%)	95.4	94.93	0.61	

 Table 4.3: Predicted and experimental values of the studied responses obtained at optimum conditions.

^a Represents the average values of the repeated experiments.

4.2.3 Reusability of the sorbent prepared under optimum conditions

Though having high CO_2 adsorption and desorption capacities is essential for practical application of the adsorbents, stable adsorption/ desorption performance is also a crucial issue. To evaluate the reusability of the optimal activated carbon adsorbent (OXA sorbent that was aminated at 425 °C for 2.12 h) in cyclic processes, four consecutive adsorption/regeneration experiments at 105 °C were carried out, and the results are presented in Fig. 4.5. The desorption capacities presented here were calculated based on the time necessary to fully desorb CO_2 from untreated carbon. As can be seen in the figure, the adsorption performance and the desorption capacity show slight changes after first

regeneration (less than 4%). However, the alterations in these capacities after next regenerations are almost negligible (i.e., remain more or less unchanged). The regeneration potential of this adsorbent and its stability under the investigated conditions would be promising characteristics for further study toward practical applications.



Figure 4.5: Cyclical adsorption and desorption of CO₂ at 105 °C by the optimal activated carbon adsorbent.

4.3 A semi-empirical model to predict adsorption equilibrium of carbon dioxide on ammonia modified activated carbon

4.3.1 Adsorption equilibrium study

The volumetric method of measuring adsorption involves measuring the pressure change in a known volume of gas introduced to an adsorbent sample. As the gas is adsorbed and allowed to reach equilibrium, the measured decrease in the closed-system pressure yields the amount of gas adsorbed under the given conditions (Li & Tezel, 2008). Adsorption isotherms obtained at temperatures of 30, 45 and 60 °C and at pressures up to 1 atm for the ammonia-modified and untreated carbon are graphically represented in Figs. 4.6a and b, respectively. The ranges for temperature and pressure were chosen on the basis that a typical post-combustion flue gas contains approximately 10–15% CO₂ at a total pressure of 1 bar and a temperature range of 40-60 °C (Mason et al., 2011). As evident in these figures, for both modified and unmodified adsorbent, the concentration of adsorbed CO₂ increases with increasing pressure. At a given temperature, the shape of the function M = f(P) (where M is the uptake at equilibrium pressure *P*) corresponded to a type I isotherm (i.e., a monotonically concave isotherm) according to the International Union of Pure and Applied Chemistry (IUPAC) classification scheme. A type I isotherm is typically characteristic of a microporous adsorbent.

The adsorption of CO_2 is well established to possibly result from physical and/or chemical factors both influencing the adsorbent performance (Drage et al., 2007; Maroto-Valer et al., 2005; Pevida, Drage, & Snape, 2008). For the unmodified adsorbent, CO_2 capture with a pure physisorption process was proposed as a controlling mechanism. As evident in Figs. 4.6a and b, in comparison with the unmodified adsorbent, OXA-GAC exhibited a higher CO_2 uptake, particularly at low partial pressures, as reflected by the steep initial slope of the isotherms. This enhanced uptake stems from adsorption (Belmabkhout & Sayari, 2009). As the CO_2 partial pressure increases, the slopes of the OXA-GAC adsorption isotherms decrease steeply compared with the slopes of the GAC adsorption isotherms, possibly in response to a decrease of CO_2 -nitrogen surface-group interactions. At low CO_2 partial pressures, the contribution of chemisorption to the total CO_2 uptake is more pronounced, whereas physisorption within the pores becomes significant at higher concentrations (Serna-Guerrero et al., 2010b). Nevertheless, the general trend for both adsorption isotherms was a steady increase in CO_2 uptake with increasing CO_2 pressure.

As evident in the figures, the CO₂ adsorption was strongly dependent on the nitrogen functionalities: OXA-GAC exhibited a higher uptake than GAC over the whole pressure range studied. Notably, the reported values for the amount of carbon dioxide adsorbed were also in good agreement with the data from the cited literature for various ammonia-modified activated carbons (Pevida, Plaza, et al., 2008; Plaza et al., 2011; M. G. Plaza et al., 2009; Przepiórski et al., 2004).



Figure 4.6: Experimental adsorption isotherms of CO₂ on (a) modified and (b) untreated activated carbon measured at 30, 45 and 60 °C.

Moreover, as shown in Figs. 4.6a and b, both of the adsorbents adsorbed less CO₂ at 45 and 60 °C than at 30 °C. The observed decrease in CO₂ uptakes at higher temperatures is attributed to the exothermic nature of the adsorption process, where both the molecular diffusion rate and the surface adsorption energy increase with increasing temperature (Maroto-Valer et al., 2005; Shafeeyan et al., 2011). According to the isothermal data, at 30 °C and atmospheric feed pressure, the amount of CO₂ adsorbed by OXA-GAC was approximately 1.71 mol kg⁻¹, whereas only approximately 1.12 mol of CO₂/kg of the adsorbent was adsorbed at 60 °C. However, interestingly, the results at 40 °C and especially at 60 °C indicate that, compared to the unmodified adsorbent, OXA-GAC exhibited a smaller decrease in the amount adsorbed with increasing temperature. The modified adsorbent did exhibit a decrease in the amount adsorbed with increasing temperature; however, this decrease was not as pronounced as the uptake reduction observed for GAC, where physisorption is the only retention phenomenon. Because in the high-temperature adsorption, the contribution of chemisorption to the total adsorption is more significant; a possible explanation for this observation is the occurrence of strong chemical reactions between CO₂ and incorporated nitrogen-functionalities on the surface. These observations are consistent with the studies performed by Do et al. (1998) and Na et al. (2001), who reported a decrease in the amount of CO₂ adsorbed onto commercial and Ajax-activated carbon from 3.2 to 1.6 mmol g^{-1} and from 0.75 to 0.11 mmol g^{-1} when the temperature was increased from 288 to 328 K and from 298 to 373 K, respectively, at 1 atm.

4.3.2 Equilibrium isotherms modeling

Isotherm data analysis is generally a prerequisite to establish an equation that accurately predicts adsorption and that can be used to designing an adsorption separation process (Ho, 2004). As previously suggested, the overall CO_2 adsorption on the ammonia-

modified activated carbon could be the result of both physical adsorption within the pores and chemical adsorption onto the nitrogen surface groups (Shafeeyan et al., 2011; Shafeeyan et al., 2012). Distinguishing between these two mechanisms is useful in identifying the factors that may affect the rate of the adsorption process (Lu, Bai, Wu, Su, & Hwang, 2008; Su et al., 2010). Therefore, in the analysis of the adsorption equilibrium in the present study, we implemented an approach that takes into account the physical adsorption as well as the enhanced adsorption due to chemical interactions. A semiempirical model that considers the simultaneous occurrence of two independent chemical and physical adsorption mechanisms for CO₂ adsorption can be expressed as the following equation:

$$q = q_{chem} + q_{phys} \tag{4-5}$$

where q is the overall adsorption of CO₂ on ammonia-modified activated carbon, q_{chem} represents the CO₂ uptake by nitrogen functionalities and q_{phys} denotes the physical adsorption onto the porous structure.

To differentiate the contribution of each independent mechanism to the total adsorption capacity of the modified adsorbent, a procedure was used to calculate q_{phys} on the basis of the CO₂ adsorption data for the untreated activated carbon adsorbent. In our previous study, we demonstrated that, under the conditions used for the modification of the adsorbent, the ammonia treatment did not significantly alter the pore structure of the studied adsorbent. In fact, the ammonia modification yielded a material with textural characteristics very similar to those of the parent carbon (Shafeeyan et al., 2012). Given that the amount of physical adsorption is proportional to the adsorbent textural properties, the difference in the amount of CO₂ physisorbed onto both modified and untreated activated carbon under the same operating conditions is reasonably assumed to be almost

negligible. Therefore, the contribution from chemical adsorption (q_{chem}) to the total CO₂ uptake is estimated by subtracting the amount of q_{phys} for the untreated adsorbent from the overall adsorption uptake (q), measured experimentally for the modified sample. The same approach was applied successfully by Serna-Guerrero et al. to differentiate the contributions of each independent mechanism of CO₂ adsorption on amine-grafted mesoporous silica (Serna-Guerrero et al., 2010b). They also provided further evidence of the applicability of the assumption used in the present study and extended it to other adsorption processes that combine the contribution of physisorption and chemisorption mechanisms.

Accordingly, using the equilibrium adsorption data for GAC, and subtracting the measured contribution of physisorption from the overall uptake, the corresponding values of q_{chem} for OXA-GAC at different temperatures were calculated; the results are represented in Fig. 4.7. As evident in this figure, the amount of CO₂ chemisorbed by the nitrogen functional groups corresponds to a type I isotherm in the IUPAC classification scheme, consistent with adsorption due to chemical interactions.



Figure 4.7: Calculated adsorption isotherms for CO₂ chemisorption onto the modified adsorbent at 30, 45 and 60 °C.

After the contributions of the two independent physisorption and chemisorption mechanisms were distinguished, different equilibrium isotherm equations could potentially be used to correlate equilibrium data. Applying various conventional isotherm models, we can express the terms on the right side of Eq. (4-5) in different forms to describe each of the mechanisms with a proper isotherm model. The main distinguishing feature for selecting an appropriate isotherm model arises from the accuracy as well as simplicity of the equations. Non-idealities in the isotherms mainly stem from the heterogeneity of the adsorbent surface, and such heterogeneities often occur in the case of activated carbon (Malek & Farooq, 1996). From this point of view, three of the most common pure-species isotherm models (i.e., Freundlich, Sips and Toth) that have been previously applied to describe adsorption equilibria on heterogeneous adsorbents were employed in this work.

For each temperature and mechanism, a non-linear regression method was applied to independently determine the parameters corresponding to the aforementioned isotherm models. The optimal values of the Freundlich, Sips and Toth isotherm parameters are summarized in Tables 4.4-4.6, respectively. To quantitatively compare the quality of the nonlinear regressions for these three models, the average relative error (ARE) and nonlinear regression coefficient (R^2) were calculated; the results are included in Tables 4.4-4.6. As evident in these tables, the parameters of each isotherm model varied when the adsorption mechanism and temperature were changed. In the Freundlich model, the K_r parameter refers to the adsorption capacity and represents the adsorption quantity per unit equilibrium concentration, whereas the exponent $1/m_r$ is a measure of the adsorbent-adsorbate binding energy and expresses both the adsorption intensity and the surface heterogeneity (Ferraro, Cruz, Jorge, Pintado, & Castro, 2013). The higher values of K_r obtained for the physisorption mechanism indicated greater adsorption compared

to chemisorption, whereas the higher values of $m_{\rm F}$ observed for chemisorption denoted a more favorable adsorption and a stronger bond between CO₂ and the modified adsorbent $(5 < m_F < 25)$. The heterogeneity of the adsorbent surface can also be described with the exponent m_i in the Sips and Toth isotherm equations (Esteves et al., 2008). When the surface is homogeneous, m_i is equal to unity and the isotherm expressions reduce to the Langmuir equation. In agreement with the Freundlich model, the obtained values of the surface heterogeneity parameter in the Toth $(m_T < 1)$ and Sips $(m_S > 1)$ equations showed a higher degree of heterogeneous adsorption for CO₂ chemisorption. From Tables 4.5 and 4.6, the values of the saturation capacity parameter (q_m) , which indicate the maximum amount that can possibly be adsorbed, decreased with increasing temperature. The observed decrease is associated with the exothermicity of the adsorption process (Ning et al., 2012). Because η is the parameter that reflects the temperature dependence of q_m in the Toth isotherm (Table 4), the lower values of η obtained for the chemisorption compared to the physisorption imply that, within the temperature range studied, the chemical adsorption mechanism exhibited a greater tendency to remain the same with changes in temperature. Furthermore, compared with physical adsorption, the chemisorption mechanism presented larger values of K_i (for instance, q_{chem} (1.05×10⁵ atm⁻¹) compared to q_{phys} (7.1×10⁻¹ atm⁻¹) at 303 K), which is attributed to the strength of the adsorbate-adsorbent interactions for the chemisorption mechanism (Do, 1998).

	Temperatures		
	303 K	318 K	333 K
Physical adsorption			
$K_F\left(molkg^{-1}atm^{-1/m_F} ight)$	1.22	0.93	0.68
m_F (dimensionless)	1.40	1.45	1.24
R^2	0.996	0.995	0.996
ARE (%)	5.56	5.95	5.02
Chemical adsorption			
$K_F\left(molkg^{-1}atm^{-1/m_F} ight)$	0.48	0.47	0.46
m_F (dimensionless)	24.57	14.75	8.41
R^2	0.962	0.965	0.865
ARE (%)	0.86	0.79	5.16

Table 4.4: Freundlich isotherm parameters with R^2 and ARE for each independent mechanism at temperatures of 303, 318, and 333 K

Table 4.5: Sips isotherm parameters with R^2 and ARE for each independent mechanism at temperatures of 303, 318, and 333 K

	Temperatures		
	303 K	318 K	333 K
Physical adsorption			
$q_m(mol kg^{-1})$	3.43	2.12	1.76
$K_{s}(atm^{-1})$	0.69	0.39	0.23
m_s (dimensionless)	1.11	1.08	0.98
R^2	0.998	0.998	0.999
ARE (%)	2.93	2.08	1.71
Chemical adsorption			
$q_m(molkg^{-1})$	0.53	0.48	0.44
$K_{s}(atm^{-1})$	8.73×10^{2}	1.96×10^{2}	50.76
m_s (dimensionless)	2.12	2.02	1.83
R^2	0.970	0.981	0.991
ARE (%)	0.54	0.67	0.99

	Temperatures		
	303 K	318 K	333 K
Physical adsorption			
$q_m(mol kg^{-1})$	5.01	2.58	1.32
$K_T(atm^{-1})$	0.71	1.09	1.50
m_T (dimensionless)	0.59	0.63	0.68
R^2	0.999	0.999	0.999
ARE (%)	2.65	2.01	1.65
Chemical adsorption			
$q_m(mol kg^{-1})$	0.54	0.49	0.46
$K_T(atm^{-1})$	1.05×10^{5}	3.76×10 ⁵	1.20×10^{6}
m_T (dimensionless)	0.29	0.30	0.31
R^2	0.971	0.982	0.992
ARE (%)	0.52	0.64	0.93

Table 4.6: Toth isotherm parameters with R^2 and ARE for each independent mechanism at temperatures of 303, 318, and 333 K

On the basis of the calculated values of ARE and R^2 tabulated in Tables 4.4-4.6, both the Sips and Toth isotherms were capable of fitting the equilibrium data over a broad range of experimental conditions. However, the Toth equation, which involves a symmetrical quasi-Gaussian distribution of adsorption sites, provided a slightly better fit. The low values obtained for the ARE (in no case greater than 3%), as well as the high values of the nonlinear regression coefficient (very near unity, $R^2 \ge 0.97$), indicate the goodness of the fit. Therefore, compared to the Freundlich and Sips models, the Toth equation is more accurate and more capable of describing the CO₂ adsorption isotherms over the ammonia-modified adsorbent. Thus, only the Toth model was used here to illustrate the quality of its fit to the experimental equilibrium data (Fig. 4.8). The complete form of the proposed semi-empirical model is expressed as follows:

$$q = \left[\frac{q_m K_T P}{\left(1 + \left(K_T P\right)^{m_T}\right)^{1/m_T}}\right]_{phys} + \left[\frac{q_m K_T P}{\left(1 + \left(K_T P\right)^{m_T}\right)^{1/m_T}}\right]_{chem}$$
(4-6)

where the subscripts "*phys*" and "*chem*" indicate the contributions of each independent mechanism to the total CO₂ uptake.

To express the temperature dependence of the Toth isotherm parameters for the purpose of interpolating or extrapolating the equilibrium data to various temperatures, as well as determining the isosteric enthalpy of adsorption, the parameters, K_i , q_m and m_i are described by the following equations (Li & Tezel, 2008):

$$K_{T} = K_{T_{0}} \exp\left[-\frac{\Delta H}{RT_{0}}\left(\frac{T_{0}-T}{T}\right)\right]$$
(4-7)

$$q_m = q_{m_0} \exp\left[\eta\left(\frac{T_0 - T}{T_0}\right)\right] \tag{4-8}$$

$$m_T = m_{T_0} + \alpha \left(\frac{T - T_0}{T}\right) \tag{4-9}$$

Where *T* is the absolute temperature (K), T_0 is the reference temperature and was taken as 303 K, K_{i0} and m_{i0} are the affinity and heterogeneity parameters at the reference temperature, respectively, α and η are constant parameters, $(-\Delta H)$ is the isosteric heat of adsorption at zero coverage (kJ mol⁻¹), and *R* is the gas-law constant (J mol⁻¹ K⁻¹). The optimal parameter values for both the chemical and physical adsorption mechanisms were obtained by nonlinear regression; the results are presented in Table 4.7. The surface obtained from the global fitting of the aforementioned model to the experimental data (Fig. 4.8) shows that the equilibrium data are described well for all temperatures when the adsorption isotherm plotted according to the proposed Toth equation is used.

	α	η	$\left(-\Delta H ight)$
	(dimensionless)	(dimensionless)	(kJ mol ⁻¹)
Physical adsorption	0.96	13.43	23.05
Chemical adsorption	0. 25	1.67	68.11

Table 4.7: Optimal values of the proposed Toth temperature-dependent parameters



Figure 4.8: Graphical evaluation of the fit of the experimental equilibrium data to the proposed model for the modified adsorbent, whose parameters are presented in Tables 4.6 and 4.7 The surface is the global isotherm model, and the black and white circles show the experimental data at 303, 318 and 333 K.

4.3.3 Isosteric heat of adsorption

The isosteric heat of adsorption is a critical parameter for designing and operating an adsorption-based separation process such as PSA. It is a measure of the interaction between the adsorbate molecules and the surface of the adsorbent, and it can be obtained from the temperature dependence of the adsorption isotherm (Esteves et al., 2008; Ning et al., 2012). The isosteric heat of adsorption, Q_{st} , also denoted as $-\Delta H$ (kJ mol⁻¹), at a specific adsorbate loading can be estimated using the Clausius–Clapeyron equation as follows (Park, Lee, Choi, Lee, & Kim, 2002):

$$-\frac{Q_{st}}{R} = \left(\frac{\partial \ln P}{\partial 1/T}\right)_{q^a} \tag{4-10}$$

where q^a is a specific surface loading (mol kg⁻¹).

Integration of Eq. (4-10) enables the calculation of Q_{st} by knowing the differential partial pressure change as a function of a differential change in system temperature for a given adsorbed amount:

$$Q_{st} = R \left[\frac{\ln P_1 - \ln P_2}{1/T_2 - 1/T_1} \right]_{q^a}$$
(4-11)

Therefore, the isosteric heats are obtained from the slopes of the lines in the plots of $\ln P$ vs. 1/T for the same adsorption amount for the different isotherms. In the present study, the experimental isotherms at the three temperatures of 303, 318, and 333 K for a given adsorption amount (from 0.2-1.2 mol kg⁻¹ in 0.2 mol kg⁻¹ increments) were used to measure the isosteric enthalpies of adsorption for OXA-GAC and GAC adsorbents (see Figs. 4.9a and b).

The observed linear behavior is based on the assumption that, over the temperature range investigated, the isosteric heat of adsorption is independent of temperature and depends only on the surface coverage. In practice, the assumption of temperature independence of the heat of adsorption was shown to be valid for the pressure and loading estimations (Berlier & Frère, 1996; Esteves et al., 2008). Notably, consistent with the exothermic character of the adsorption process, negative Q_{st} values were attained at 303-333 K, indicating that the entropy was reduced in the system.



Figure 4.9: Adsorption isosteres of CO_2 for (a) modified and (b) untreated adsorbent in the temperature range from 303 to 333 K. The points were calculated by numerical interpolation, and the lines represent the linear fit. All of the isosteres are marked with the corresponding amount of CO_2 adsorbed in units of (cm³ STP/g).

The isosteric enthalpy changes accompanying adsorption can be used to examine the molecular-scale interactions between the adsorbate molecules and the adsorbent. Moreover, information regarding the magnitude of the adsorption enthalpy and its surface coverage dependence can be used as criteria for determining the energetic heterogeneity of a solid surface (Belmabkhout & Sayari, 2009; Park et al., 2002). The isosteric heat is independent of the surface coverage when no interaction occurs between adsorbed molecules, and the surface is energetically homogeneous. However, a variation of the isosteric heats with the surface loading indicates the existence of different levels of surface energy and heterogeneity of the adsorbent surface (Ning et al., 2012). Figure 4.10 depicts the variation of Q_{st} (calculated using the Clausius-Clapeyron equation) as a function of the amount of CO₂ adsorbed for the OXA-GAC and GAC adsorbents. As evident in the figure, the unmodified adsorbent exhibited approximately constant values of Q_{st} for different CO₂ loadings, indicating the uniform nature of the adsorbent surface. Notably, the obtained initial heat of adsorption for GAC (25 kJ mol⁻¹) corresponds closely to the value of $-\Delta H$ (23 kJ mol⁻¹) calculated from the temperature-dependent parameters of the proposed model for q_{phys} and is also of the same order of magnitude as the typical values for commercial activated carbon reported in the literature (Chue et al., 1995; Esteves et al., 2008; Garcés et al., 2013; Himeno, Komatsu, & Fujita, 2005; Van Der Vaart, Huiskes, Bosch, & Reith, 2000). In contrast, OXA-GAC exhibited rather high values of Q_{st} at the low range of loading, which reflects the relatively strong chemical interactions between CO₂ and the basic nitrogen functionalities.



Figure 4.10: Concentration dependence of the isosteric enthalpy for CO_2 adsorption on the GAC and the OXA-GAC

The isosteric enthalpy of adsorption over OXA-GAC decreases with increasing CO₂ surface loading until it approaches values similar to those of the untreated carbon adsorbent. The observed decrease with increasing loading is attributed to the high degree of heterogeneity of the adsorbent surface (Belmabkhout & Sayari, 2009). At the beginning of the adsorption process, only the most active adsorption sites are filled, the activation energy is low and the heat of adsorption is high. With increasing surface coverage, less-active adsorption sites start to be occupied, thereby increasing the activation energy and decreasing the heat of adsorption (Himeno, Tomita, Suzuki, & Yoshida, 2007; Ning et al., 2012; Zhao, Xu, Sun, Zhang, & Liu, 2009). The observed heats of adsorption for the ammonia-modified adsorbent clearly demonstrate the strong and weak interactions of the nitrogen surface groups with CO₂ at low and high coverages, respectively. Notably, the Q_{st} value obtained at low coverage (70.4 kJ mol⁻¹) is consistent with the value calculated from the proposed Toth parameter K_i for q_{clean} (68.1 kJ mol⁻¹)

at zero loading and is in the range of values for typical cases of CO_2 chemisorption (60 to 90 kJ mol⁻¹) (Samanta et al., 2011).

Indeed, this value is higher than that for a physical interaction but lower than that for a strong chemical interaction. In addition, in agreement with the observed inflection in the isotherms for CO₂ adsorption over the OXA-GAC adsorbent (see Fig. 4.6a), a corresponding curvature in the plot of the isosteric heat vs. coverage was detected, coinciding with the saturation of the most active adsorption sites (indicated by an arrow in Fig. 4.10). The observed variation of the slope of the Q_{st} vs. CO₂ loading curve clearly reflects the occurrence of two independent adsorption mechanisms.

4.4 Modeling of carbon dioxide adsorption onto ammonia-modified activated carbon: Kinetic analysis and breakthrough behavior

4.4.1 Adsorption kinetics

Fig. 4.11 displays the experimental data of CO₂ uptake as a function of time over the ammonia-modified and untreated adsorbents at 30, 45, and 60 °C determined using the gravimetric uptake method. As shown in Fig. 4.11, the adsorption kinetics over OXA-GAC were faster than the untreated carbon, particularly at high temperatures, and the uptake capacity increased rapidly during the first minutes of the process. For example, at 60 °C, OXA-GAC reached approximately 90% of its maximum capacity within the first 70 seconds of CO₂ exposure (Fig. 4.11c). However, the CO₂ uptake rate then decelerates, and approximately 1,200 seconds are required to reach a constant equilibrium value. The obtained experimental kinetic data for CO₂ adsorption were fitted to three different kinetic models (i.e., pseudo-first-order, pseudo-second-order, and Avrami models), and the corresponding predicted kinetic curves are presented in Fig. 4.11. These three kinetic models are frequently used for the mass transfer coefficient determination of various

gases on different adsorbents, including activated carbon (Heydari-Gorji & Sayari, 2011; Loganathan et al., 2014; Serna-Guerrero & Sayari, 2010; Stevens et al., 2013). A nonlinear regression method was used to determine the parameters corresponding to the mentioned kinetic models; the values obtained are summarized in Tables 4.8 and 4.9. To quantify and compare the quality of the nonlinear regressions for these three models, the associated coefficients of determination (R^2) and the normalized standard deviation (Δq) were calculated using Eqs. 3.23 and 3.24; the results are presented in Tables 4.8-4.9. These tables reveal that the parameters of each kinetic model varied when the adsorption temperature and type of adsorbent were changed.



Figure 4.11: Experimental CO₂ adsorption onto modified and untreated adsorbents at (a) 30 °C, (b) 45 °C, and (c) 60 °C along with the corresponding fit to kinetic models.

'Figure 4.11, continued'




Although because of the exothermic nature of the adsorption process, the CO₂ uptake decreased with increasing temperature, as evident in Tables 4.8 and 4.9, for each adsorbent, the rate constants (k_F , k_S , and k_A) obtained using the kinetic models increased when the higher adsorption temperatures were applied. The favorable adsorption kinetics observed at higher temperatures can be attributed to the faster migration of CO₂ molecules inside of the pores due to the increase in their kinetic energy, which results in an increase in the diffusion rate (Loganathan et al., 2014). The increase in the mass transfer coefficient with the rise in temperature is reflected in the faster kinetics of adsorption and the significantly steeper kinetic curves. For example, at 30 °C, the time required for GAC to reach approximately 85% of its equilibrium capacity was ($\sim <350$ s), whereas this time was reduced to ($\sim <140$ s) when the temperature was increased to 60 °C (Fig. 2a and c). The trend observed agrees well with the data presented in the literature in which the kvalues were reported to increase with an increasing adsorption temperature (Loganathan et al., 2014; Serna-Guerrero & Sayari, 2010; Stevens et al., 2013; Zhijuan Zhang, Zhang, Chen, Xia, & Li, 2010). In addition, consistent with the previous observation that the kinetic behavior of the modified sample was comparatively faster than that of untreated adsorbent, the calculated values of the rate constants (k_F , k_S , and k_A) for OXA-GAC were higher than those for GAC. As shown in Tables 4-5, GAC exhibited a k_A of 1.14×10^{-2} at 60 °C, whereas it increased to 4.01×10^{-2} for the modified sample under the same conditions. In addition, the estimated values of parameter n in the Avrami model for untreated adsorbent were lower compared with the values for its modified counterpart. A variation in the value of this parameter, which describes how fast the adsorption process occurs, suggests a change in the adsorption mechanism (Cestari, Vieira, Vieira, & Almeida, 2006; Wang et al., 2012). This result may be attributed to the occurrence of chemical reactions between the CO₂ and nitrogen functionalities incorporated onto the adsorbent surface, in addition to CO2 capture by physical adsorption. Notably, the

calculated values of the kinetic rate constants for the untreated and modified adsorbents (~ $8 \times 10^{-3} < k < ~ 4 \times 10^{-2}$) were consistent with the typical values reported in the literature for various activated carbons (Do & Wang, 1998; Dreisbach, Staudt, & Keller, 1999).

Fig. 4.11 shows that for both of the adsorbents, the pseudo-first-order kinetic model reasonably fit the experimental kinetic curves of CO₂ adsorption at all of the studied temperatures. However, a slight deviation was observed at high surface coverage (underestimation of the CO₂ uptake) because as previously reported by other authors (Ho, 2006; Loganathan et al., 2014; Serna-Guerrero & Sayari, 2010), the most accurate range for the pseudo-first-order model to fit the kinetic curves occurs in the early stages of adsorption. Moreover, in the case of adsorption onto OXA-GAC, the observed deviation may also be due to the simultaneous occurrence of physisorption and chemisorption in which the assumption that the uptake rate is proportional to the linear difference between the equilibrium and the actual concentration is no longer satisfactory (Stevens et al., 2013). However, the estimated values of Δq for the pseudo-first-order kinetic model varied by less than 3% for the temperature range studied, suggesting that the model can be applied with a reasonable degree of confidence to describe the CO₂ capture kinetics of both of the adsorbents. The advantage of using the pseudo-first-order kinetic model is the mathematical simplicity of its expression. Using this expression of the adsorption rate, which does not involve the spatial coordinates, the partial differential equation describing mass conservation for gas penetrating pores can transform into a significantly simpler ordinary differential equation; thus, the solutions are mathematically simpler and faster than the solution of the diffusional models (Shafeeyan et al., 2014). Yang and Lee have demonstrated that although this adsorption rate model is relatively simple, it can predict the experimental data with satisfactory accuracy (Yang & Lee, 1998). Consequently, this approximation has found widespread application in modeling fixed-bed and cyclic CO₂ adsorption processes.

In contrast, as shown in Fig. 4.11, the pseudo-second-order model significantly deviates from the experimental results at the beginning of the process (low surface coverage) and only appears to fit the kinetic data when the adsorbate loadings become sufficiently high. Based on the calculated values of R^2 and Δq (%) presented in Tables 4-5 for both of the studied adsorbents, the Avrami model provided the best fit to the experimental kinetic data over the range of temperatures considered. The high values obtained for the nonlinear regression coefficient (close to unity, $R^2 \ge 0.99$) and the low values of the Δq (%) (in no case greater than 2%) indicate the goodness of fit over the range of the recorded data. Therefore, compared with the pseudo-first-order and pseudo-second-order kinetic models, the Avrami equation is more accurate and more capable of describing CO₂ adsorption kinetics over the studied adsorbents. The excellent quality of the fit of the Avrami model to the experimental kinetic data at the low and high surface coverages is most likely associated with its potential to account for the occurrence of complex reaction pathways (Cestari et al., 2006; Lopes, dos Anjos, Vieira, & Cestari, 2003; Serna-Guerrero & Sayari, 2010; Wang et al., 2012). The findings are in good agreement with previous studies, which have demonstrated that the pseudo-first-order and Avrami kinetic models successfully described the adsorption kinetics of CO₂ on activated carbon (Zhijuan Zhang, Zhang, et al., 2010) and MCM-41 adsorbents (Berenguer-Murcia et al., 2003) adsorbents) and functionalized pore-expanded mesoporous (physical silica (physicochemical adsorbent) (Loganathan et al., 2014), respectively. Because the Avrami model provided the best experimental simulation fit, we employed this equation for the modeling of the fixed bed CO₂ adsorption.

Adsorption temperature (°C)		30	45	60
Pseudo-first order model				
	$K_F(s^{-1})$	7.99×10 ⁻³	1.01×10 ⁻²	1.13×10 ⁻²
	R^2	0.9970	0.9961	0.9942
	$\Delta q(\%)$	1.71	1.84	1.96
Pseudo-second order				
model	$K_S \pmod{\mathrm{kg}^{-1} \mathrm{s}^{-1}}$	8.01×10 ⁻³	9.97×10 ⁻³	1.16×10 ⁻²
	R^2	0.9860	0.9885	0.9897
	$\Delta q(\%)$	6.61	3.94	3.46
Avrami model				
	K_A (s ⁻¹)	7.97×10 ⁻³	1.01×10 ⁻²	1.14×10 ⁻²
	n	1.03	0.96	0.91
	R^2	0.9979	0.9981	0.9994
	$\Delta q(\%)$	1.40	1.39	1.01

Table 4-8: The calculated parameters of the kinetic models and associated R^2 and Δq (%) for the CO₂ adsorption onto GAC at different temperatures.

Table 4-9: The calculated parameters of the kinetic models and associated R^2 and Δq (%) for the CO₂ adsorption onto OXA-GAC at different temperatures.

Adsorption temperature (°C)		30	45	60
Pseudo-first order model				
	$K_F(s^{-1})$	3.15×10 ⁻²	3.68×10 ⁻²	3.97×10 ⁻²
	R^2	0.9961	0.9947	0.9909
	$\Delta q(\%)$	2.08	2.68	3.09
Pseudo-second order				
model	$K_S (\mathrm{mol} \mathrm{kg}^{-1} \mathrm{s}^{-1})$	2.78×10 ⁻²	3.29×10 ⁻²	4.02×10 ⁻²
	R^2	0.9841	0.9958	0.9942
	$\Delta q(\%)$	6.72	1.94	2.27
Avrami model	K_A (s ⁻¹)	3.11×10 ⁻²	3.53×10 ⁻²	4.01×10 ⁻²
	n	1.74	1.53	1.38
	R^2	0.9970	0.9981	0.9943
	$\Delta q(\%)$	1.63	1.48	2.03

To describe the temperature dependence of the rate constants, the following Arrheniustype equation was employed:

$$k = k_0 \exp\left(-E_a / RT\right) \tag{4-12}$$

where k_0 is the pre-exponential factor, R denotes the universal gas constant, E_a is the apparent activation energy, and T expresses the absolute temperature.

Because the fit statistics of the Avrami kinetic model were adequate, its corresponding k constants at 30, 45, and 60 °C were used to estimate the parameters of the Arrhenius equation (Fig. 4.12). From the slope and the intercept of the straight line of the plot of the natural logarithm of k_A against the inverse of the absolute temperature, E_a values of 7.11 and 10.06 kJ mol⁻¹ and k_0 values of 5.21×10^{-1} and 4.38×10^{-1} (s⁻¹) were determined for the OXA-GAC and GAC adsorbents, respectively. The lower value of E_a on the OXA-GAC sample compared with the untreated adsorbent indicates the stronger adsorbate-adsorbent interaction potential of the modified adsorbent compared with the GAC sample. The plots of ln k_A vs 1/T were linearly fitted, with coefficients of determination (R^2) greater than 0.97, indicating a good linearity between ln k_A and 1/T.



Figure 4.12: Arrhenius plots for the estimation of the CO₂ adsorption activation energies on the GAC and OXA-GAC adsorbents.

4.4.2 Column breakthrough experiments and model validation

Breakthrough adsorption experiments for the ammonia-modified and untreated adsorbents were performed at different adsorption temperatures and feed flow rates. Fig. 4.13 represents the effluent CO₂ concentration as a function of time (breakthrough curve) for each sample at 30, 45, and 60 °C with varying CO₂ flow rates from 50 to 100 ml min⁻¹ in the feed gas. To describe the breakthrough behavior of the CO₂ adsorption, the characteristic parameters of the breakthrough time, t_b , saturation time, t_s , length of the mass transfer zone, L_{MTZ} , and the dynamic adsorption capacity, q, were determined. The obtained results for each experimental breakthrough curve are tabulated in Table 4.10. The breakthrough and saturation times were considered as the required times to reach 10% and 90% of the inlet CO₂ concentration at the column outlet, whereas the equilibrium adsorption capacity was taken as the loading amount until the point at which the effluent CO₂ concentration was equal to the feed concentration. Assuming a constant pattern profile, the length of the mass transfer zone, which is the active part of the fixed bed where adsorption occurs, was estimated using the following correlation (Zhao, Shen, Bai, & Ni, 2012):

$$L_{MIZ} = L \left[\frac{t_{s} - t_{b}}{t_{s} - 0.5(t_{s} - t_{b})} \right]$$
(4-13)

where *L* is the bed height (cm).

The equilibrium dynamic adsorption capacity of the bed was determined as a function of the feed concentration applying a mass balance to the bed, as shown in Eq. (4-14):

$$q = \frac{Q_F c_0 t_q}{m_S} \times \frac{T_0}{T} \times \frac{1}{V_m}$$
(4-14)

where $q \pmod{\text{kg}^{-1}}$ is the calculated CO₂ loading at a certain time, $m_S \pmod{\text{kg}}$ denotes the mass of adsorbent loaded in the column, $Q_F \pmod{\text{min}^{-1}}$ is the volumetric flow rate of the

feed gas, c_0 is the inlet CO₂ concentration (vol.% or mol%), *T* (K) expresses the gas temperature, T_0 is 273.15 (K), V_m is (22.4 ×10³ ml mol⁻¹), and t_q is the time equivalent to the CO₂ stoichiometric adsorption capacity of the bed and is calculated from the area between the breakthrough profile and a line at $c_t/c_0 = 1.0$, as described in Eq. (20) (Liu et al., 2010; Y. Liu et al., 2011; Monazam et al., 2013; Serna-Guerrero & Sayari, 2010):

$$t_q = \int_0^\infty (1 - \frac{c_t}{c_0}) dt$$
 (4-15)

where c_t (vol %) is the concentration of CO₂ at time *t*.

The efficiency of the column (Φ) was calculated based on the conventional concept of the length of unused bed (LUB) (Beaver, Caram, & Sircar, 2009):

$$\frac{LUB}{L} = 1 - \frac{t_b}{t^*} = 1 - \Phi \tag{4-16}$$

where t^* is the stoichiometric breakthrough time and can be determined by extending a vertical line on the breakthrough profile at a point at which the adsorbed quantity of adsorbate is equal to its unadsorbed quantity in the effluent. Assuming a symmetric breakthrough curve, t^* can also be estimated as $t^* = (t_s + t_b)/2$ (Monazam et al., 2013). Based on the above definition, in Eq. 4-16, Φ is the portion of the bed capacity used for CO₂ capture when incipient breakthrough occurs at the column effluent ($t=t_b$) (Beaver et al., 2009; García et al., 2011).

conditions										
ID	Т	Q	$t_{b(exp)}$	$t_{b(sim)}$	$t_{s(exp)}$	$t_{s(sim)}$	L _{MTZ}	Φ	q_{exp}	q_{sim}
	(°C)	(ml min ⁻¹)	(min)	(min)	(min)	(min)	(cm)	-	(mol kg ⁻¹)	(mol kg ⁻¹)
GAC	30	50	3.3	3.3	7.0	6.9	14.37	0.64	0.28	0.28
GAC	45	50	2.7	2.7	5.6	5.5	13.97	0.65	0.21	0.21
GAC	60	50	1.6	1.6	3.2	3.1	13.33	0.67	0.12	0.12
GAC	30	100	1.5	1.5	3.4	3.5	15.51	0.61	0.26	0.27
GAC	45	100	1.1	1.1	2.4	2.5	14.86	0.63	0.19	0.20
GAC	60	100	0.7	0.7	1.5	1.5	14.54	0.64	0.11	0.11
OXA- GAC	30	50	10.9	10.8	13.5	13.3	4.26	0.89	0.67	0.63
OXA- GAC	45	50	9.7	9.6	11.9	11.8	4.07	0.90	0.57	0.55
OXA- GAC	60	50	8.3	8.2	10.0	9.9	3.72	0.91	0.46	0.45
OXA- GAC	30	100	4.4	4.4	6.4	6.4	7.40	0.81	0.60	0.60
OXA- GAC	45	100	3.9	4.0	5.6	5.7	7.16	0.82	0.51	0.52
OXA- GAC	60	100	3.3	3.3	4.7	4.8	7.00	0.83	0.40	0.41

 Table 4-10: Breakthrough curve characteristic parameters under different operating conditions

4.4.2.1 Effect of temperature on the breakthrough profile of CO₂ adsorption

The effect of temperature on the breakthrough curves of CO₂ adsorption on both of the studied adsorbents at feed rates of 50 and 100 ml min⁻¹ is illustrated in Fig. 4.13 As shown in this figure (and in Table 4.10), for both of the adsorbents and flow rates investigated, the CO₂ breakthrough times and dynamic adsorption capacities significantly decreased with increasing adsorption temperature due to the exothermic nature of the adsorption process (Auta & Hameed, 2014; Lua & Yang, 2009; Mulgundmath et al., 2012; Y. Zhao et al., 2012). However, the results indicated that compared with OXA-GAC, the breakthrough curves obtained by GAC shows a shorter breakthrough time with increasing temperature but not as pronounced as the time reduction observed for GAC in which the physisorption is the dominant feature. For example, under a feed rate of 50 ml

min⁻¹ (Fig. 4.13b), the breakthrough time by OXA-GAC exhibited a 25% decrease (from 10.9 s to 8.2 min) for an increase in temperature from 30 to 60 °C, whereas a decrease of 51% (from 3.3 s to 1.6 min) was observed for the GAC over the 30–60 °C temperature range (Fig. 4.13a). Similarly, compared with the GAC, the modified adsorbent exhibited a lower decrease in the equilibrium dynamic capacity with a rising temperature from 30 to 60 °C (31% decrease vs 57% decrease at 50 ml min⁻¹ and 33% decrease vs 58% decrease at 100 ml min⁻¹). A plausible interpretation of this observation is that in contrast to the expected reduction in the physisorption capacity of both of the adsorbents with increasing temperature, chemisorption through the nitrogen functionalities incorporated onto the surface of OXA-GAC aid the retention of the adsorption capacity at higher temperatures.

The findings are in accordance with the studies of García et al. and Thote et al. reporting that the breakthrough time and the dynamic CO_2 adsorption capacity of a nitrogenenriched carbon and a commercial activated carbon decreased with increasing temperature (García et al., 2011; Thote et al., 2010).



Figure 4.13: Effect of temperature on breakthrough curves of CO_2 adsorption onto (a) GAC and (b) OXA-GAC adsorbents at a 50 ml min⁻¹ feed flow rate and (c) GAC and (d) OXA-GAC at a 100 ml min⁻¹ feed flow rate.



Moreover, for a given adsorbent sample and the feed flow rate studied, a lower adsorption temperature leads to longer saturation times. For example, at a feed flow rate of 100 ml min⁻¹ when the temperature decreased from 60 to 30 °C, the time required to reach the saturation level increased from 1.5 to 3.4 min and from 4.7 to 6.4 min for the GAC and

the OXA-GAC adsorbents, respectively (Fig. 4.13 c and d). Notably, the slope of the OXA-GAC breakthrough curves was steeper than that obtained for the untreated sample. In addition, the breakthrough profile became steeper as the temperature increased. Therefore, the use of modified adsorbent and increasing temperatures favor mass transfer and decrease the mass transfer zone (L_{MTZ}), consequently increasing the column efficiency (Φ) (Fig. 4.14). For example, the efficiency of the column packed with GAC adsorbent under a 50 ml min⁻¹ feed flow rate at 30 °C increased from 0.64 to 0.67 when the temperature increased to 60 °C, whereas it increased to 0.89 when the OXA-GAC adsorbent was at 30 °C and the same feed flow rate was used (Table 4.10). This observation is consistent with the results presented in section 4.4.1 in which the rate constants increased with increasing adsorption temperature and/or the use of a modified sample.



Figure 4.14: The efficiency of the adsorption bed for the breakthrough experiments with GAC and OXA-GAC adsorbents at varying adsorption temperatures and feed flowrates.

4.4.2.2 Effect of feed flow rate on the breakthrough curve for the adsorption of CO₂ Fig. 4.15 shows the effect of the feed flow rate on the breakthrough curves obtained at 30, 45, and 60 °C for both of the studied adsorbents. This figure and Table 4.10 show that for both of the adsorbents over the temperature range studied, a considerably shorter breakthrough and saturation time and a relatively steeper breakthrough curve were achieved as the flow rate increased. This result was obtained because with an increasing flow rate, the degree of turbulence and mixing increases, and breakthrough occurs faster with more adsorbate entering the bed per unit time (Lua & Yang, 2009; Monazam et al., 2013; Mulgundmath et al., 2012). A faster feed flow rate results in a higher Reynolds number and may cause the mass transfer zone to move faster, leading to a reduction in the column efficiency (Fig. 4.14). Conversely, the lower flow rate of the feed gas was accompanied by the slower transport of the adsorbate molecules, leading to a prolonged breakthrough time, thereby increasing the diffusion coefficient (Auta & Hameed, 2014). In line with this finding, for both of the adsorbents over the temperature range studied, the calculated values of the dynamic adsorption capacity decreased as the feed flow rate varied from 50 to 100 ml min⁻¹. These observations are in accordance with those reported elsewhere (Ahmed, Mohammed, & Kadhum, 2011; Auta & Hameed, 2014; K.-S. Hwang et al., 2010; Sabouni, Kazemian, & Rohani, 2013). The use of OXA-GAC adsorbent at 60 °C (Table 6) resulted in the most significant decrease in the measured breakthrough time (from 8.3 s to approximately 3.3 min) and dynamic adsorption capacity (from 0.46 to approximately 0.4 mol kg⁻¹) when the flow rate was increased from 50 to 100 ml min⁻ ¹. Note that the enhanced adsorption capacity of the OXA-GAC adsorbent compared with the GAC sample at higher adsorption temperatures is primarily attributed to chemisorption with the attached nitrogen surface groups. Therefore, with an increasing flow rate and a decreasing residence time of the adsorbate in the bed, the observed chemisorption significantly decreases because the contact time for the CO_2 molecules to

react with the incorporated N-functionalities decreases. The results also indicated that at the tail-end part of the experimental breakthrough curves in which the c_t/c_0 ratio reaches 1, the higher feed flow rates led to the slower approach of the ratio to unity compared with the lower flow rates (Fig. 4.15).



Figure 4.15: Effect of feed flow rate on breakthrough curves of CO_2 adsorption onto GAC and OXA-GAC adsorbents at (a) 30 °C (b) 45 °C and (c) 60 °C.



4.4.2.3 Validation of the proposed model

The breakthrough curves were simulated using the adsorption bed characteristics and the model parameters presented in Tables 3.3 and 3.5, respectively. To test the validity of the proposed mathematical model, a series of experimental breakthrough profiles was compared with the theoretical breakthroughs for different adsorption temperatures, flow rates, and types of adsorbents. Fig. 4.16 illustrates the representative experimental data (discrete symbols) and the corresponding simulations (solid lines) under different operating conditions. An evaluation of the three verification cases demonstrates a reasonable agreement between the measured and simulated breakthrough curves over the entire c_t/c_0 time. This result indicates that the estimated parameters and the hypothesis on which the model is based are valid for the system studied. However, a slight discrepancy was detected between the predicted and experimental breakthrough curves, which can be attributed to the experimental errors. The simulated values of the breakthrough time,

 $t_{b(sim)}$, exhibited relative error deviations of less than 2% from the experimental data, $t_{b(exp)}$, whereas relative error deviations of up to ca. 4% and 6% were observed in the estimation of the saturation time, $t_{s(sim)}$, and the maximum sorption capacity, q_{sim} , respectively (Table 4.10). Therefore, considering the level of accuracy achieved, it is deduced that the use of the proposed model based on the semi-empirical Toth equation for the adsorption equilibrium and the Avrami model for adsorption kinetics was suitable to describe the dynamics of CO₂ adsorption in a fixed bed packed with GAC and OXA-GAC adsorbents.



Figure 4.16: Verification of the proposed model with experimental breakthrough curves at three representative operating conditions. Discrete symbols are experimental data, and solid lines are predictions by the model.

CHAPTER 5: CONCLUSION AND RECOMMANDATIONS

5.1 Conclusion

The first objective of this study was to optimize the amination conditions of activated carbon adsorbents in an effort to maximize their CO₂ adsorption/desorption capacities. The effects of amination temperature, amination time, and the type of starting materials (variables) on the CO₂ adsorption/desorption capacities of modified adsorbents (responses) were investigated using a central composite design. Among the process parameters studied, the temperature of ammonia treatment was found to have the most significant positive and negative influence on the CO_2 adsorption and desorption capacity, respectively. The use of a pre-oxidized sorbent as a starting material and amination at 425 °C for 2.1 h were found to be the optimum conditions for obtaining an efficient carbon dioxide adsorbent. This material exhibited CO₂ adsorption and desorption capacity values of 26.47 mg/g and 95.4%, respectively. The experimental values of the responses were found to agree satisfactorily with the values predicted by the models. This indicates that the second-order response surface models were suitable and sufficient to predict CO_2 adsorption/desorption capacities within the investigated range of chosen variables. The adsorption performance of the optimal adsorbent, as well as its desorption capacity under remained essentially the investigated condition. unchanged during cyclic adsorption/desorption operations.

The adsorption equilibria of carbon dioxide on an untreated GAC and its ammoniamodified counterpart were investigated over the temperature range of 303–333 K and up to pressures of 1 atm. Compared to untreated carbon, the OXA-GAC adsorbent exhibited a higher CO₂ uptake, particularly at low partial pressures. To distinguish the contribution of chemisorption and physisorption mechanisms to the overall CO₂ adsorption, we developed a semi-empirical equilibrium model. A non-linear regression method was employed to estimate the best fitting parameters corresponding to the isotherm model. An analysis of the calculated statistical parameters indicated that the proposed model successfully fit the experimental data over the entire analyzed ranges of temperature and pressure. The initial isosteric enthalpy of adsorption calculated using the Clausius– Clapeyron equation indicated a sharp increase in CO_2 –adsorbent interaction after ammonia modification of the untreated adsorbent, consistent with a dramatic uptake of CO_2 at low partial pressures. The heats of adsorption calculated using the temperaturedependent parameters of the proposed model for physisorption and chemisorption of CO_2 onto the modified adsorbent were in excellent agreement with the heats of adsorption obtained from the experimental data.

The kinetics of CO₂ adsorption on ammonia-modified and untreated activated carbon adsorbents over the temperature range of 30-60 °C were studied using the pseudo-firstorder, pseudo-second-order, and Avrami kinetic models. The best fit with the experimental kinetic data for both of the studied adsorbents was obtained by applying the Avrami kinetic model, with average relative errors of less than 2%. The kinetic rate constants of CO₂ capture on both of the adsorbents increased with increasing adsorption temperature. Fixed-bed breakthrough experiments for CO₂ adsorption onto the GAC and OXA-GAC adsorbents were performed by changing the adsorption temperature over the range of 30 to 60 °C and the feed flow rate from 50 to 100 ml min⁻¹. An investigation of the effect of the column adsorption temperature, feed flow rate, and type of adsorbent revealed that using OXA-GAC adsorbent under the operating conditions of 30 °C under a 50 ml min⁻¹ feed flow resulted in the longest breakthrough time (10.9 min) and the highest breakthrough adsorption capacity (0.67 mol kg⁻¹). In addition, to predict the breakthrough behavior of CO₂ adsorption in the fixed-bed column, a simple model was developed, including the Toth and Avrami equations to describe the equilibrium and kinetics of adsorption, respectively. The set of coupled differential equations was solved

using the finite element method implemented in computational fluid dynamics software. The validity of the model predictions was evaluated by a comparison with the experimental data. The results showed that the simulated breakthrough profiles reproduce the experimental breakthrough curves reasonably well under the different operating conditions examined.

5.2 **Recommendations**

Based on the findings of this research, the following recommendations can be utilized for the development of future studies:

- 1. One of the typical variables in the design of an ammonia-modified adsorbent is the type of nitrogen containing functional groups. Considering the established role of the amine groups and pyridine-like functionalities on the high-temperature CO₂ adsorption performance of activated carbon, it could be of interest to explore other nitrogen-related modification techniques, which may produce adsorbents with interesting properties after optimization of synthesis conditions.
- 2. Although the concentration breakthrough curves were predicted very well by the proposed isothermal model, efforts could be placed to develop a nonisothermal model with the inclusion of the radial bed gradients. Since the energy and mass balances are tightly coupled, the radial temperature gradients created by a non-isothermal operation could have a significant effect on the concentration breakthrough curve.
- 3. For the flue gas separation, moisture is a bulk component which was not included in this study. Ternary mixture separations including moisture have to be incorporated in the experiments. The possibility of using layered adsorbent beds to target specific gas components should be investigated.

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LIST OF PUBLICATIONS

Academic Journal

- Mohammad Saleh Shafeeyan, Wan Mohd Ashri Wan Daud, Ahmad Shamiri, "A review of mathematical modeling of fixed-bed columns for carbon dioxide adsorption", Chemical Engineering Research and Design, 92 (2014), Pages 961-988.
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A review of mathematical modeling of fixed-bed columns for carbon dioxide adsorption

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ABSTRACT

Carbon dioxide emissions must be stabilized to mitigate the unfettered release of greenhouse gases into the atmosphere. The removal of carbon dioxide from flue gases, an important first step in addressing the problem of CO_2 emissions, can be achieved through adsorption separation technologies. In most adsorption processes, the adsorbent is in contact with fluid in a fixed bed. Fixed-bed column mathematical models are required to predict the performance of the adsorptive separation of carbon dioxide for optimizing design and operating conditions. A comprehensive mathematical model consists of coupled partial differential equations distributed over time and space that describe material, energy, and the momentum balances together with transport rates and equilibrium equations. Due to the complexities associated with the solution of a coupled stiff partial differential equation system, the use of accurate and efficient simplified models is desirable to decrease the required computational time. The simplified model is primarily established based on the description of mass transfer within adsorption systems. This paper presents a review of efforts over the last three decades toward mathematical modeling of the fixed-bed adsorption of carbon dioxide. The nature of various gas-solid equilibrium relationships as well as different descriptions of the mass transfer mechanisms within the adsorbent particle are reviewed. In addition to mass transfer, other aspects of adsorption in a fixed bed, such as heat and momentum transfer, are also studied. Both single- and multi-component CO2 adsorption systems are discussed in the review. © 2013 The Institution of Chemical Engineers. Published by Elsevier B.V. All rights reserved.

Keywords: Adsorption; Carbon dioxide; Fixed bed; Modeling; Mass transfer; Linear driving force approximation

Contents

1.	Intro	Introduction					
2.	Over	Overview of the prediction of adsorption column dynamics					
3.	Development and analysis of a mathematical model						
	3.1.	3.1. Fluid phase material balance					
	3.2.	Compl	exity of kinetic models	976			
		3.2.1.	Local equilibrium model	976			
		3.2.2.	Mass transfer resistance models	977			
	3.3.	Energy balance		982			
		3.3.1.	Gas phase energy balance	983			
		3.3.2.	Solid-phase energy balance	983			
		3.3.3.	Wall energy balance	983			
	3.4.	Mome	ntum balance	984			
4.	Conclusion						
	Acknowledgments						
	References						

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The application of response surface methodology to optimize the amination of activated carbon for the preparation of carbon dioxide adsorbents

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Keywords: Activated carbon Amination Carbon dioxide adsorption Response surface methodology Optimization

ABSTRACT

In this comparative study, the application of response surface methodology (RSM) in predicting and ontimizing the amination conditions of activated carbon adsorbent toward CO₂ adsorption was investi gated. The adsorbents were prepared based on the central composite design (CCD) with three independent variables (i.e., amination temperature, amination time, and the use of pre-heat treated (HTA) or pre-oxidized (OXA) sorbent as the starting material), and the responses studied were CO2 adsorption and desorp tion capacity. Two quadratic models were developed to calculate the optimum amination conditions of activated carbon that provide a compromise between the studied responses (dependent variables). From the analysis of variance (ANOVA), the temperature of ammonia treatment was found to be the most important factor; it had a positive influence on CO_2 adsorption capacity but a negative effect on desorption capac-ity. The optimal point indicated by numerical optimization corresponded to an OXA sorbent that had been aminated at 425 °C for 2.1 h. The activated carbon modified at optimum conditions had a CO₂ adsorption capacity of 26.47 mg/g and a CO2 desorption capacity of 95.4%. The experimental values of the responses were in good agreement with the amounts predicted by the regression models, indicating that the devel oped models could adequately predict the responses from the amination variables. The stable adsorp-tion/desorption performance of the optimal activated carbon adsorbent during cyclical operations showed its potential for practical applications.

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1. Introduction

Carbon dioxide (CO₂) is one of the most important greenhouse gases, and it is primarily released during the combustion of fossil fuels (coal, oil, and natural gas). Power plants that generate electricity from carbon-based fuels are responsible for approximately 1/3 of all anthropogenic CO2 emissions [1]. The growing environmental concern over global warming and climate change has prompted a global research effort to stabilize the atmospheric concentration of CO2 and mitigate the unfettered release of greenhouse gases into the atmosphere [2,3]. To date, the most common method for capturing CO₂ from gas streams is low temperature absorption with liquid solvents. Because this process has serious drawbacks, such as an energy-intensive regeneration process and a tendency to corrode process equipment, there is a need to develop alternative energy-efficient separation techniques [2,4-6].

The removal of CO2 with solid adsorbents may be a more attractive approach than conventional liquid absorbents because it could reduce the cost associated with the capture step. However, the success of this approach is dependent on the development of an easily

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regenerated and durable adsorbent with high CO2 selectivity and adsorption capacities [4,7,8]. One of the most widely used adsorbents for CO_2 capture from flue gas streams is activated carbon. Activated carbon has a number of attractive characteristics, such as its high adsorption capacity, high hydrophobicity, low cost, and low energy requirement for regeneration [1,9].

Introducing basic nitrogen functionalities into the carbon surface through reactions with nitrogen-containing reagents (e.g., ammonia) can increase the capacity of activated carbon to adsorb $\rm CO_2$ [10-14]. However, the graphitic surface of carbon does not demonstrate a high reactivity toward ammonia. Therefore, carbon is usually oxidized prior to introducing N-containing functionalities to the carbon surface [15,16]. This oxidation-amination process may also enlarge pores and increase surface basicity [17]; such changes in the carbon structure and surface chemistry can increase $\rm CO_2$ adsorption capacity [12,14].

Most of the processes that produce CO_2 in product streams occur at elevated temperatures (up to 100 °C), and the stream must be cooled before separation takes place. Therefore, developing an adsorbent with a high adsorption capacity at relatively high temperatures can drastically reduce the cooling cost of separation and make CO2 capture from power plants feasible [7,12]. The CO2 capture capacity of activated carbon is a combination of both physical and chemical adsorption (i.e., the capacity is governed by

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Modification of Activated Carbon Using Nitration Followed by Reduction for Carbon Dioxide Capture

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Activated carbon (AC) samples were modified using nitration followed by reduction to enhance their CO_2 adsorption capacities. Besides characterization of the samples, investigation of CO_2 capture performance was conducted by CO_2 isothermal adsorption, temperature-programmed (TP) CO_2 adsorption, cyclic CO_2 adsorption, and dynamic CO_2 adsorption tests. Almost all modified samples showed a rise in the amount of CO_2 adsorbed when the comparison is made in unit surface area. On the other hand, some of the samples displayed a capacity superior to that of the parent material when compared in mass unit, especially at elevated temperatures. Despite ~65\% decrease in the surface area, TP- CO_2 adsorption of the best samples exhibited increases of ~10 and 70\% in CO_2 capture capacity at 30 and 100° C, respectively.

Keywords: Activated carbon, Surface modification, Anchoring, Amino groups, Nitration

Introduction

Carbon dioxide is the most significant greenhouse gas that contributes to global warming. This is due to its high emission rate. CO2 capture and sequestration (CCS) is one of the possible stra tegies for mitigating the amount of emitted CO2. However, these approaches are considered as short-term solutions.2 Adsorption is the viable method among the various methods currently in use or undergoing investigation for the initial stage of CCS (i.e., CO_2 capture). Adsorption does not have the drawbacks of amine scrubbing as with most conventional technique for CO₂ capture and has lower energy requirement.³⁻¹² Adsorption capacity of an adsorbate such as CO2 is determined by the surface area, pore size, adsorbent's surface chemistry, temperature, and partial pressure/concentration of the adsorbate. 12,13 In this regard, surface functional groups may strongly increase the adsorption capacity via the creation of specific interactions between the adsorbent and the adsorbate. However, it is worth noting that textural characteristics have a governing effect on the capacity, whereas surface chemistry has an influencing effect.1

Activated carbon (AC) is a suitable adsorbent that has a number of advantages over other porous materials. 10,12,15,16 Modification of AC in terms of surface chemistry has received great attention in producing adsorbents with high capture capacity. It is widely accepted that nitrogen functionalities, in the form of amino or other nitrogen groups, enhance CO₂ adsorption capacity on the AC surface. $^{4,12,13,19-28}$ There are a number of methods that can be used to modify the AC surface with nitrogen functionalities (amino groups, in particular), including impregnation with amine-containing compounds. 12,13,22 amination, $^{13,19-23}$ silylation with aminosilance, nitration followed by reduction, anchoring diaminest polyamines, anchoring halogenated amines, surface-initiated polymerization of ethyleneinine and its derivatives, and plasma treatment. The possible techniques for modifying AC with amino groups have been reviewed.²⁹ Nitration followed by reduction as one of these potential

Nitration followed by reduction as one of these potential methods has been studied in this work. Nitration of AC with the mixture of nitric acid and sulfuric acid (the so-called nitrating mixture³⁰ or mixed acid³¹) creates –NO₂ groups directly attached to the benzene ring of the AC surface.^{30,32} Electrophilic aromatic substitution is believed to be the mechanism of nitration^{31,33,34}: sulfuric acid is stronger than nitric acid and can protonate the latter with the following reaction:

$HNO_3 + 2H_2SO_4 \rightarrow NO_2^+ + H_3O^+ + 2HSO_4^-$

Nitronium ion (NO₂⁺), which is an electrophile center for nitration, is produced by the above reaction.^{50,31,34} This ion is highly reactive and can attach to the aromatic rings with activating or deactivating groups.³⁵ Theoretically, nitric acid alone is not able to nitrate the benzene ring.³⁰ However, some studies show that it can produce a small amount of NO₂⁻ by electrophilic aromatic substitution.^{31,34-36} Reduction of nitro groups to amino groups may be carried out via a variety of methods. Na₂S₂O₁^{33,34} or iron powder³⁷ has been employed for reduction of nitrated AC.

Experimental

Materials. The starting material for experimental study was a commercial AC produced from palm shell by Bravo Green Sdn Bhd, Malaysia. After sieving to the size range of 500–850 µm, it was washed with distilled water to remove

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Anchoring a halogenated amine on the surface of a microporous activated carbon for carbon dioxide capture

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Keywords Activated carbon Surface modification Anchoring Amino groups Halogenated amine ABSTRACT

Activated carbon (AC) surface may be modified by grafting amine-containing compounds to improve the CO₂ adsorption capacity. Two solid sorbents were prepared by anchoring a halogenated amine, *i.e.*, 2-chloroethylamine hydrochloric acid (CEA) on the surface of a microporous AC using a two-stage modification. At the first stage, the samples of AC were oxidized by nitric acid to increase the amount of oxygen surface groups and at the second stage, the oxidized samples were modified by anchoring CEA on the surface to produce a superior CO_2 adsorbent. The oxidized samples were compared with the aid of proximate and ultimate analysis, nitrogen adsorption–desorption at –196 °C (77 K) and temperature programmed desorption (TPD) to decide on the best oxidation conditions. The amine-modified samples were analyzed in terms of texture, surface chemistry and CO₂ adsorption. The latter was studied using isothermal CO₂ capture, temperature-programmed (TP) $\overline{\text{CO}}_2$ adsorption and cyclic operation. The modified samples had a lower surface area than the parent sample. The best modified sample presented an increase of 45% in CO₂ capture capacity at 100 °C. Based on unit surface area, the modified samples showed great CO_2 capture capacities, compared to the virgin sample. Moreover, the modified samples presented a less dependency of CO_2 capacity on temperature. This indicates that the adsorption mechanism shifts from physisorption to chemisorption by increasing temperature, © 2013 Taiwan Institute of Chemical Engineers, Published by Elsevier B.V. All rights reserved.

1. Introduction

Climate change and global warming as a concern of this century is led by emission of greenhouse gases such as carbon dioxide, methane, chlorofluorocarbons (CFCs), ozone and nitrous oxide into the atmosphere [1,2].

Greenhouse effects of chlorofluorocarbons and methane are much higher than carbon dioxide when they are compared in mass unit [2]. However, CO₂ is by far the most important contributor to the global warming due to its high emitted amount by using fossil fuels, which supply around 98% of energy requirement worldwide [3]. According to IPCC, global average of anthropogenic CO_2 emission in the last decade of twentieth century was around 27×10^9 metric tons per year [4]. Growth in demand for energy in 21st century with the

consideration of the point that fossil fuel is still the major source of energy caused a great attraction toward developing solutions to mitigate anthropogenic CO2 emission.

Three general strategies can be thought to cut down the released amount of CO2 [2]. The first one is increasing energy efficiency, which decreases "energy intensity". The second is using energy sources other than fossil fuels to decrease "carbon intensity" and the last one is capture and sequestration of CO2 (CCS).

CCS can be considered as a short- or mid-term, but not longterm, solution to the global warming problem, while new sources of energy such as nuclear, solar and biomass, are matured and developed enough to replace fossil fuels safely and economically [5,6]. Different approaches are currently being developed for both stages of CCS. For sequestration of CO₂, a variety of geoengineering technologies including storage in terrestrial ecosystems, geological formation (e.g. emptied fields of oil and gas and deep saline formations) and deep ocean are possible [7,8]. CO₂ capture is the first and most consuming stage of CCS [2]. It takes approximately 70-80% of CCS total cost [8].

Several methods including absorption (amine scrubbing), adsorption, cryogenic distillation and membrane separation are under use and progressive development for CO_2 capture [6,9,10]. Of these ways, absorption by different amine solutions has been higher developed compared with other methods and so is the most conventional one. However, it has several drawbacks such as being energy sensitive (due to high energy requirement for regeneration step), corrosion, solvent degradation (causing restricted lifetime),

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Carbon Dioxide Capture with Amine-Grafted Activated Carbon

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Abstract There are several possible methods by which amine groups can be grafted on the surface of activated carbon (AC) to improve their capacity for CO2 adsorption. Ethylenediamine and diethylenetriamine were selected as amino compounds for anchoring on the surface of an oxidized AC. Oxidation of AC was carried out by concentrated nitric acid. For each amino compound, two "in-solvent" and "solventfree" methods with a number of grafting times were studied. Nitrogen adsorption-desorption at 77 K and proximate and ultimate analysis were used to determine physical and chemical characteristics of the samples. Temperature-programmed (TP) CO2 adsorption test from 30°C to 120°C were performed to investigate the effect of modification on CO2 capture. The modification clearly had a negative

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M.-G. Lee Department of Chemical Engineering, Pukyong National University, Busan 608-739, Republic of Korea e-mail: mglee@pknu.ac.kr effect on the textural characteristics of the samples, so the samples showed a less CO2 uptake at lower temperatures. However, the decrease of capture capacity with increasing temperature is to somewhat softer for amine-grafted samples, so that they have a capacity comparable to the parent sample or even more than that at elevated temperatures. This property may give the new adsorbents this opportunity to be used at flue gas temperature with a higher efficiency. CO2 capture capacity per unit surface area of all the amine-modified samples, however, was significantly improved, compared to the parent sample presenting a great influence of amino groups on the CO₂ capture capacity. Moreover, the used amine compounds and grafting methods were compared in terms of adsorbent characteristics and CO2 uptake curves. Cyclic adsorption-desorption tests showed a satisfactory regeneration for the modified samples.

Keywords Activated carbon · Surface modification · Amine grafting · Ethylenediamine · Diethylenetriamine

1 Introduction

Energy consumption continues to grow worldwide in this century. Considering that fossil fuel is still the major source of energy, a great attention has been concentrated on developing solutions to mitigate anthropogenic CO_2 emission. Available methods

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Statistical Modeling and Optimization of Amination Conditions of Activated Carbon for Carbon Dioxide Adsorption Using Response Surface Methodology

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Abstract. The surface modification of activated carbon with gaseous ammonia for obtaining an efficient carbon dioxide (CO₂) adsorbent was optimized using the response surface methodology. The variables chosen were amination temperature, amination time, and the use of pre-heat treated (HTA) or pre-oxidized (OXA) sorbent as the starting material, while the responses were CO₂ adsorption and desorption capacity. Among process parameters studied, amination temperature was found to have the most significant effect on both the studied responses. Quadratic models developed for the two responses studied (dependent variables) indicated the optimum conditions could be obtained by using OXA sorbent with amination temperature of 425 °C and amination time of 2.1 h. The activated carbon modified at optimum condition showed CO₂ adsorption capacity of 26.1 mg/g and CO₂ desorption capacity of 94.9%. The experimental values of the responses at the optimum condition were found to agree satisfactorily with the amounts predicted by the models.

Keywords: Activated carbon, Ammonia modification, Carbon dioxide adsorption, Response surface methodology, Optimization

1. Introduction

Carbon dioxide (CO₂) is one of the most important greenhouse gases, and it is primarily released during the combustion of fossil fuels (coal, oil, and natural gas). The growing environmental concern over global warming and climate change has prompted a global research effort to stabilize the atmospheric concentration of CO₂ and mitigate the unfettered release of greenhouse gases into the atmosphere [1]. The removal of CO₂ with solid adsorbents may be a more attractive approach than conventional liquid absorbents because it could reduce the cost associated with the capture step. However, the success of this approach is dependent on the development of an easily regenerated and durable adsorbent with high CO₂ selectivity and adsorption capacities [2]. One of the most widely used adsorbents for CO₂ capture from flue gas streams is activated carbon.

Introducing basic nitrogen functionalities into the carbon surface through reactions with nitrogencontaining reagents (e.g., ammonia) can increase the capacity of activated carbon to adsorb CO_2 [3]. However, the graphitic surface of carbon does not demonstrate a high reactivity toward ammonia. Therefore, carbon is usually oxidized prior to introducing N-containing functionalities to the carbon surface [4]. In the case of high-temperature adsorption, although the role of surface chemistry and the presence of basic nitrogen functionalities is more considerable [5], only the optimization of textural properties is well reached in literature. Although the effect of surface chemistry modification on CO_2 adsorption performance depends on the modification conditions, as far as we know, no study has been done to optimize these modification conditions.

Accordingly, in the present study, RSM based on central composite design (CCD) was used to design experiments, build models and determine the optimum modification conditions for desirable responses. The main objective of this work was to investigate the effect of modification parameters (amination temperature,

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